

NICKEL SPECIATION OF URBAN PARTICULATE MATTER

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ABSTRACT

A four-step sequential Ni extraction method, summarized in Table AB-1, was evaluated for identifying and quantifying the Ni species occurring in urban total suspended particulate (TSP) matter and fine particulate matter ($<10\text{ }\mu\text{m}$ [PM_{10}] and $<2.5\text{ }\mu\text{m}$ [$\text{PM}_{2.5}$] in aerodynamic diameter). The extraction method was originally developed for quantifying soluble, sulfidic, elemental, and oxidic forms of Ni that may occur in industrial atmospheres. X-ray diffraction (XRD) and x-ray absorption fine structure (XAFS) spectroscopy were used to evaluate the Ni species selectivity of the extraction method. Uncertainties in the chemical speciation of Ni in urban PM_{10} and $\text{PM}_{2.5}$ greatly affect inhalation health risk estimates, primarily because of the large variability in acute, chronic, and cancer-causing effects for different Ni compounds.

Table AB-1. Four-Step Ni Extraction Procedure

Step	Description	Ni Form
I	15-mL deionized H_2O , mix for 10 s, expose for 30 min, centrifuge 10 min at 3000 rpm	Soluble
II	15-mL 1:1 30% H_2O_2 –0.1 M citrate, mix for 10 s, expose for 30 min, centrifuge 10 min at 3000 rpm	Sulfidic
III	10-mL 50:1 methanol–bromine solution, mix for 10 s, expose for 20 min, centrifuge 10 min at 3000 rpm	Metallic
IV	0.5-mL HF–1 mL HNO_3 –0.5 mL HCl, microwave 4 min at 600 W; wait 1 min, then 4 min at 600 W	Oxidic

A mixture of nickel sulfate hexahydrate ($\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$), nickel subsulfide (Ni_3S_2), metallic nickel (Ni^0), and nickel oxide (NiO) was analyzed eight times using the sequential Ni extraction method. Chemical mass balance considerations and XRD and XAFS spectroscopy analyses of extraction residues indicated that the soluble Ni extraction procedure provided a reliable estimate of $\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$ concentration but underestimated Ni_3S_2 and NiO and overestimated Ni^0 concentrations.

A National Institute of Standards and Technology urban TSP standard reference material (SRM) 1648 (urban TSP from St. Louis, Missouri) was analyzed in quadruplicate using the sequential Ni extraction method. Extraction results indicated that soluble and oxidic Ni forms accounted for about 80% of the total Ni in SRM 1648 and sulfidic and metallic Ni forms accounted for 20% of the remaining Ni. XAFS spectroscopy analyses of SRM 1648 extraction residues, however, were inconclusive for corroborating Ni species results because of an uneven and insufficient filter particle loading.

TSP, PM₁₀, and PM_{2.5} were sampled continuously using a high-volume air sampler, an automatic cartridge collection unit, and sequential air sampler, respectively, during August 26–31, 2002, from an urban air monitoring site in Davie, Florida. Total mass loadings on the filters corresponded to average concentrations of airborne TSP, PM₁₀, and PM_{2.5} of 26.8, 10.4, and 6.3 µg/m³, respectively. An insufficient amount of PM_{2.5} was sampled to quantify Ni and perform speciation measurements. TSP and PM₁₀ Ni concentrations (86 ppm and ≈140 ppm) correspond to airborne Ni concentrations of 2.3 and 1.5 ng/m³, respectively. These low Ni concentrations probably represent the atmospheric burden of “background” Ni derived from long-range transport via the North Atlantic trade winds.

Sequential Ni extraction and XAFS results indicated that NiSO₄ · xH₂O and a nickel oxide compound, possibly NiFe₂O₄, were the dominant Ni species occurring in Davie, Florida, TSP. Both methods also indicated that a small proportion of NiS was present. XAFS spectroscopy analyses indicated that the sequential Ni extraction procedure selectively and quantitatively removed NiSO₄ · xH₂O from Davie TSP but overestimated the sulfidic Ni and Ni⁰ fractions and underestimated oxidic Ni. Davie PM₁₀ lacked NiS, but NiSO₄ · xH₂O was more abundant than in TSP, probably because of the greater surface area of PM₁₀ on which NiSO₄ · xH₂O heterogeneously condensed or adsorbed.

Automated scanning electron microscopy analyses of 800–1100 particles/sample indicated that 8%–15% of the total number of particles sampled from Davie, Florida, contained significant Ni concentrations, ≥10 wt%. Most of these Ni-bearing particles were <10 µm in diameter suggesting that they primarily originated from combustion sources.

Direct metal speciation techniques such as XAFS spectroscopy should be used to verify the applicability and reliability of sequential Ni extraction methods before they are employed on a specific sample type (raw materials, workplace dust, ambient air TSP, PM₁₀, PM_{2.5}, etc.). The limitations of sequential Ni extraction methods applied to urban TSP, PM₁₀, and PM_{2.5} samples need to be considered when using such indirect Ni speciation results to assess the potential inhalation health risks associated with individual Ni species.

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NOMENCLATURE

ACCU	automatic cartridge collection unit
ASEM	automated scanning electron microscopy
CAAA	Clean Air Act Amendments
EPA	U.S. Environmental Protection Agency
EPMA	electron probe microanalysis
GFAAS	graphite furnace atomic absorption spectroscopy
HVAS	high-volume air sampler
ICP–AES	inductively coupled plasma–atomic emission spectroscopy
m ³	cubic meter
NAMS	National Air Monitoring System
ng	nanogram (10 ^{−9} g)
Ni ⁰	metallic nickel
Ni ₃ S ₂	nickel subsulfide
NSLS	National Synchrotron Light Source
NIST	National Institute of Standards and Technology
PM	particulate matter
PM ₁₀	particulate matter with aerodynamic diameters <10 μm
PM _{2.5}	particulate matter with aerodynamic diameters <2.5 μm
ppm	part per million
s	second
SAS	sequential air sampler
SLAMS	State and Local Air Monitoring Station
SRM	standard reference material
SSRL	Stanford Synchrotron Radiation Laboratory
TRI	Toxics Release Inventory
TSP	total suspended particulate
μg	microgram (10 ^{−6} g)
wt%	weight percent
XAFS	x-ray absorption fine structure
XANES	x-ray absorption near edge structure
XRD	x-ray diffraction

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1.0 INTRODUCTION

Nickel (Ni) is a transition metal that occurs throughout the environment as it is discharged into air, water, and soil from various natural and industrial sources. A significant natural source of atmospheric Ni is windborne dust particles derived from the weathering of rocks and soils and from volcanic eruptions.¹ The average Ni concentration of crustal rocks is 75 ppm.² Although Ni is a ubiquitous metal, industrialization has increased its flux into the environment.³ Anthropogenic stationary sources that release Ni into ambient air include 1) combustion and incineration sources (oil- and coal-burning units in utility, industrial, and residential sectors and medical, municipal, and sewage sludge incinerators); 2) high-temperature metallurgical operations (steel and Ni alloy manufacturing, secondary metals smelting, and coproduct Ni recovery); 3) primary production operations (mining, milling, smelting, and refining); and 4) chemical and catalyst sources (Ni chemical manufacturing, electroplating, Ni–Cd battery manufacturing, and catalyst production, use, and reclamation).^{4,5} The mobile source contribution to Ni emission inventories is small and derives primarily from engine wear and impurities in engine oil and fuel additives.⁶ However, commercial marine vessels are significant mobile sources of Ni in areas near harbors.⁷

Ni-bearing particles occur in the atmosphere as part of suspended particulate matter (PM) and, rarely, of mist aerosols. Ni is most commonly associated with the fine PM fraction of ambient air samples with diameters ranging from 0.6 to 10 μm .⁸ These fine Ni-bearing particles are susceptible to human inhalation and, therefore, pose significant health concerns. Particles with aerodynamic diameters less than 10 μm (PM_{10}) are likely to penetrate beyond the nose, and those less than 2.5 μm ($\text{PM}_{2.5}$) are likely to penetrate deep into the lung.

The Ni concentration of PM in the U.S. atmosphere ranges from 0.01 to 60, 0.6 to 78, and 1 to 328 ng/m^3 in remote, rural, and urban areas, respectively.⁹ Sunderman and Oskarsson¹⁰ reported that Ni concentrations in ambient air average 6 ng/m^3 in rural areas, 25 ng/m^3 in urban areas, and 170 ng/m^3 in large metropolitan or industrial areas. Average ambient air Ni concentrations in U.S. and Canadian cities range from 5 to 50 ng/m^3 and 1 to 20 ng/m^3 , respectively.^{11,12}

In the United States, the impetus for focusing on individual elements, such as Ni, in air pollution derives from the 1990 Clean Air Act Amendments (CAAA)¹³ and, more recently, the attainment of National Ambient Air Quality Standards.¹⁴ Similar legislation has been enacted or proposed in many other countries as well. Title III of the CAAA identifies 189 chemicals, including Ni and 15 other inorganic trace elements (As, Be, Cd, Cl, Co, Cr, F, Hg, Mn, P, Pb, Sb, Se, Th, and U), as potential hazardous air pollutants or air toxics. Many stationary sources have had to report Ni emissions as part of the U.S. Environmental Protection Agency's (EPA) Toxics Release Inventory (TRI).¹⁵ Although TRI and similar reporting provide estimates of the amounts of Ni released into the environment, they are not an indicator of toxicity because the acute, chronic, and cancer-causing effects vary significantly for the different chemical species of Ni. For example, nickel subsulfide (Ni_3S_2) is considered the most carcinogenic Ni species on the basis of available human epidemiology and animal studies.^{5,11,12,16} The carcinogenicity of water-soluble Ni salts such as $\text{NiSO}_4 \cdot x\text{H}_2\text{O}$ and

insoluble nickel oxide compounds such as Ni-containing spinels (e.g., NiFe_2O_4), however, is controversial.^{5,16,17} Therefore, understanding the speciation of Ni in ambient air is of the utmost importance for assessing potential respiratory health risks associated with Ni.

The primary purpose of this investigation was to evaluate an improved sequential Ni extraction method, summarized in Table 1, for its application to Ni ambient air samples. This method, developed by Luk et al.¹⁸ for the Nickel Producers Environmental Research Association, improved the Zatka et al.¹⁹ sequential Ni extraction method by incorporating the following modifications: 1) replaced filtration with centrifugation, 2) replaced conventional hot plate sample dissolution approach with microwave digestion, 3) optimized extraction durations, and 4) modified the leaching reagents. Bolt et al.²⁰ and Füchtjohann et al.²¹ have also recently proposed improved sequential Ni extraction methods. These methods are based on treating a small but representative sample of PM with successive leaching steps to separate analyte Ni species from the sample matrix. Ni species concentrations in the extracted fractions are then determined using sensitive analytical techniques, such as graphite furnace atomic absorption spectroscopy (GFAAS) and inductively coupled plasma–atomic emission spectroscopy (ICP–AES). Sequential extractions provide an operationally defined (i.e., based on solubility differences) determination of Ni species (soluble, sulfidic, metallic, and oxidic). These methods are also useful for evaluating the environmental mobility and bioavailability of Ni.

Table 1. Four-Step Ni Extraction Procedure Developed by Luk et al.¹⁸

Step	Description	Ni Form
I	15-mL deionized H_2O , mix for 10 s, expose for 30 min, centrifuge 10 min at 3000 rpm	Soluble
II	15-mL 1:1 30% H_2O_2 –0.1 M citrate, mix for 10 s, expose for 30 min, centrifuge 10 min at 3000 rpm	Sulfidic
III	10-mL 50:1 methanol–bromine solution, mix for 10 s, expose for 20 min, centrifuge 10 min at 3000 rpm	Metallic
IV	0.5-mL HF–1 mL HNO_3 –0.5 mL HCl, microwave 4 min at 600 W; wait 1 min, then 4 min at 600 W	Oxidic

Ni speciation analyses were conducted on a known Ni compound mixture, a National Institute for Standards and Technology (NIST) standard reference material (SRM 1648), and an urban total suspended particulate (TSP) sample collected from Davie, Florida, using Luk et al.’s¹⁸ improved sequential Ni extraction method. Two complementary methods: x-ray diffraction (XRD) and/or x-ray absorption fine structure (XAFS) spectroscopy were also used for determining the Ni speciation of TSP, PM_{10} , and $\text{PM}_{2.5}$ samples and Ni extraction residues from TSP samples. The XAFS technique is especially well suited for determining Ni speciation because it can directly and nondestructively analyze TSP, PM_{10} , and $\text{PM}_{2.5}$ filter samples with ppm sensitivity.^{22–25} XAFS spectroscopy, a definitive speciation technique, involves recording the variation of the x-ray absorption coefficient, as a function of energy, in the vicinity of a characteristic absorption edge for the element of interest. The x-ray flux is supplied from a very intense synchrotron x-ray source. Analysis of variations in the

x-ray absorption coefficient provides structural information concerning the local environment of the analyte element, in terms of oxidation state, interatomic distances, and the type and number of coordinating ligands. The primary limitations of this direct speciation analysis technique are that the structural information obtained represents a weighted average for all the different chemical species of an element that may exist in a sample and the requirement of a synchrotron x-ray source. The capability of XAFS spectroscopy to directly determine Ni speciation was used to evaluate the Ni species selectivity of sequential extractions and to corroborate speciation results. Automated scanning electron microscopy (ASEM) was also used to determine the number concentrations (frequency %) of Ni-containing particles in TSP samples and to evaluate the particle-size distributions of Ni.

2.0 EXPERIMENTAL

2.1 Samples

2.1.1 Model Ni Compound Mixture

Four reagent-grade Ni compounds, 1) nickel sulfate hexahydrate ($\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$), 2) Ni_3S_2 , 3) metallic nickel (Ni^0), and 4) green nickel oxide (NiO), were used to prepare a mixture of known Ni compounds. An XRD analysis of the Ni_3S_2 compound indicated that it is contaminated with a very small amount of an additional sulfidic Ni compound, Ni_7S_6 . Approximately 200 mg of each compound was weighed in a 20-mL capacity plastic scintillation vial. Several glass mixing beads were added, and the vial was shaken in a SPEX mixer mill for 15 min to homogenize the mixture. The model nickel compound mixture consisted of 25 wt% $\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$, 26 wt% Ni_3S_2 , 23 wt% Ni^0 , and 26 wt% NiO .

2.1.2 NIST SRM 1648 Urban TSP

SRM 1648 consists of atmospheric TSP collected in an urban location of St. Louis, Missouri. NIST collected TSP in a specially designed baghouse for a period in excess of 12 months and, therefore, it is a time-integrated sample. SRM 1648 was used in this investigation, as intended by NIST, to typify the analytical problems of TSP samples obtained from industrialized urban areas. SRM 1648 is sold in 2-g portions, and Ni is a sufficiently concentrated, certified concentration of 82 ± 3 ppm, in it for sequential Ni extraction and XAFS speciation analyses.

2.1.3 Urban TSP, PM_{10} , and $\text{PM}_{2.5}$

Ambient TSP, PM_{10} , and $\text{PM}_{2.5}$ were sampled continuously during August 26–31, 2002, from an urban State and Local Air Monitoring Stations (SLAMS) and National Air Monitoring System (NAMS) site located in Davie, Florida (26.08°N, 80.24°W). The Broward County Air Quality Division utilizes this urban (citywide scale, 4–50 km) site to determine representative PM_{10} and $\text{PM}_{2.5}$ concentrations in an area of high population density. According to the Broward County Planning Services Division,²⁶ the 2002 population density of Broward County was 3837 persons/mi².

The following equipment and EPA Compendium Methods²⁷ were used to sample ambient TSP, PM₁₀, and PM_{2.5}:

- Tapered electronic oscillating microbalance, Rupprecht & Patashnick Model 1400a (EPA Compendium Method IO-1.3).
- Automatic cartridge collection unit (ACCU), Rupprecht & Patashnick
- Sequential air sampler (SAS), Rupprecht & Patashnick Model 2025 (EPA Compendium Method IO-2.3)
- High-volume air sampler (HVAS) and General Metal Works (EPA Compendium Method IO-2.1)

TSP was collected on Gelman Scientific Type A/E glass-fiber filters, whereas PM₁₀ and PM_{2.5} were collected on Whatman GF/C glass fiber filters. Urban TSP was also collected on a moving carbon tape using a Burkard Seven-Day Recording Volumetric Spore Trap. Field filter blank samples were collected for quality control purposes. Meteorological conditions (e.g., temperature, wind speed and direction, and relative humidity) were also recorded during the sampling period. However, late in the sampling campaign, lightning struck the weather mast at the Davie site, and the recorded meteorological data were lost.

The locations of major PM₁₀ point sources in relation to the sampling site and a wind rose diagram, created from hourly average wind speed and direction measurements from the Fort Lauderdale International Airport, for the sampling period are shown in Figure 1. The PM₁₀ emission estimates in Figure 1 are based on 2001 statistics from the Florida Department of Environmental Protection. Point sources in the vicinity of the Davie site include residual (No. 6 fuel) oil- and gas-burning power plants, cement kiln operations, and municipal waste incineration facilities. Significant mobile sources of TSP, PM₁₀, and PM_{2.5} near the Davie site include motor vehicle, aircraft, and marine vessel traffic. Marine vessels are also potential Ni emission sources because most burn distillate and residual-grade fuel oils which generally contain significant Ni contents.^{7,28} Approximately 5400 ships call at Port Everglades a year.²⁹ The wind rose diagram in Figure 1 indicates that winds were predominantly marine in origin and that mobile and point sources east-southeast and southeast from the Davie site probably contributed to TSP, PM₁₀, and PM_{2.5} during the sampling period. Two meteorological processes that occurred during the sampling period were synoptic southeast and easterly winds associated with the tropical North Atlantic trade winds and the formation of deep convective thunderstorm cells. Tall (12–16 km) convective thunderstorms were generally generated during the afternoon when moist air from the ocean breeze was carried aloft by the hot air masses rising off the southern peninsula of Florida. In contrast to low-altitude frontal storms, these tall convective thunderstorms scavenge PM from the middle and upper troposphere.

2.2 Sequential Ni Extractions and Analyses

The four-step extraction procedure summarized in Table 1 was evaluated for determining Ni speciation. The extraction procedure was designed specifically for determining the soluble, sulfidic,

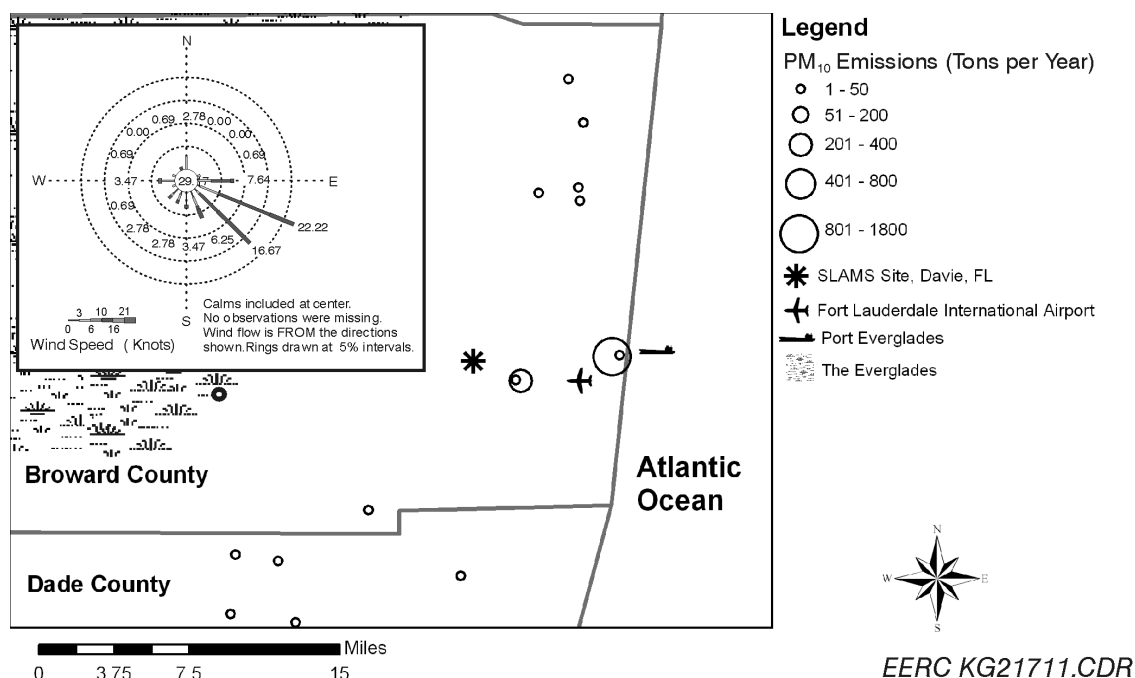


Figure 1. Davie, Florida, SLAMS/NAMS site, major PM₁₀ point source, Fort Lauderdale International Airport, and Port Everglade locations and a wind rose diagram for the sampling period.

metallic, and oxidic forms of Ni found in industrial atmospheres. Ni in each extract was quantified using GFAAS or ICP–AES. In addition, urban TSP, PM₁₀, and PM_{2.5} samples and filter blanks were prepared using American Society for Testing and Materials Method D6357³⁰ and analyzed using GFAAS. The sulfur concentration of TSP was determined by performing the first two Ni extraction steps in Table 1 and analyzing the extracts by ion chromatography.

Ni extraction residues were prepared for XAFS spectroscopy analyses by performing the first three Ni extraction steps (Table 1). However, instead of separating the residue from the extraction fluid by centrifugation, they were separated by filtering on Whatman GF/C glass-fiber filters. Residues were then dried in an oven at 50°C before analyzing them using XAFS spectroscopy and XRD.

2.3 XRD

XRD analyses were performed on extraction residues and the Davie TSP sample to identify crystalline phases. XRD patterns were collected over 5°–70° 2θ with a Philips X'Pert theta-theta x-ray diffractometer system operating at 40 kV and 50 mA (graphite monochromatized Cu K-α radiation, 0.02° 2θ steps, 3 sec/step). Incident beam optics consisted of a programmable divergence slit fixed at ¼° and a 15-mm wide-beam mask. Diffracted beam optics consisted of an antiscatter slit programmed for an automatic 10-mm irradiated length, programmable receiving slit set to a height of 0.35 mm, soller slits, and a secondary monochromator. Diffraction peaks were identified using

a Minerals Data Incorporated JADE+ processing software utilizing the ICDD PDF-2 inorganic and organic powder diffraction database, Sets 1–45 on CD-ROM.

2.4 XAFS Spectroscopy

XAFS spectroscopy was performed at beam-line X-18B of the National Synchrotron Light Source (NSLS), Brookhaven National Laboratory, New York, during July 26–30, 2002, and at beam line 4-3 of the Stanford Synchrotron Radiation Laboratory (SSRL), Stanford University, California, during January 16–23 and February 18–24, 2003. The NSLS was operated at a constant voltage of 2.85 GeV, and the current generally decayed from 270 to 150 mA between fills every 12 hours. The synchrotron at SSRL was operated at a constant voltage of 3.00 GeV, and the current generally decayed from 100 to 65 mA between fills every 24 hours. Ni XAFS spectra were measured in absorption and fluorescence geometries. The latter spectra were obtained using a PIPS detector or a 12-GeV array detector, depending on Ni concentration. Sulfur K-edge XAFS spectra were collected at the NSLS in fluorescence geometry using a PIBS detector, whereas at the SSRL a Lytle detector was used. A Co filter was also used to enhance the signal-to-noise ratio of the Ni fluorescent XAFS spectra.³¹ The Ni K-absorption edge at 8,333 eV was defined as the major peak in the first derivative XAFS spectrum of a thin Ni⁰ foil acquired simultaneously in absorption geometry behind the main experiment. The S K-absorption edge at 2,472 eV was defined by the maximum white line absorption peak for an elemental sulfur calibration standard. Ni XAFS spectra were generated as a function of x-ray energy by rotating a double Si(220) crystal monochromator from about 100 eV below the Ni K-edge to about 500 eV above. For S, a Si(111) channel cut crystal monochromator was rotated from 72 eV below and 250 eV above the S K-edge. Multiple scans, typically between 3 and 5 scans, were combined to produce the overall spectrum.

Ni and S XAFS spectra were analyzed using the procedures described by Brown et al.,³² Lee et al.,³³ and Koningsberger and Prins.³⁴ First, the energy scale was calibrated with respect to the zero-point of energy defined by the calibration standard spectrum. Then the XAFS spectrum of the unknown sample was divided into separate x-ray absorption near-edge structure (XANES) and extended XAFS regions. The edge step parameter provided an approximation of Ni concentration. XANES spectra of reagent-grade Ni compounds and synthesized Ni compounds, acquired in previous investigations,^{25,35,36} were used essentially as “fingerprints” for identifying and quantifying Ni species. The following Ni compounds are included in this spectral database: Ni⁰, NiS, NiS₂, Ni₃S₂–Ni₇S₆ mixture, “green” NiO, Ni(OH)₂, NiSO₄ · xH₂O, C₄H₆NiO₄ · 4H₂O, NiCO₃, NiFe₂O₄, MgNiO₂, (Mg_{0.2}Ni_{0.8})₂SiO₄, and (Mg_{0.85}Ni_{0.15})₂SiO₄.

2.5 ASEM

The 20-mm-wide carbon tape from the Burkard Seven-Day Recording Volumetric Spore Trap was cut into 48-mm-long pieces representing a 24-hour period. Starting times were known, and the tapes were cut to represent 8:00 a.m.–5:00 p.m. sampling periods. These strips, representing 9 hours of TSP collection, were then mounted on a standard petrographic slide and inserted directly into the ASEM for analysis without further sample preparation.

The ASEM method utilizes digital image analysis and electron probe microanalysis (EPMA) techniques to locate, size, and chemically analyze individual ash particles with submicron diameters as small as 0.3 μm . The analyses were conducted using a JEOL JSM-5800 SEM equipped with a Noran Instruments Voyager x-ray analysis and automation system. The SEM was operated at an accelerating voltage of 15 kV and a beam current of 600 pA. EPMA of each particle was conducted for 10 s. The following elements were analyzed: Na, Mg, Al, Si, P, S, Cl, K, Ca, Ti, V, Fe, and Ni using the atomic number, absorption, and fluorescence (ZAF) method which corrects x-ray intensities for differences in the atomic number (Z), absorption (A), and fluorescent (F) effects of the calibration standards relative to the sample. The chemical compositions obtained in this manner are semiquantitative at best because of the short x-ray counting time employed (10 s), the use of flat mineral standards for calibration, and the fact that no matrix corrections for particle diameter, shape, or density were applied.

3.0 RESULTS AND DISCUSSION

3.1 Model Ni Compound Mixture

3.1.1 Sequential Ni Extraction Analysis

Eight representative portions, 10–15 mg, of the known Ni compound mixture were analyzed to evaluate the selectivity of the four-step sequential Ni extraction method summarized in Table 1. The Ni extraction recoveries presented in Figure 2 (additional extraction data are also presented in

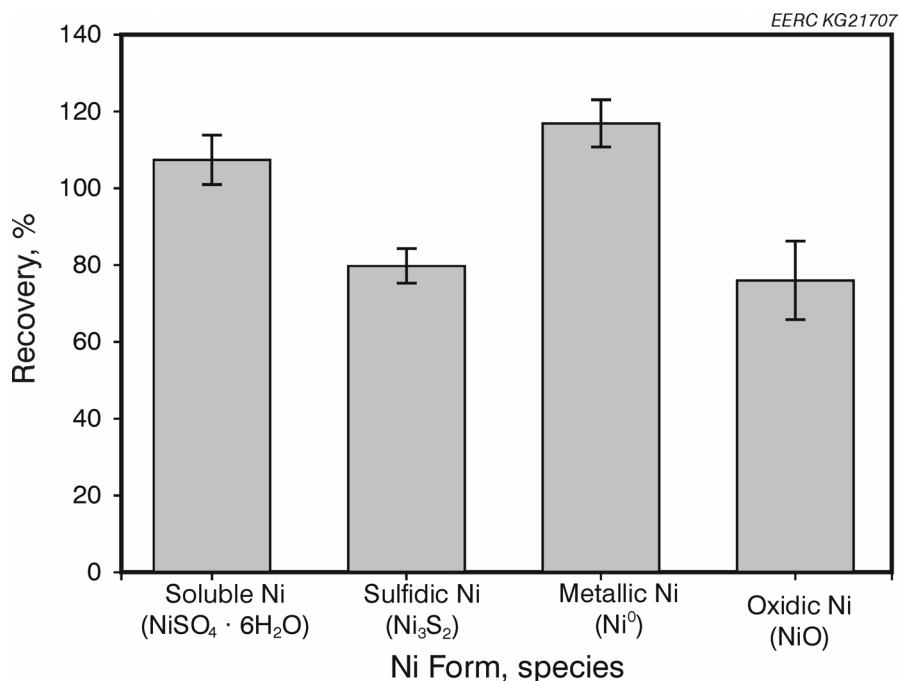


Figure 2. Mean Ni extraction recoveries for model Ni compound mixture. The error bars represent $\pm 95\%$ confidence limits calculated from eight sequential Ni extraction measurements.

Appendix A, Table A-1) indicate that the first extraction step recovered, on average, 107% of the soluble $\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$. Recoveries of the sulfidic (Ni_3S_2) and oxidic (NiO) forms were biased low by about 20%; conversely, Ni^0 was biased high by about 20%. Sulfidic Ni is probably biased low because of incomplete dissolution and/or sulfidic Ni separation resulting in 20 wt% less Ni extracted and 20 wt% more Ni remaining in the residue that is incorrectly measured as Ni^0 during the subsequent extraction step. In addition, the partial dissolution of Ni_3S_2 and its coextraction with soluble Ni during the first extraction step may have contributed to the low sulfidic Ni recovery. Oxidic Ni is presumably biased low because of incomplete sample digestion during the fourth extraction step. Results presented by Luk et al.¹⁸ for known Ni compound mixtures (Tables 11a–11c in Reference 18) also indicated greater soluble Ni and Ni^0 recoveries relative to sulfidic and oxidic Ni recoveries.

3.1.2 XRD and XAFS Spectroscopy Analyses

The selectivity of the four-step sequential Ni extraction procedure was also investigated by repeating Step I, Step I and II, and Step I–III extractions on the model Ni compound mixture and then analyzing the residues using XRD and XAFS spectroscopy. XRD spectra of the three extraction residues in Figures 3–5 indicate that the first extraction step effectively removed soluble $\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$; the sulfidic Ni extraction (Step II) effectively removed Ni_3S_2 ; and the Ni^0 extraction (Step III) removed most but not all the Ni^0 . Presented in Figure 6 are the XANES spectra of the three extraction residues. Similar spectra obtained on reagent-grade $\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$, Ni_3S_2 , Ni^0 , and “green” NiO were compared to the spectra obtained on the residues using a least-squares fitting technique to quantify Ni speciation. Quantitative XANES results for the soluble, sulfidic, and metallic Ni extraction residues are compared to expected Ni species proportions in Figures 7–9. The expected compositions were calculated assuming that the first three extraction steps selectively and quantitatively removed $\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$, Ni_3S_2 , and Ni^0 . XANES measurements of Ni^0 and NiO concentrations in Figures 7–9 compare favorably to the corresponding calculated concentrations, with relative percent differences of $\leq 14\%$. The XANES estimate of the Ni_3S_2 proportion in the soluble Ni extraction residue, however, is 25% less than expected. XANES results in Figures 7–9 indicate that the first three extraction steps were effective in removing $\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$, Ni_3S_2 , and Ni^0 , respectively. Similar to the Ni recovery results in Figure 2 and Table A-1, the XANES results suggest that the extraction procedure underestimated sulfidic Ni. However, contrary to the Ni recovery results in Figure 2, XANES results indicate that the extraction procedure did not significantly overestimate Ni^0 contents. This implies, assuming that the XANES analysis of Ni_3S_2 is accurate, that the soluble Ni extraction removed a portion of Ni_3S_2 in addition to $\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$.

3.2 NIST SRM 1648 (Urban TSP)

3.2.1 Sequential Ni Extraction Analysis

In addition to the known Ni compound mixtures, SRM 1648 was analyzed in quadruplicate (approximately 16-mg samples) using the four-step sequential Ni extraction method. Extraction results are summarized in Figure 10 and additional extraction data are presented in Tables A-2 and A-3. Total Ni recoveries, based on the NIST certified Ni concentration of 82 $\mu\text{g/g}$, were acceptable with values ranging from 86% to 100% and averaging 92%. The results in Figure 10 indicate that

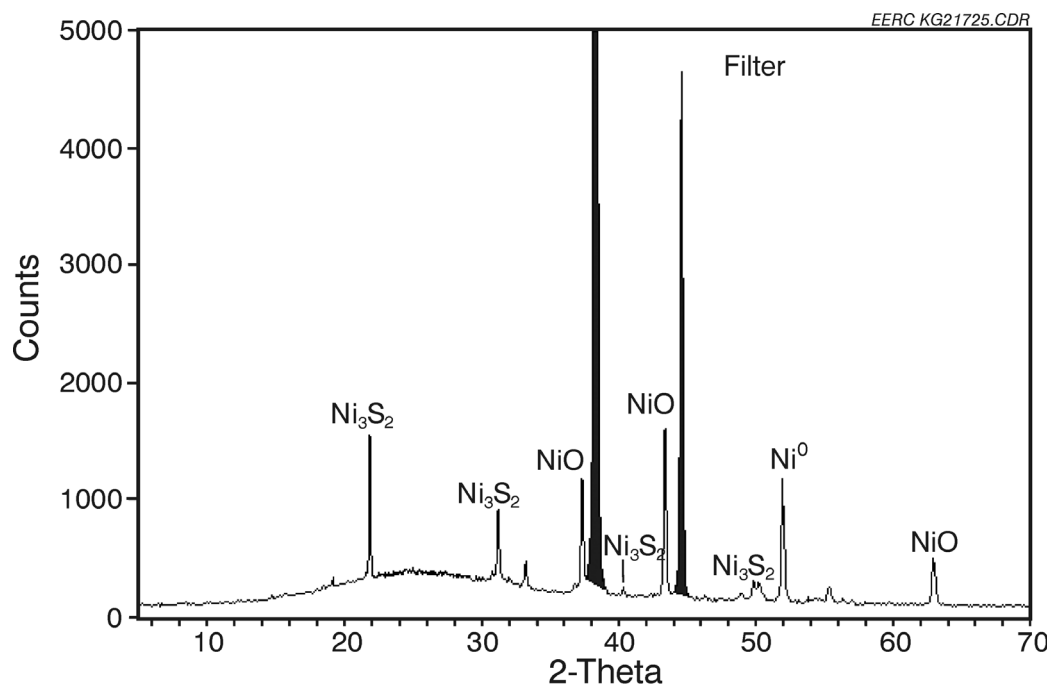


Figure 3. XRD pattern of soluble Ni extraction residue prepared from a standard $\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$ – Ni_3S_2 – Ni^0 – NiO compound mixture.

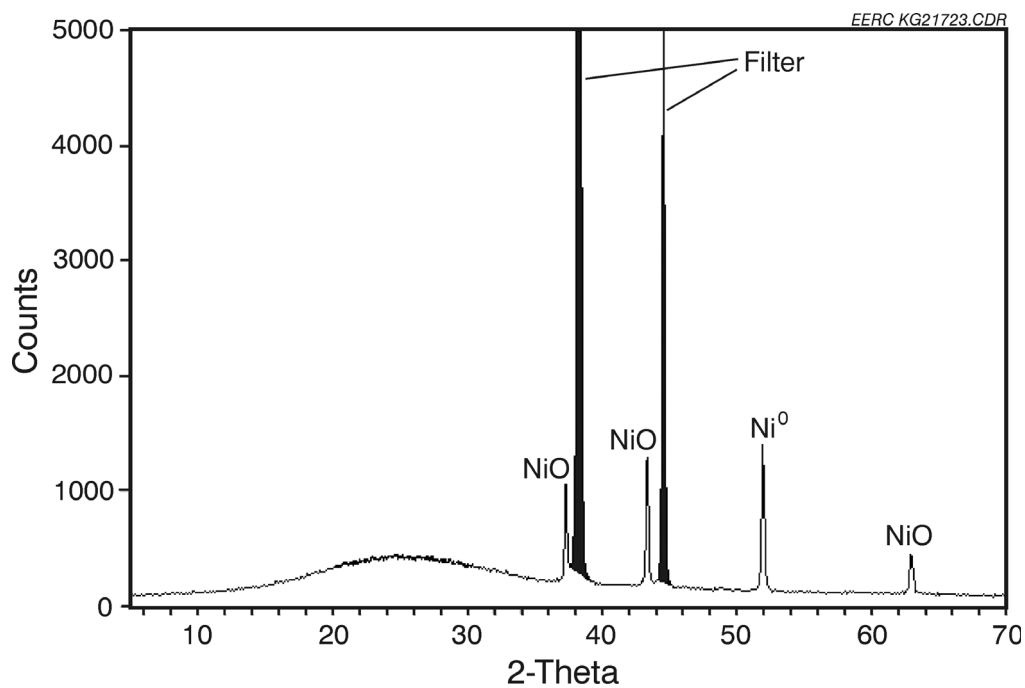


Figure 4. XRD pattern of sulfidic Ni extraction residue prepared from a standard $\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$ – Ni_3S_2 – Ni^0 – NiO compound mixture.

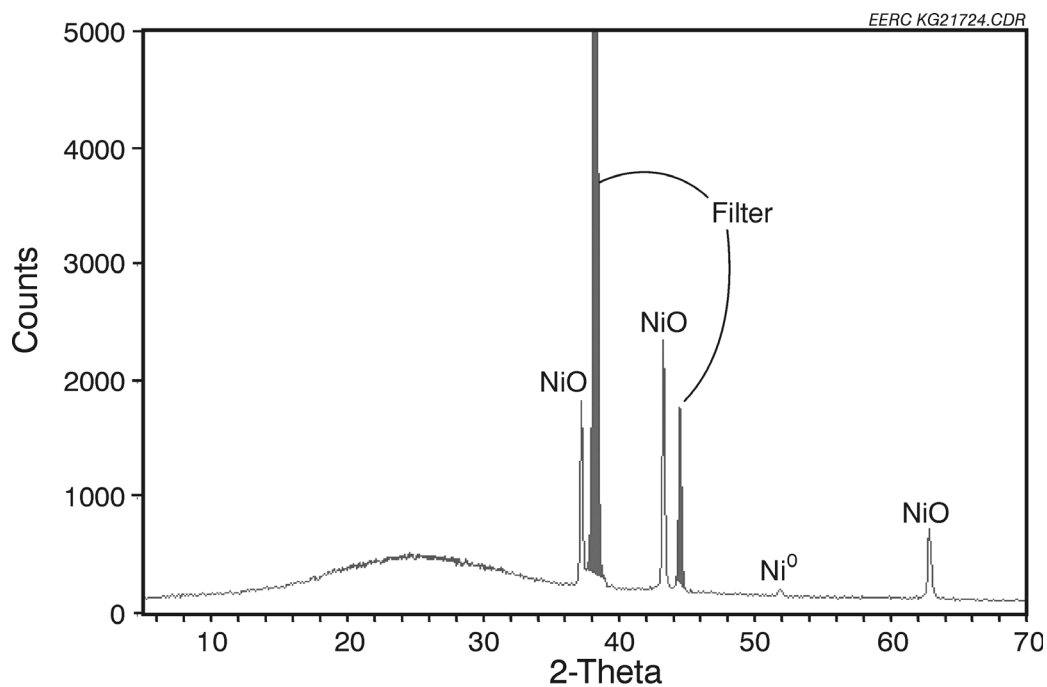


Figure 5. XRD pattern of Ni⁰ extraction residue prepared from a standard NiSO₄ · 6H₂O–Ni₃S₂–Ni⁰–NiO compound mixture.

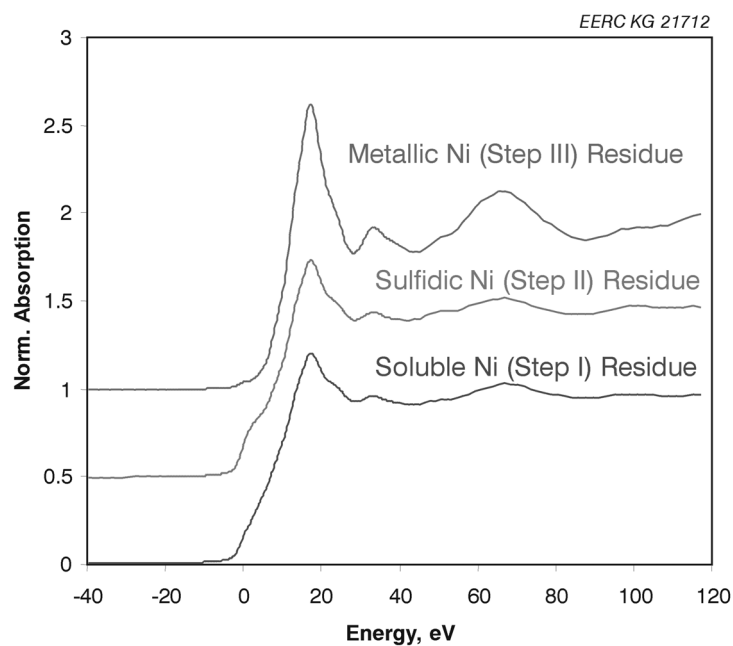


Figure 6. Ni K-edge XANES spectra of three Ni extraction residues prepared from a standard Ni compound mixture. Value of 0 on the abscissa corresponds to the actual Ni K-edge absorption energy of 8333 eV.

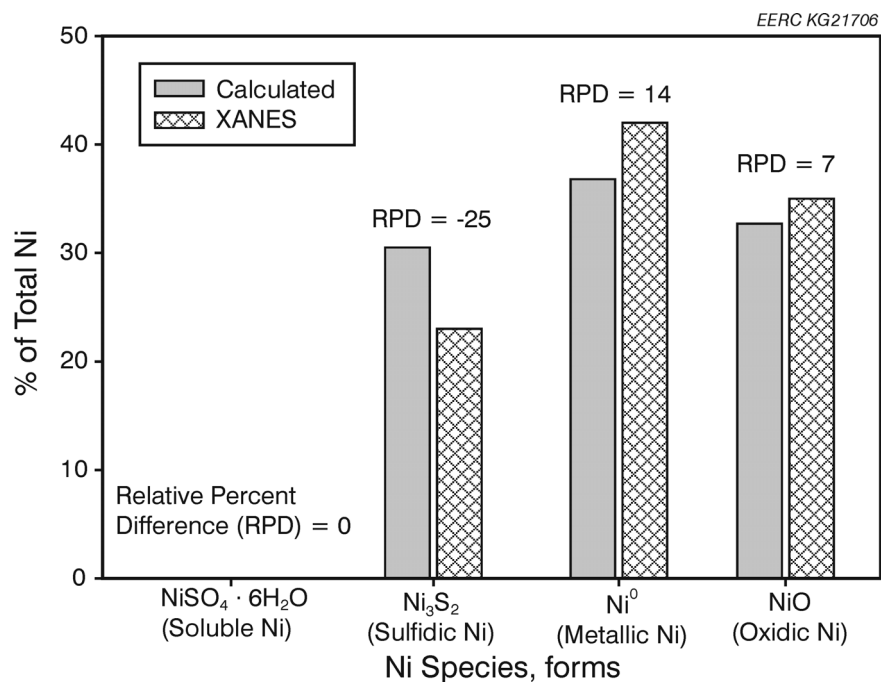


Figure 7. Comparison of calculated and measured (XANES) soluble Ni extraction residue speciation.

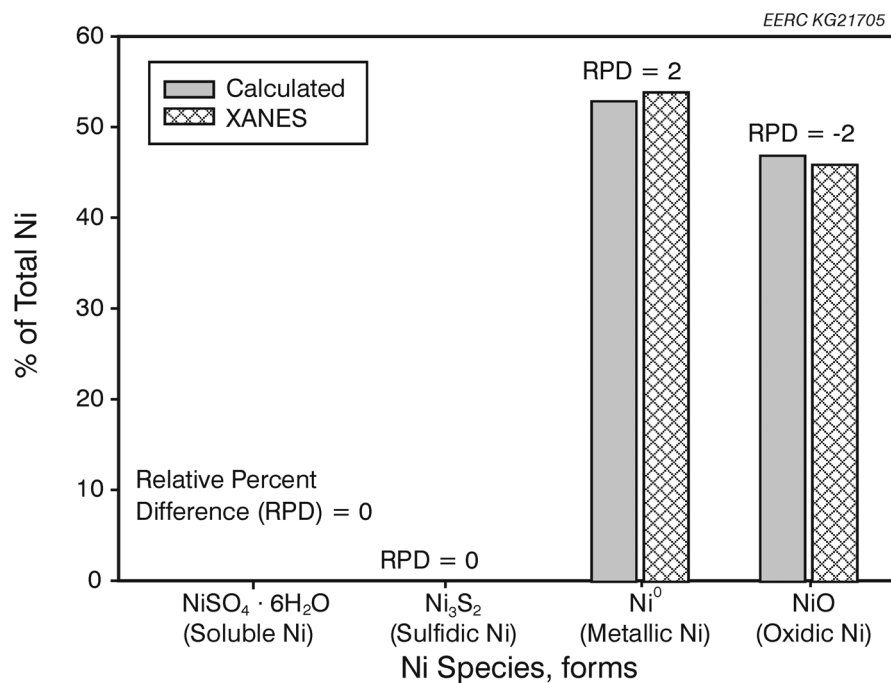


Figure 8. Comparison of calculated and measured (XANES) sulfidic Ni extraction residue speciation.

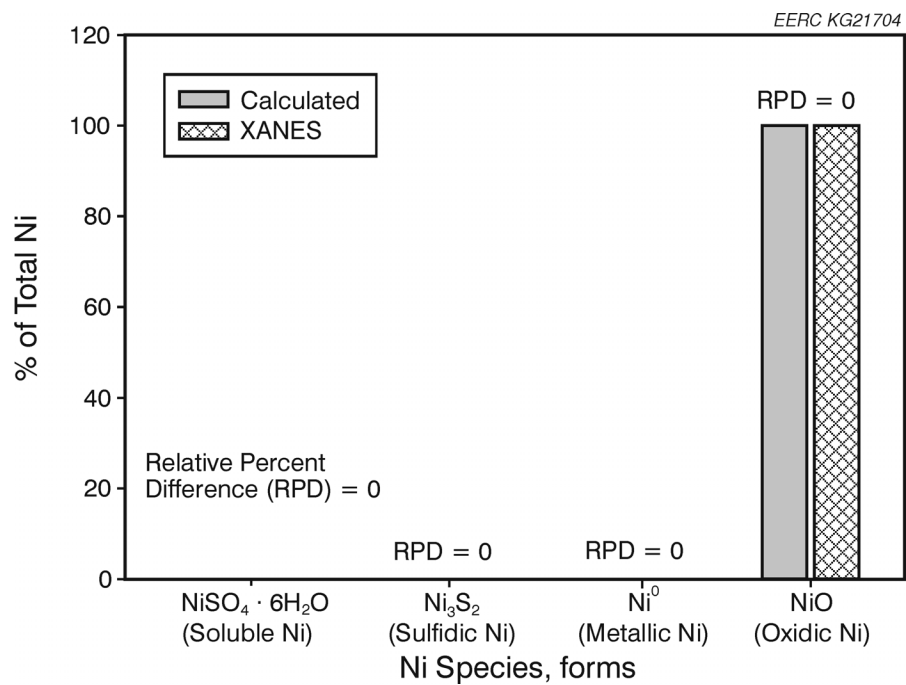


Figure 9. Comparison of calculated and measured (XANES) Ni^0 extraction residue speciation.

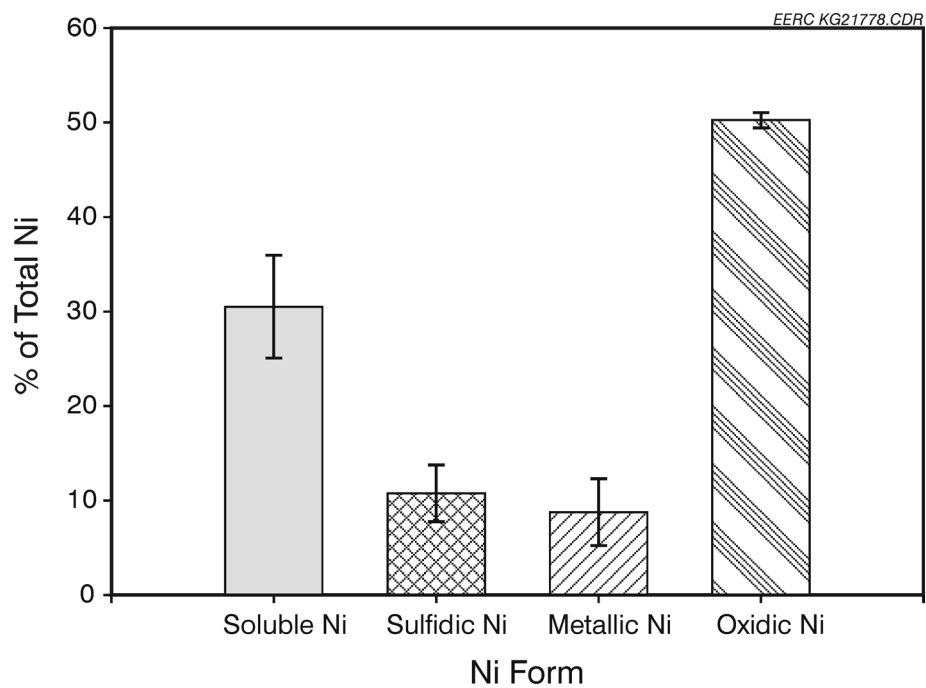


Figure 10. Mean Ni extraction results for SRM 1648. The error bars represent $\pm 95\%$ confidence limits calculated from four sequential Ni extraction measurements.

soluble and oxidic Ni forms account for about 80 % of the total Ni in SRM 1648 and sulfidic and metallic Ni forms account for 20 % of the remaining Ni.

3.2.2 XAFS Spectroscopy Analysis

Three extraction (Steps I–III) residues from SRM 1648 were analyzed using XAFS spectroscopy. However, the quality (signal-to-noise ratio) of the XAFS spectra obtained on the three residues was insufficient for processing and analysis. Mass balance calculations, based on the certified Ni content of SRM 1648, the extraction results presented in Figure 10, and GFAAS analyses of the residues, indicated that Ni was present in concentrations (41–57 ppm) well above the XAFS spectroscopy detection limit of ~5 ppm. Apparently, the residue samples were too sparsely distributed on the filter substrate to obtain high-quality XAFS spectra. Visual inspection of the filters indicated that the residues were not homogeneously distributed which probably also contributed to the lack of XAFS spectroscopy sensitivity.

3.3 Urban (Davie, Florida) TSP, PM₁₀, and PM_{2.5}

3.3.1 Mass Loadings and Ni Concentrations

Filter mass loading information, exemplified in Table 2 for the TSP sample, was used to calculate average concentrations of airborne TSP, PM₁₀, and PM_{2.5} of 26.8, 10.4, and 6.3 µg/m³, respectively. Compared in Table 3 are the average measured (in duplicate) and certified Ni concentrations for SRM 1648 and Ni concentrations for a filter blank and TSP collected at the Davie, Florida, site. The measured Ni concentration for the SRM 1648 is biased 22% high relative to the NIST-certified Ni concentration. The filter blank contained significant Ni but was much less than the Ni concentration of Davie TSP. The Ni concentration of Davie TSP is apparently much greater than SRM 1648. However, after the Ni contamination in the filter blank is accounted for and it is assumed that, similar to SRM 1648 the Ni measurement is biased 22% high, the corrected Ni concentration of Davie TSP presented in Table 3 is very similar to SRM 1648 Ni concentration.

Table 2. High-Volume Aerosol Sampling Parameters

Parameter	Value
Filter Weight, initial	4.43 g
Filter Weight, final	4.69 g
Filter Loading	0.26 g
Total Sample Volume	9,700 m ³
TSP	26.8 µg/m ³

Table 3. Comparison of SRM 1648 and Davie TSP Ni Concentrations, ppm

NIST SRM 1648				Davie, Florida		
	Measured	NIST Certified	RPD ¹	Filter Blank	Urban TSP and Filter	Urban TSP ²
Ni	100	82	22	31	150	86

¹ Relative percent difference.

² Corrected Ni concentration.

3.3.2 XRD Analysis

A blank HVAS filter and the HVAS filter from Davie were analyzed using XRD to identify crystalline components of PM. The XRD pattern presented in Figure 11 indicates that halite (NaCl) and calcite (CaCO₃) are present in Davie TSP. Halite and calcite are known tracers of marine aerosols. The broad, diffuse peak, ranging from about 14° to 36° 2θ, indicates that the TSP is composed largely of amorphous or crystallite (<0.1 μm in diameter) components.

3.3.3 Sequential Ni Extraction Analysis

Assuming a uniform distribution of TSP, the HVAS filter was sectioned into equal portions for subsequent sequential Ni extraction analysis, with each portion representing approximately 22 mg of TSP. Sequential Ni extraction was performed in duplicate on two TSP sections. Method blanks were also analyzed, and all blank results were below GFAAS detection. Presented in Figure 12 are four-step Ni extraction results for the urban TSP collected in Davie, Florida. These

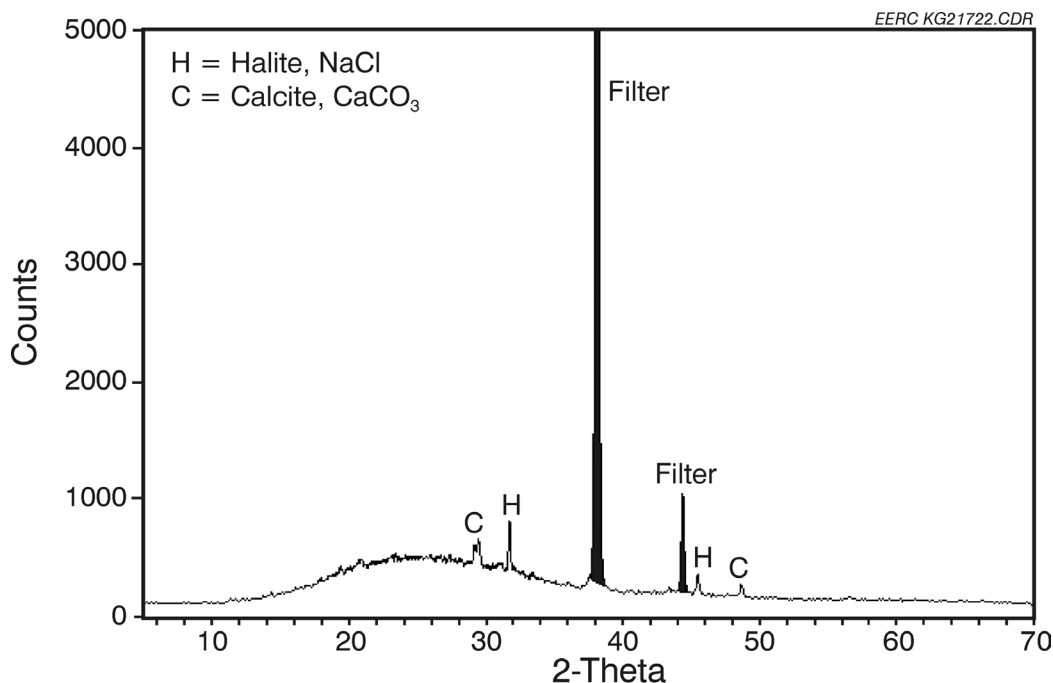


Figure 11. XRD pattern of HVAS filter containing TSP from Davie, Florida.

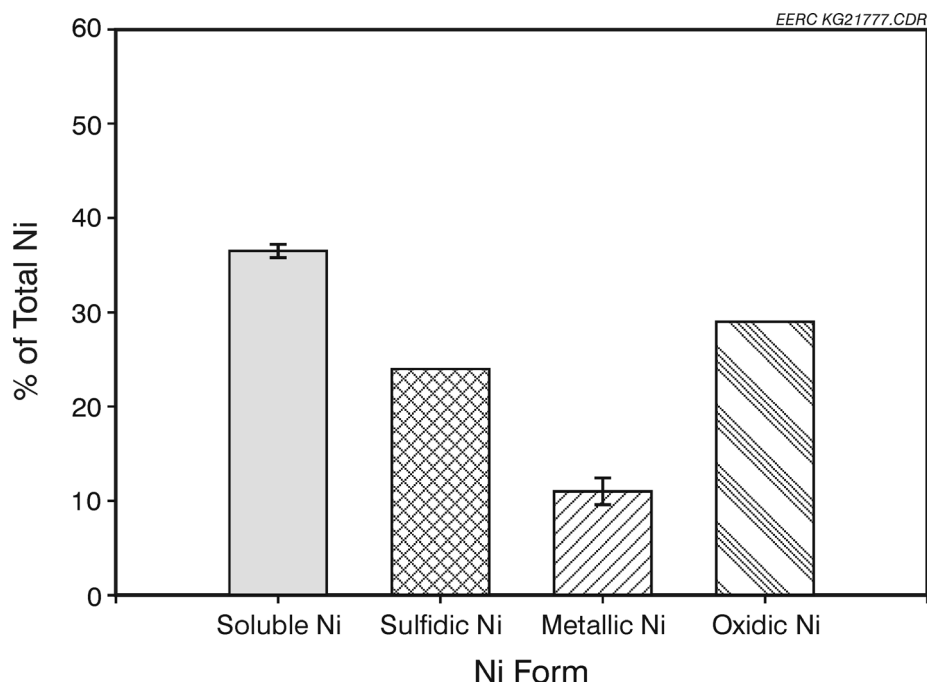


Figure 12. Average of duplicate Ni extraction results for urban TSP collected in Davie, Florida.

extraction results indicate that soluble, oxidic, and sulfidic Ni forms account for most of the Ni in Davie TSP. Ni^0 accounts for about 10% of the remaining total Ni. In comparison to SRM 1648 (Figure 10), Davie TSP contains much greater proportions of soluble- and sulfidic-Ni and less oxidic Ni.

3.3.4 Sulfur K-Edge XAFS Spectroscopy Analyses

Presented in Figure 13 are unprocessed sulfur K-edge XAFS spectra for the ACCU (PM_{10}) and SAS ($\text{PM}_{2.5}$) filter blanks and filters containing Davie TSP, PM_{10} , and $\text{PM}_{2.5}$. A spectrum for a blank filter from the HVAS is not presented in Figure 13 because it was inadvertently not analyzed for sulfur using XAFS spectroscopy. Although the PM_{10} - and $\text{PM}_{2.5}$ -containing filters exhibit much smaller x-ray absorption intensities than the TSP-containing filter, the absorption intensities for these samples are significantly greater than the absorption signals for the corresponding blank filters. Hence, the sulfur absorptions are primarily caused by the PM_{10} and $\text{PM}_{2.5}$ trapped within the filters. Sulfur XANES spectra for the SAS filter blank and the filters containing TSP, PM_{10} , and $\text{PM}_{2.5}$ are presented in Figure 14. An XANES spectrum for the ACCU filter blank is not included in Figure 14 because the sulfur K-edge absorption signal was too weak to be processed. The XANES features in Figure 14 are insensitive to differences in sulfur concentration because of the data reduction process used to create the XANES spectra. The energy location of the sulfur peaks in Figure 14 are all consistent with the presence of the sulfate (SO_4) species. The XANES sulfur species detection limit is approximately 3% of the total sulfur present, thus >97% of the sulfur in Davie TSP, PM_{10} , $\text{PM}_{2.5}$, and SAS filter blank is present as SO_4 -based compounds. An ion chromatography analysis of the Step I and II (Table 1) extracts from Davie TSP indicated that it contains 1.61 wt% SO_4 .

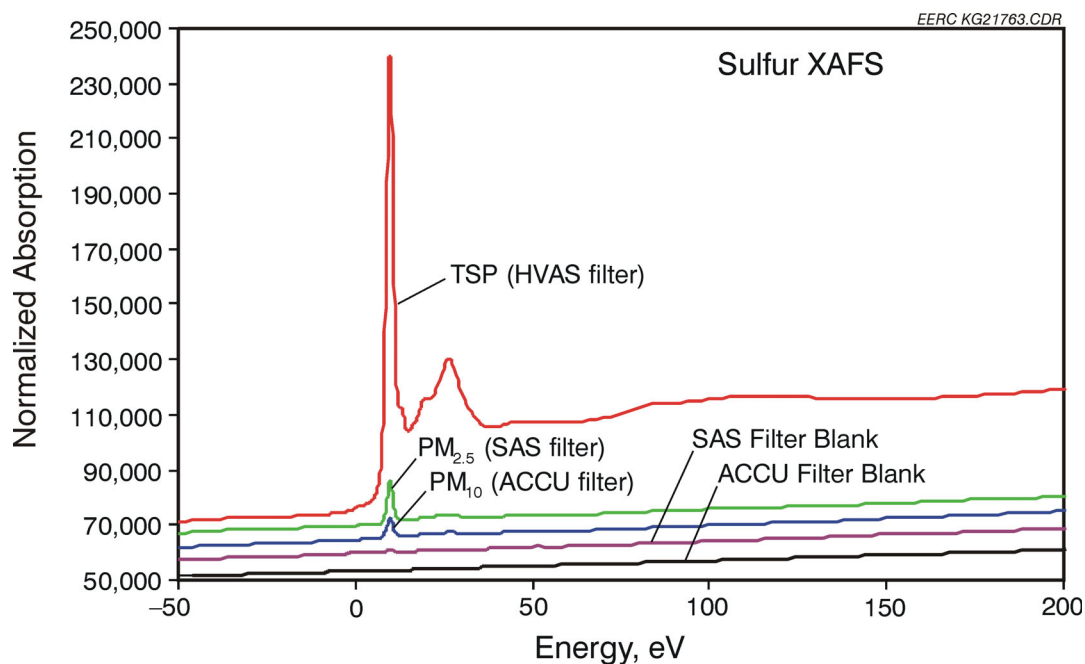


Figure 13. Sulfur K-edge XAFS spectra of ACCU (PM₁₀) and SAS (PM_{2.5}) filter blanks and filters containing Davie TSP, PM₁₀, and PM_{2.5}. Value of 0 on the abscissa corresponds to the actual sulfur K-edge absorption energy of 2472 eV.

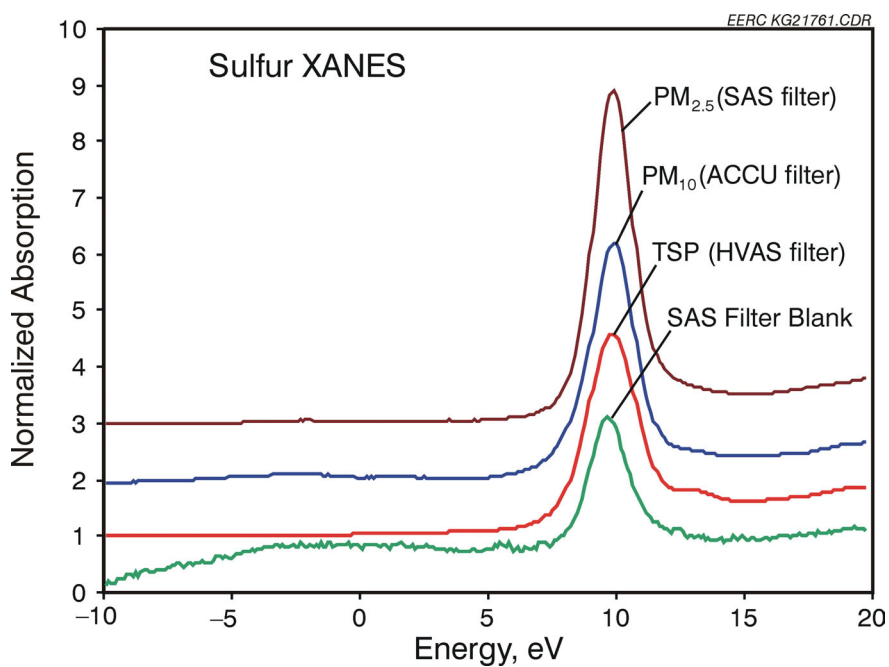


Figure 14. Sulfur XANES spectra of SAS filter blank and filters containing Davie TSP, PM₁₀, and PM_{2.5}. Value of 0 on the abscissa corresponds to the actual S K-edge absorption energy of 2472 eV.

3.3.5 Nickel K-Edge XAFS Spectroscopy Analyses

Presented in Figure 15 are Ni K-edge XAFS spectra for the HVAS (TSP), ACCU (PM₁₀), and SAS (PM_{2.5}) filter blanks and filters containing Davie TSP, PM₁₀, and PM_{2.5}. Also included in Figure 15 is a Ni K-edge XAFS spectrum collected when no sample was in the x-ray beam path. The relative Ni absorption intensity relationships are PM₁₀ > TSP > PM_{2.5} \approx filter and beam blanks indicating that Ni is mostly concentrated in PM₁₀. Duplicate GFAAS analyses of Davie TSP indicated that it contains 86 ppm Ni (Table 3). Ni absorption is generally proportional to Ni concentration; thus the Davie PM₁₀ sample is estimated to contain approximately 140 ppm Ni. These Davie TSP and PM₁₀ Ni concentrations correspond to airborne Ni concentrations of 2.3 and 1.5 ng/m³, respectively, which are very similar to the annual average ambient Ni concentration of 1.9 ng/m³ estimated for Broward County by EPA.³⁷ These Ni concentrations are much lower in comparison to average urban ambient air Ni concentrations which range from 5 to 50 ng/m³ and are more representative of Ni concentrations found in remote and rural areas.⁹⁻¹² The low airborne Ni concentrations at the Davie, Florida, air-monitoring site probably represent the atmospheric burden of “background” Ni derived from long-range transport via the southeast and easterly North Atlantic trade winds (Figure 1).

The Ni K-edge absorption signals from the filters containing Davie TSP and PM₁₀ were sufficiently intense to process and create the XANES spectra presented in Figure 16. Spectra from the Ni XANES database, described previously in the experimental section of this report, were compared to the spectra shown in Figure 16 using a least-squares fitting procedure. Estimates, based

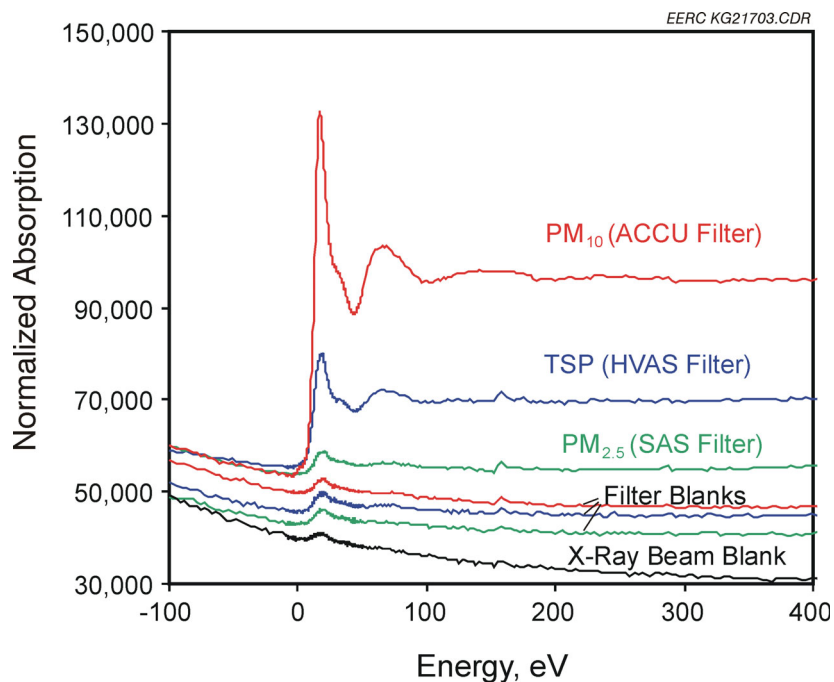


Figure 15. Ni K-edge XAFS spectra of HVAS, ACCU, and SAS filter blanks and filters containing Davie TSP, PM₁₀, and PM_{2.5}. Value of 0 on the abscissa corresponds to the actual Ni K-edge absorption energy of 8333 eV.

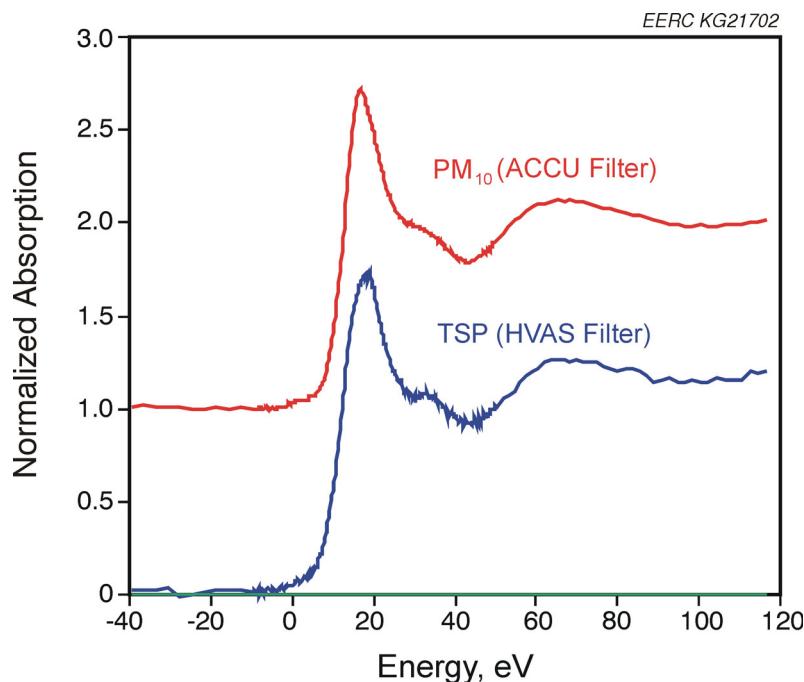


Figure 16. Ni K-edge XANES spectra of the filters containing Davie TSP and PM₁₀. Value of 0 on the abscissa corresponds to the actual Ni K-edge absorption energy of 8333 eV.

on the least-squares fitting, of the Ni species compositions of Davie TSP and PM₁₀ are provided in Table 4. Davie PM₁₀ contains a much greater proportion of NiSO₄ · xH₂O relative to TSP, probably because of the greater surface area of PM₁₀ on which NiSO₄ · xH₂O heterogeneously condensed or adsorbed.

Table 4. XANES Ni Speciation Results for Davie TSP and PM₁₀, % of total Ni

Ni Species	TSP	PM ₁₀
NiSO ₄ · xH ₂ O	40	78
NiFe ₂ O ₄	50	22
NiS	10	<5

The Ni speciation results in Table 4 were used to calculate the concentrations of Ni compounds in Davie TSP and PM₁₀ and ambient air. The calculations were performed assuming pure stoichiometric nickel hexahydrite (NiSO₄ · 6H₂O), NiFe₂O₄, and NiS compounds. Results of the calculations are presented in Table 5. The results in Table 5, especially those for PM₁₀, can be used to estimate an inhalation dose and are, therefore, more useful than those in Table 4 for evaluating the potential health risks associated with individual Ni species.

Table 5. Ni Species Concentrations for Davie TSP and PM₁₀, ppm

Ni Species	TSP		PM ₁₀	
	ppm	ng/m ³	ppm	ng/m ³
NiSO ₄ · 6H ₂ O	154	4.1	489	5.1
NiFe ₂ O ₄	172	4.6	123	1.3
NiS	13	0.4	<11	<0.1

Compared in Figure 17 are sequential Ni extraction and XANES analysis results for Davie TSP. The proportion of NiSO₄ · xH₂O in Davie TSP measured by XANES is similar to the proportion of soluble Ni measured by the Ni extraction method. In contrast to the four-step Ni extraction results for Davie TSP, the XANES results indicate a much lower proportion of sulfidic Ni as NiS and a much greater proportion of oxidic Ni, possibly in the form of NiFe₂O₄. The XANES method did not detect any Ni⁰ which was estimated by the Ni extraction method to comprise about 10% of the total Ni. Contrary to XAFS spectroscopy results for the Ni compound mixture residue samples, these results suggest that oxidic Ni in Davie TSP was extracted during the sulfidic Ni and Ni⁰ (second and third) extraction steps rather than in the last step. Scanning electron microscopy analyses indicated that Ni-bearing particles in the model Ni compound mixture are much larger relative to those found in Davie TSP, suggesting that the sequential Ni extraction method overestimates sulfidic Ni and Ni⁰ fractions when smaller particles are present.

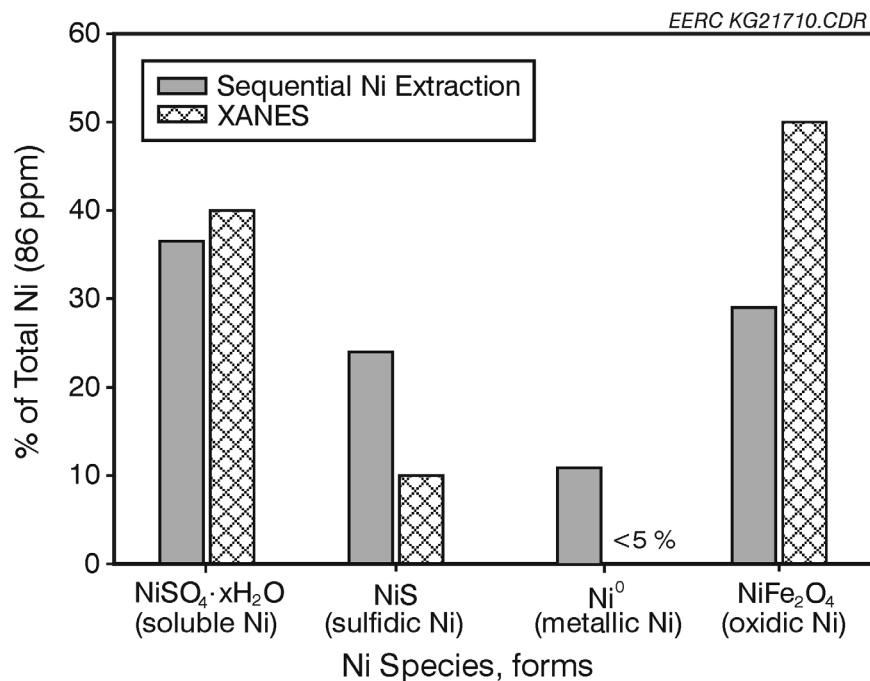


Figure 17. Comparison of sequential Ni extraction and XANES analysis results for Davie TSP.

The soluble Ni extraction procedure (Table 1) was performed on a filter blank and on a representative portion of the Davie TSP filter sample. Presented in Figure 18 are Ni K-edge XAFS spectra for an HVAS filter blank, an HVAS filter blank from the soluble Ni extraction, and Davie TSP residue from the soluble Ni extraction procedure. Spectra for both filter blanks are similar, and the Ni K-edge absorption signal for the Davie TSP residue is only about three times more intense relative to the corresponding filter blank. Consequently, spectral processing resulted in a poor-quality XANES spectrum for the Davie TSP residue sample that was insufficient for quantitative analysis.

3.3.6 ASEM

Urban TSP deposited on portions of the carbon tape, corresponding to a daily time period of 8:00 a.m.–5:00 p.m., of the Burkard Seven-Day Recording Volumetric Spore Trap were analyzed using ASEM. Approximately 800–1100 particles were analyzed at random on the samples collected on August 27–30. Compared in Figure 19 are the number concentrations of Ni-bearing particles, arbitrarily defined as containing ≥ 10 wt% Ni, sampled during each of 4 days (August 27–30, 2002). Approximately 8% of the particles sampled on August 27, 29, and 30 contained ≥ 10 wt% Ni, whereas on August 28, about 15% of the particles were Ni bearing. Apparently, additional emission sources of Ni were contributing to the TSP sample collected on August 28 or the Ni emission rate for a particular source in the Davie, Florida, area increased on August 28.

The particle-size distributions of Ni-bearing particles are compared in Figure 20. Most of the Ni-bearing particles are <10 μm in diameter, indicating that they were released from combustion

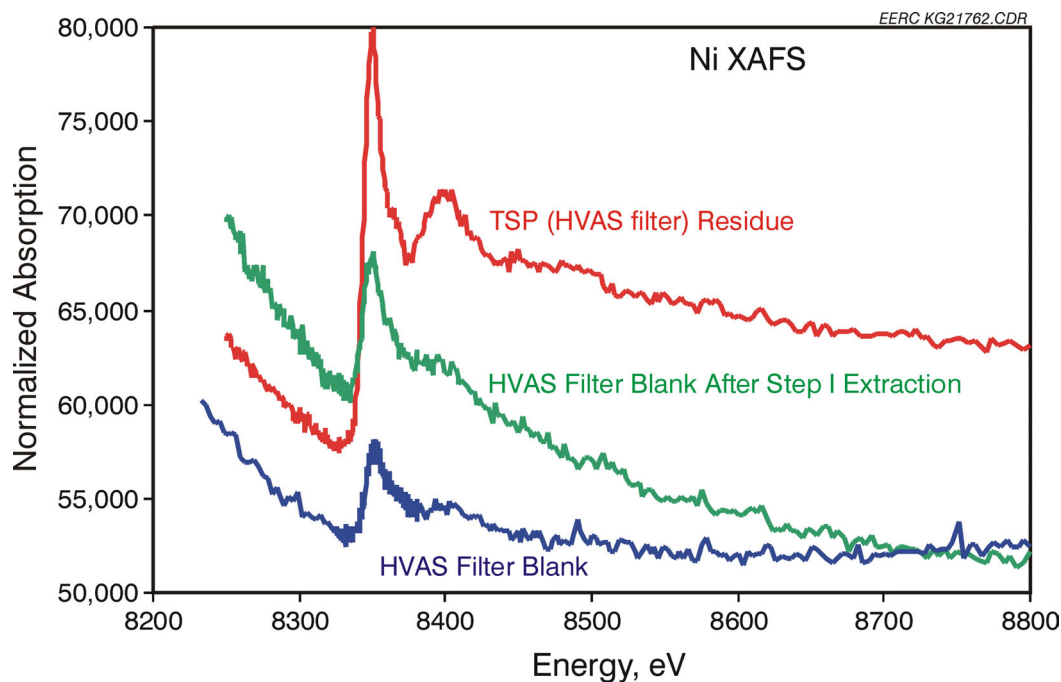


Figure 18. Ni K-edge XAFS spectra of a HVAS filter blank, HVAS filter blank from the soluble Ni (Step I) extraction, and a HVAS filter containing Davie TSP residue from the soluble Ni (Step I) extraction.

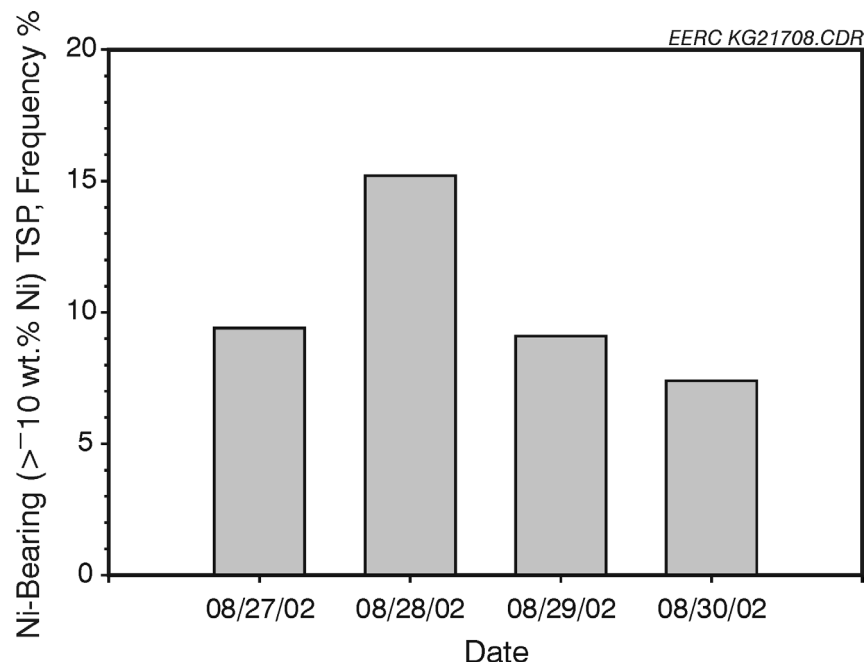


Figure 19. Number concentrations of Ni-bearing (≥ 10 wt %) particles in Davie TSP samples.

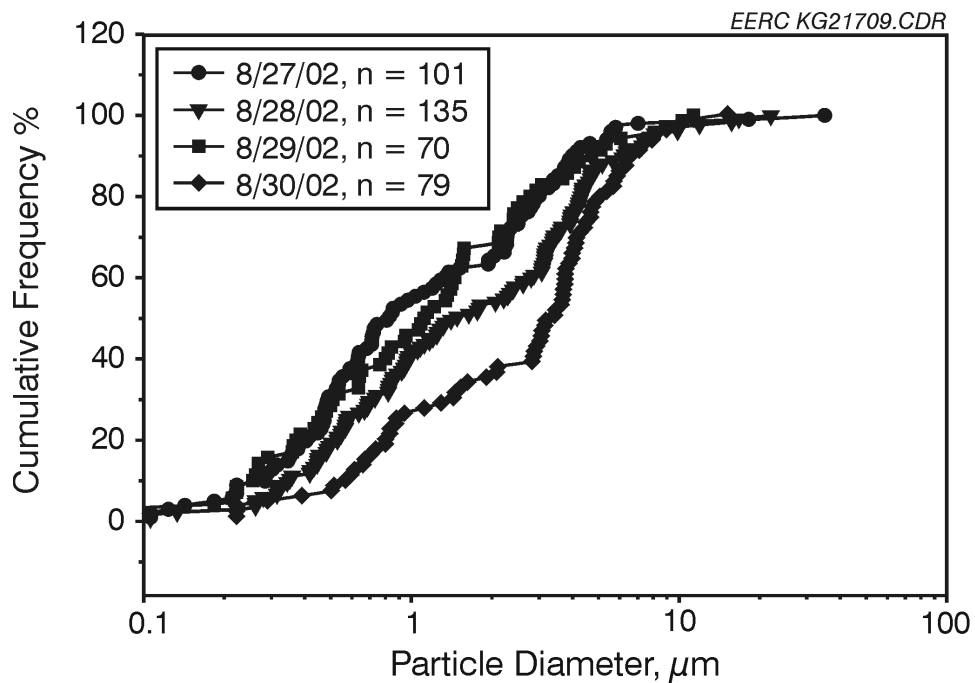


Figure 20. Particle-size distribution of Ni-bearing (≥ 10 wt %) particles in Davie TSP.

sources. Ni-bearing particles sampled on August 30 were significantly coarser than those sampled on August 27–29.

4.0 CONCLUSIONS

- Chemical mass balance considerations, XRD, and XANES analyses indicated that the soluble Ni extraction procedure (Step I) of Luk et al.¹⁸ provided a reliable estimate of $\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$ concentration in a $\text{NiSO}_4 \cdot 6\text{H}_2\text{O}$ – Ni_3S_2 – Ni^0 – NiO compound mixture but overestimated Ni^0 and underestimated Ni_3S_2 and NiO concentrations.
- Sulfur K-edge XANES analyses indicated that >97% of the total sulfur in Davie, Florida, TSP, PM_{10} , and $\text{PM}_{2.5}$ is present as SO_4 .
- The airborne Ni concentrations of 2.3 and 1.5 ng/m^3 measured for TSP and PM_{10} , respectively, at the Davie, Florida, SLAMS/NAMS site during August 26–31, 2002, is similar to the EPA estimated annual average ambient Ni concentration of 1.9 ng/m^3 for Broward County.³⁷
- Sequential Ni extraction and XANES results indicated that $\text{NiSO}_4 \cdot x\text{H}_2\text{O}$ and a nickel oxide compound, possibly NiFe_2O_4 , were the dominant Ni species occurring in Davie, Florida, TSP. Both methods also indicated that a small proportion of NiS was present.
- XAFS spectroscopy analyses indicated that the sequential Ni extraction procedure of Luk et al.¹⁸ selectively and quantitatively removed $\text{NiSO}_4 \cdot x\text{H}_2\text{O}$ from Davie, Florida, TSP but overestimated the sulfidic Ni and Ni^0 fractions and underestimated oxidic Ni.
- Davie, Florida, PM_{10} lacked NiS , but $\text{NiSO}_4 \cdot x\text{H}_2\text{O}$ was more abundant in comparison to TSP, probably because of the greater surface area of PM_{10} on which $\text{NiSO}_4 \cdot x\text{H}_2\text{O}$ heterogeneously condensed or adsorbed.
- An insufficient amount of $\text{PM}_{2.5}$ was sampled from Davie, Florida, to quantify Ni and perform speciation analyses.
- Approximately 8%–15% of the total number of particles sampled from the Davie, Florida, SLAMS/NAMS site contained significant Ni concentrations, ≥ 10 wt%. Most of these Ni-bearing particles were <10 μm in diameter.
- Direct metal speciation techniques such as XAFS spectroscopy should be used to verify the applicability and reliability of sequential Ni extraction methods before they are employed on a specific sample type (raw materials, workplace dust, ambient air TSP, PM_{10} , $\text{PM}_{2.5}$, etc.).
- The limitations of sequential Ni extraction methods applied to urban PM samples need to be considered when using such indirect Ni speciation results to assess the potential inhalation health risks associated with individual Ni species.

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APPENDIX A

SEQUENTIAL NICKEL EXTRACTION RESULTS

Table A-1. Ni Extraction Recoveries for Model Ni-Compound Mixtures

Sample	Soluble Ni (NiSO ₄)		Sulfidic Ni (Ni ₃ S ₂)		Metallic Ni (Ni ⁰)		Oxidic Ni (NiO)	
	Expected, µg	Recovery, %	Expected, µg	Recovery, %	Expected, µg	Recovery, %	Expected, µg	Recovery, %
A	883	101	3032	75	4257	122	2885	67
B	873	118	2996	80	4207	115	2851	90
C	873	113	2996	83	4207	110	2851	88
D	883	101	3032	90	4257	129	2885	59
E	690	100	2369	73	3327	124	2255	68
F	720	116	2472	79	3471	108	2352	76
G	684	110	2347	82	3295	114	2233	91
H	679	100	2330	76	3270	113	2217	69
Average	786	107	2697	80	3786	117	2566	76
STD ¹	100	8	342	5	480	7	325	12
RSD, ² %	13	7	13	7	13	6	13	16

¹ Standard deviation.

² Relative standard deviation.

Table A-2. Quadruplicate Sequential Ni Extraction Results for Urban Particulate Matter (NIST SRM 1648)

	Soluble Ni, µg	Sulfidic Ni, µg	Ni ⁰ , µg	Oxidic Ni, µg	Total Ni, µg	Expected Recovery, ¹ µg	Recovery, %
A	0.74	0.24	0.46	1.44	2.88	3.19	90
B	1.02	0.42	0.26	1.68	3.38	3.39	100
C	1.05	0.36	0.24	1.63	3.28	3.80	86
D	1.07	0.24	0.24	1.59	3.14	3.41	92
Average	0.97	0.32	0.30	1.58	3.17	3.45	92
STD ²	0.15	0.09	0.11	0.10	0.22	0.26	5.9
RSD, ³ %	15	28	37	6.3	6.9	7.5	6.4

¹ Expected recovery is based on a nominal 40-mg sample of SRM 1648 with a certified value of 82±3 µg/g Ni.

² Standard deviation.

³ Relative standard deviation.

Table A-3. Quadruplicate Sequential Ni Extraction Results for SRM 1648 (Urban Particulate Matter), % of total Ni

	Soluble Ni	Sulfidic Ni	Ni ⁰	Oxidic Ni
A	26	12	12	50
B	30	12	8	50
C	32	11	7	50
D	34	8	8	51
Average	31	11	9	50
STD ¹	3.4	1.9	2.2	0.5
RSD, ² %	11	18	25	1.0

¹ Standard deviation.

² Relative standard deviation.

Table A-4. Duplicate Sequential Ni Extraction Results for Fort Lauderdale Bulk Urban PM

Soluble Ni, µg	Sulfidic Ni, µg	Ni ⁰ , µg	Oxidic Ni, µg	Total Ni, µg	Expected total Ni, ¹ µg	Recovery, %
1.72	1.10	0.48	1.33	4.63	4.40	105
1.70	1.12	0.55	1.37	4.74	4.40	108

Table A-5. Duplicate Sequential Ni Extraction Results for Fort Lauderdale Bulk Urban PM, % of total Ni

	Soluble Ni	Sulfidic Ni	Ni ⁰	Oxidic Ni
A	37	24	10	29
B	36	24	12	29
Average	37	24	11	29
STD ¹	0.7	0.0	1.4	0.0
RSD, ² %	1.9	0.0	12.9	0.0

¹ Standard deviation.

² Relative standard deviation.