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**Decontamination of HEPA Filters:  
July-September 1977**

**Don F. Luthy and Edward L. Lewis**

**February 20, 1978** P



**Monsanto**

**MOUND FACILITY**

Miamisburg, Ohio

operated by

**MONSANTO RESEARCH CORPORATION**

a subsidiary of Monsanto Company

for the

**U. S. DEPARTMENT OF ENERGY**

Contract No. EY-76-C-04-0053

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# Abstract

Dissolution parameters (in various reagents) of americium-241 and plutonium-239 oxide mixtures, uranium-233 oxide, neptunium-237 oxide, curium-244 oxide, thorium-232 oxide, and plutonium dioxide were determined. The reagents used were various concentrations of  $\text{HNO}_3$ -HF- $\text{H}_2\text{SO}_4$ ,  $\text{HNO}_3$ -( $\text{NH}_4$ ) $_2\text{Ce}(\text{NO}_3)_6$ ,  $\text{HNO}_3$ -HF, and  $\text{HNO}_3$ . Both simulated contaminated HEPA filter media and actual glovebox filter media from spent filters were used. The maximum decontamination factor achieved was 833 using a six-stage dissolution process. Also, plutonium dioxide was fused with sodium carbonate at various elevated temperatures, and a dissolution percentage was determined.

## Experimental

### I. Americium-241 and plutonium-239\* mixed oxide dissolution studies

Contaminated filter media were prepared by mixing  $\text{AmO}_2$ - $\text{PuO}_2$  powder with shredded filter media. Small samples of this mixture were then treated with various leaching agents, including 4N  $\text{HNO}_3$ -0.1M ( $\text{NH}_4$ ) $_2\text{Ce}(\text{NO}_3)_6$ , 12N  $\text{HNO}_3$ -0.05N HF-0.01N  $\text{H}_2\text{SO}_4$ , and 8N  $\text{HNO}_3$ . A sample of filter media was placed in a beaker, 250 ml of the leaching agent was added, and the temperature was increased to the boiling point of the solution. Samples were withdrawn periodically and filtered through a 4-5  $\mu\text{m}$  glass filter. The americium-241 and plutonium-239 concentrations were determined by alpha spectrometry. The solution volume and concentration were kept constant by adding acid of the proper concentration to replace evaporated acid. As can be seen in Table 1, the 4N  $\text{HNO}_3$ -0.1M ( $\text{NH}_4$ ) $_2\text{Ce}(\text{NO}_3)_6$  was an effective leaching agent, dissolving greater than

90% of both  $\text{AmO}_2$  and  $\text{PuO}_2$  in 11-3/4 hr. On the other hand, the 8N  $\text{HNO}_3$  was not a satisfactory leaching agent since only 36% of the  $\text{PuO}_2$  dissolved in 13-1/2 hr, although 97% of the  $\text{AmO}_2$  dissolved during this time (see Figure 1). The mixture of  $\text{HNO}_3$ -HF- $\text{H}_2\text{SO}_4$  was effective in dissolving the  $\text{AmO}_2$  (90% in 11-1/4 hr) but not as effective in dissolving the  $\text{PuO}_2$  (80% in 11-1/4 hr). Of the three reagents tested, the 4N  $\text{HNO}_3$ -0.1M ( $\text{NH}_4$ ) $_2\text{Ce}(\text{NO}_3)_6$  is the recommended leaching agent for mixtures of  $\text{AmO}_2$ - $\text{PuO}_2$ .

### II. Uranium-233 oxide dissolution studies

These studies were conducted using  $\text{U}_3\text{O}_8$  mixed with filter media. The oxide had been fired at 950°C for 2 hr. A small sample of the prepared filter media (approximately 3.2 g) was placed into a beaker containing 250 ml of boiling reagent. The reagents used were 8N  $\text{HNO}_3$ , 4N  $\text{HNO}_3$ -0.1M ( $\text{NH}_4$ ) $_2\text{Ce}(\text{NO}_3)_6$ , and 12N  $\text{HNO}_3$ -0.05N HF-0.01N  $\text{H}_2\text{SO}_4$ . Samples were withdrawn periodically for alpha analysis. All samples were filtered through 4-5  $\mu\text{m}$  glass filters. The solution volume and concentration were kept constant by adding acid of the proper concentration to replace evaporated acid. As can be seen from

\*This material was approximately three parts plutonium-239 oxide mixed with one part americium-241 oxide by weight. It had been fired for 2 hr at 1000°C.

Table 1 - AmO<sub>2</sub>-PuO<sub>2</sub> DISSOLUTION TESTS

Acid Test Number	Reagent Composition	Time Heated (hr)	PuO <sub>2</sub> Dissolved (wt%)	AmO <sub>2</sub> Dissolved (wt%)
Am-Pu-1A	4N HNO <sub>3</sub> -0.1M CAN <sup>a</sup>	5 3/4	72.9	63.2
Am-Pu-1B	4N HNO <sub>3</sub> -0.1M CAN <sup>a</sup>	11 3/4	93.7	111.8
Am-Pu-2A	12N HNO <sub>3</sub> -0.05N HF-0.01N H <sub>2</sub> SO <sub>4</sub>	5 3/4	68.9	72.4
Am-Pu-2B	"	11 1/4	80.2	90.4
Am-Pu-4A	8N HNO <sub>3</sub>	7 1/4	37.2	88.9
Am-Pu-4B	"	13 1/2	36.3	96.7

<sup>a</sup>CAN is an acronym for ceric ammonium nitrate.

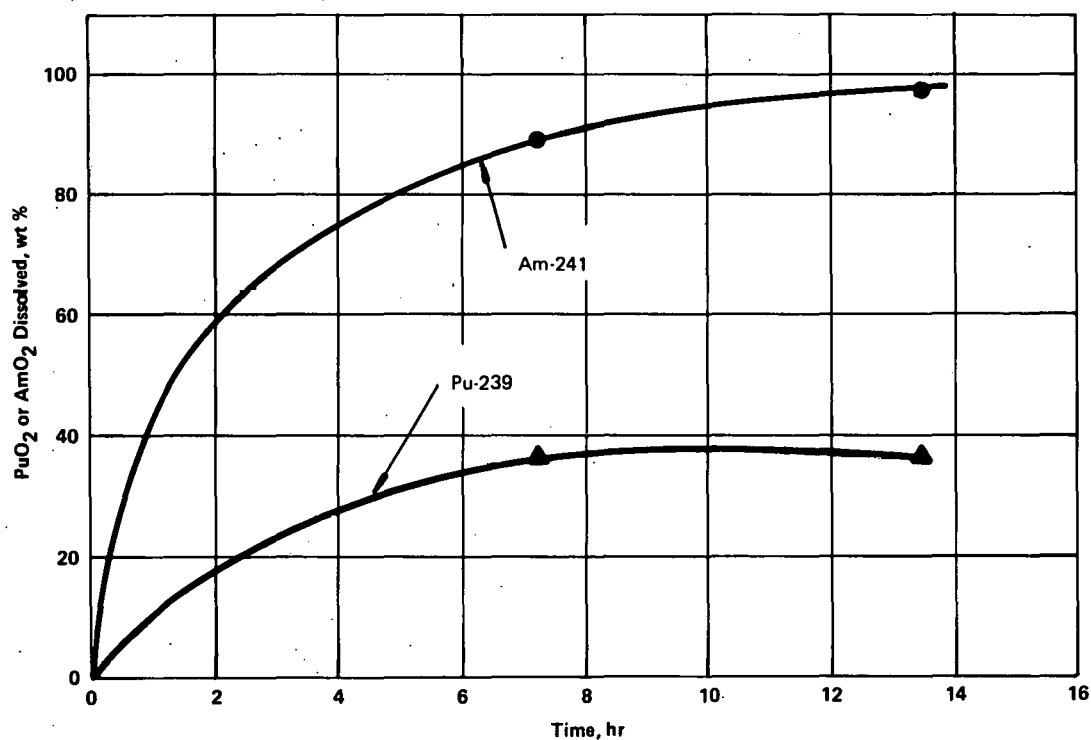


FIGURE 1 - Dissolution rate of Am<sup>241</sup>-Pu<sup>239</sup> Oxides in 8N HNO<sub>3</sub>



Table 2, in all three cases the dissolution was very rapid and exceeded 90%. For example, the dissolution rate in 8N HNO<sub>3</sub> is depicted in Figure 2. All three reagents would be acceptable dissolution reagents for uranium-233 oxide.

### III. Sodium carbonate fusion of PuO<sub>2</sub>\*

These studies were conducted in order to determine the effect of temperature on fusion efficiency. Small samples were prepared by mixing 64 mg of PuO<sub>2</sub> with 2.25 g of Na<sub>2</sub>CO<sub>3</sub>. The PuO<sub>2</sub>-Na<sub>2</sub>CO<sub>3</sub> mixture was then placed into a platinum crucible and contents were then heated to

\*The PuO<sub>2</sub> was fine powder having a composition of 80 wt% Pu-238, 16% Pu-239, 2.5% Pu-240, 0.8% Pu-241, 0.2% Pu-242, plus small amounts of other actinides. It was calcined for 2 hr at 950°C.

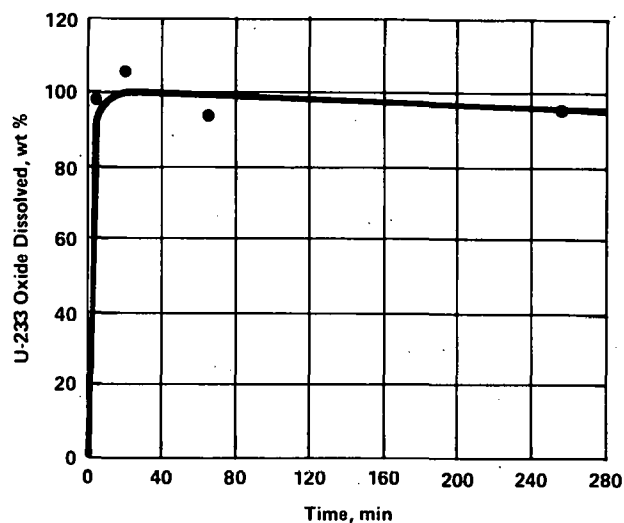


FIGURE 2 - Dissolution rate of U<sub>3</sub>O<sub>8</sub> in 8N HNO<sub>3</sub>.

Table 2 - U-233 OXIDE DISSOLUTION TESTS

Acid Test Number	Reagent Composition	Time Heated (min)	U-233 Oxide Dissolved (wt%)
U-3-1	8N HNO <sub>3</sub>	5	97.4
U-3-2	"	20	106.1
U-3-3	"	65	93.8
U-3-4	"	255	95.9
U-4-1	4N HNO <sub>3</sub> -0.1M CAN <sup>a</sup>	5	89.7
U-4-2	"	15	104.8
U-4-3	"	60	95.2
U-4-4	"	360	95.7
U-5-1	12N HNO <sub>3</sub> -0.05N HF-0.01N H <sub>2</sub> SO <sub>4</sub>	5	100.9
U-5-2	"	15	106.2
U-5-3	"	50	100.1
U-5-4	"	240	102.5

<sup>a</sup>CAN is an acronym for ceric ammonium nitrate.

the desired operating temperature (1200 and 1300°C) and maintained there for 1 hr. The crucible and melt were cooled to room temperature, and the melt was removed from the crucible and dissolved in 4N HNO<sub>3</sub>. As can be seen in Table 3, there is not a significant difference (4.6%) in the two dissolution percentages and it is probable that the difference is due to experimental error. Therefore, it can be concluded that within the temperature range under investigation, the fusion efficiency is approximately 58% (average value).

#### IV. Neptunium-237 oxide dissolution studies

The NpO<sub>2</sub> was fired for 2 hr at 950°C. Samples of contaminated filter media were prepared by mixing 0.2 g of NpO<sub>2</sub> with 3 g of filter media. The procedure was identical to that used in the uranium-233 oxide dissolution studies (Section II). Three leaching reagents were tested, 8N HNO<sub>3</sub>, 4N HNO<sub>3</sub>-0.1M (NH<sub>4</sub>)<sub>2</sub>Ce(NO<sub>3</sub>)<sub>6</sub>, and 12N HNO<sub>3</sub>-0.05N HF-0.01N H<sub>2</sub>SO<sub>4</sub>. Table 4 lists the results from these experiments. As can

Table 3 - Na<sub>2</sub>CO<sub>3</sub> FUSIONS OF PuO<sub>2</sub>

Fusion Test Number	Time Heated (hr)	Operating Temperature (°C)	PuO <sub>2</sub> Solubilized (wt%)	Dissolution Reagent
Pu-44	1	1300°C	56.4	4N HNO <sub>3</sub>
Pu-45	1	1200°C	61.0	4N HNO <sub>3</sub>

Table 4 Np-237 OXIDE DISSOLUTION TESTS

Acid Test Number	Reagent Composition	Time Heated (min)	NpO <sub>2</sub> Dissolved (wt%)
Np-3-1	8N HNO <sub>3</sub>	10	1.3
Np-3-2	"	25	1.9
Np-3-3	"	45	4.1
Np-3-4	"	75	6.1
Np-3-5	"	480	32.3
Np-4-1	4N HNO <sub>3</sub> -0.1M CAN <sup>a</sup>	10	93.7
Np-4-2	"	25	97.2
Np-4-3	"	40	88.9
Np-4-4	"	60	89.1
Np-5-1	12N HNO <sub>3</sub> -0.05N HF-0.01N H <sub>2</sub> SO <sub>4</sub>	10	64.7
Np-5-2	"	25	105.2
Np-5-3	"	45	107.4
Np-5-4	"	100	102.3

<sup>a</sup>CAN is an acronym for ceric ammonium nitrate.

be seen, the 8N  $\text{HNO}_3$  was unsuccessful in dissolving the  $\text{NpO}_2$  rapidly with only 32% solubilized in 8 hr. On the other hand, each of the other two reagents dissolved the  $\text{NpO}_2$  rapidly with greater than 90% dissolved after 1/2 hr of boiling (see Figure 3 for dissolution percentage as a function of time). Therefore, both the 4N  $\text{HNO}_3$ -0.1M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  and the 12N  $\text{HNO}_3$ -0.05N  $\text{HF}$ -0.01N  $\text{H}_2\text{SO}_4$  acid mixtures would be recommended for dissolution of  $\text{NpO}_2$  from filter media.

## V. Curium-244 oxide\* dissolution studies

Contaminated filter media were prepared by mixing 13 mg of  $\text{CmO}_2$  with 3 g of filter media. The experimental procedures were identical to those used for the uranium-233 dissolution studies (Section II). Two leaching reagents were tested, 12N  $\text{HNO}_3$ -0.05N  $\text{HF}$  and 4N  $\text{HNO}_3$ -0.1M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$ . The results from these tests are given in Table 5. As can be seen, both reagents dissolved the  $\text{CmO}_2$  very rapidly (essentially 100% dissolution in 10 min). Figure 4 shows the dissolution rate for 4N  $\text{HNO}_3$ -0.1M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$ . Both reagents are recommended for dissolution of  $\text{CmO}_2$  from contaminated filter media.

## VI. Thorium-232 oxide dissolution tests

Contaminated filter media were prepared by mixing 0.24 g of  $\text{ThO}_2$  with 3 g of filter media. The experimental procedure was identical to that used in the uranium-233 oxide studies (Section II). The  $\text{ThO}_2$

\*This was a powder which was approximately 94% Cm-244, 4% Cm-246, and the remainder Cm-245, Cm-247, and Cm-248.

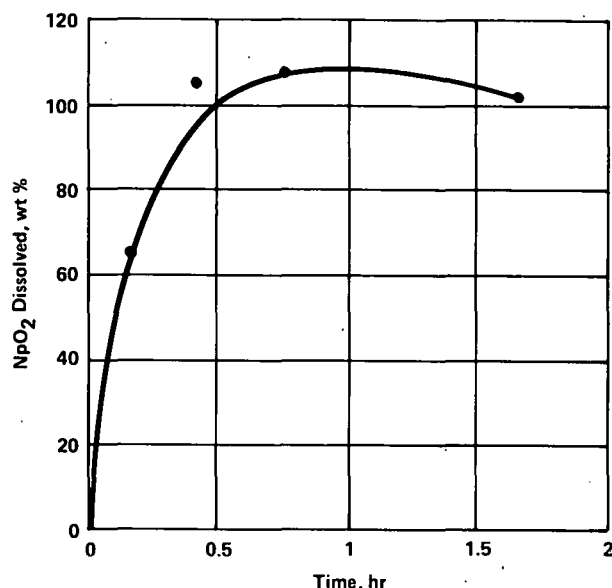


FIGURE 3 - Dissolution Rate of Np-237 Oxide in 12N  $\text{HNO}_3$ -0.05N  $\text{HF}$ -0.01N  $\text{H}_2\text{SO}_4$ .

was a fine powder which had been fired at 600°C. The purpose of the experiment was to determine if  $\text{ThO}_2$  would dissolve in 4N  $\text{HNO}_3$ -0.1M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  similarly to other actinide oxides previously tested ( $\text{PuO}_2$ ,  $\text{CmO}_2$ ,  $\text{AmO}_2$ ,  $\text{U}_3\text{O}_8$ , and  $\text{NpO}_2$ ). As can be seen from the data in Table 6 and Figure 5, the dissolution rate was slow at first (0-7 hr), rapid from 7 to 9 hr, and slow again from 9 to 20 hr. The reason for this is not known at this time, but it is thought that it took 7 hr to attack the surface of the  $\text{ThO}_2$  and open the pores so that rapid dissolution could begin. It therefore appears that 4N  $\text{HNO}_3$ -0.1M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  would be an acceptable leaching agent for  $\text{ThO}_2$  although dissolution time is longer than for other actinides. A point of interest is that in a similar experiment using just  $\text{ThO}_2$  (no filter media), the dissolution percentage obtained was only 25% after 20 hr of boiling. The reason for the difference in dissolution percentages (25% versus 100%) is not known at this time.

Table 5 - Cm-244 OXIDE DISSOLUTION TESTS

Acid Test Number	Reagent Composition	Time Heated (min)	CmO <sub>2</sub> Dissolved (wt%)
Cm-2A	12N HNO <sub>3</sub> -0.05N HF	10	103.3
Cm-2B	"	25	100.8
Cm-2C	"	60	105.0
Cm-2D	"	120	99.1
Cm-2E	"	180	104.5
Cm-2F	"	540	93.3
Cm-2G	"	540	99.0
Cm-3A	4N HNO <sub>3</sub> -0.1M CAN <sup>a</sup>	10	112.1
Cm-3B	"	25	110.0
Cm-3C	"	60	109.0
Cm-3D	"	120	107.6
Cm-3E	"	330	110.5

<sup>a</sup>CAN is an acronym for ceric ammonium nitrate.

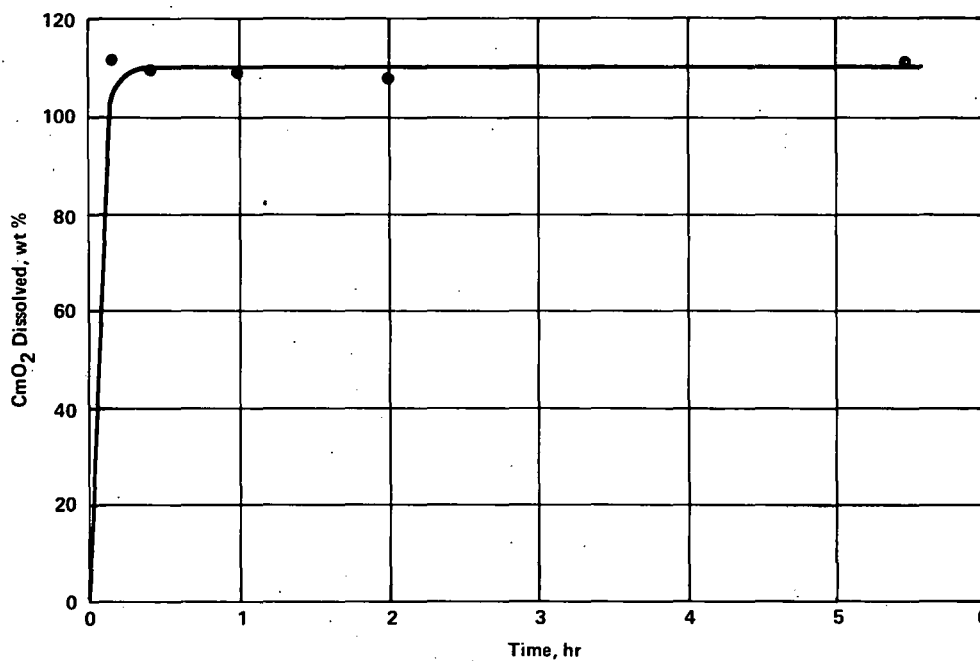
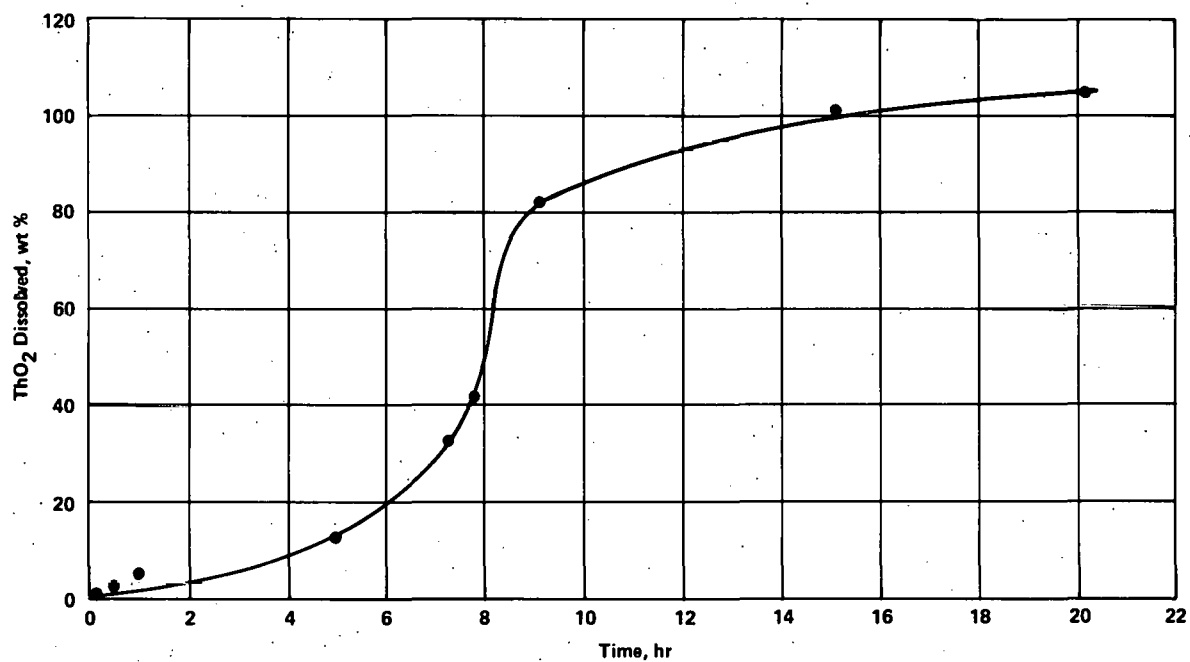
FIGURE 4 - Dissolution rate of CmO<sub>2</sub> in 4N HNO<sub>3</sub>-0.1M (NH<sub>4</sub>)<sub>2</sub>Ce(NO<sub>3</sub>)<sub>6</sub>.

Table 6 - ThO<sub>2</sub> DISSOLUTION DATA

Acid Test Number	Reagent Composition	Time Heated (min)	ThO <sub>2</sub> Dissolved (wt%)
Th-1	4N HNO <sub>3</sub> -0.1M CAN <sup>a</sup>	5	0.6
Th-2	"	10	0.4
Th-3	"	20	1.4
Th-4	"	35	2.4
Th-5	"	60	5.1
Th-6	"	180	4.7
Th-7	"	300	12.5
Th-8	"	440	31.7
Th-9	"	440	29.3
Th-10	"	470	42.1
Th-11	"	545	82.3
Th-12	"	905	100.5
Th-13	"	1205	104.3

<sup>a</sup>CAN is an acronym for ceric ammonium nitrate.

FIGURE 5 - Wt % ThO<sub>2</sub> dissolved in 4N HNO<sub>3</sub>-0.1M (NH<sub>4</sub>)<sub>2</sub>Ce(NO<sub>3</sub>)<sub>6</sub>.

## VII. Dissolution of $\text{PuO}_2$ from actual glovebox filters

These tests were conducted using samples of filter media removed from actual spent glovebox filters. The  $\text{PuO}_2$  on the filters was the same type as that used in the fusion experiments (Section III).

### EXPERIMENT A

A 9.5 g sample of filter media containing 0.18 g of  $\text{PuO}_2$  was dissolved in a six-stage dissolution procedure. The leaching solution volume was based on the flow-sheet submitted last quarter [1], namely, 18 ml of solution per gram of filter media. The ceric ammonium nitrate required was based on the initial moles of  $\text{PuO}_2$  present in the filter media (60 moles  $\text{Ce}^{+4}$  per mole  $\text{PuO}_2$ ). Using these criteria, the leaching reagent in the first five stages was 4N  $\text{HNO}_3$ -0.23M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$ , and the sixth stage contained 12N  $\text{HNO}_3$ -0.1N HF. The 12N  $\text{HNO}_3$ -0.1N HF was used because the 4N  $\text{HNO}_3$ -0.23M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  had proved ineffective in the fifth stage of dissolution. The volume of 12N  $\text{HNO}_3$ -0.1N HF used was 250 ml, and all studies were done at boiling temperature. The solution was sampled periodically for plutonium concentration, and all samples were filtered through 4-5  $\mu\text{m}$  glass filters. When the dissolution rate for a particular stage had become very slow (as indicated by  $\alpha$  analysis of solution), the filter media were separated from the leaching solution by filtration and placed in the next stage of dissolution. Results for this experiment are given in Table 7. Figures 6 through 8 show the dissolution rate for each stage, and Figure 9 shows the cumulative  $\text{PuO}_2$  dissolution percentage for the six-stage process. It should be noted that final dissolution percentage

was 99.88% (a contamination factor of 833). This was determined by gamma ray spectrometry of the residual filter media and agrees well with the final value as determined by alpha counting of the solution samples (99.80%). It should also be noted that the 12N  $\text{HNO}_3$ -0.1N HF solution used in the sixth stage was more than twice as effective as the 4N  $\text{HNO}_3$ -0.1M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  used in the fifth stage (0.079% dissolved versus 0.031%). The reason for this is not understood at this time. More studies will be done to investigate this phenomenon.

### EXPERIMENT B

A 25.6 g sample of filter media containing 0.31 g of  $\text{PuO}_2$  was dissolved in a four-stage dissolution process. The leaching reagent used in the test was 4N  $\text{HNO}_3$ -0.1M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$ . The solution volume was again based on the flowsheet (18 ml of solution per gram of filter media), and the amount of ceric ammonium nitrate was increased to 120 times the initial moles  $\text{PuO}_2$  present in the filter media. The experimental procedure was identical to that used in Experiment A of this section. The results for this experiment are given in Table 8; Figure 10 shows the cumulative  $\text{PuO}_2$  dissolution percentage for the four stage process. The dissolution percentages in Table 8 were determined by alpha counting, but the cumulative percentages used in Figure 10 were based on a final gamma spectrometry value of 98.24%. As one can see, this value compares well with the value of 97.62% obtained by alpha counting. The gamma value is, however, considered more accurate than the total obtained by simple addition of the individual alpha count values. The decontamination factor obtained using this gamma value is 55.8. The decontamination factor obtained after four stages in Experiment A of this section



Table 7 - PuO<sub>2</sub> DISSOLUTION DATA (9.5-g sample)

Acid Test Number	Reagent Composition	Time Heated (min)	PuO <sub>2</sub> Dissolved (wt%)	Process Stage
256-1 A	4N HNO <sub>3</sub> -0.23M CAN <sup>a</sup>	10	82.0	1
256-2 A	"	20	79.5	1
256-3 A	"	40	87.3	1
256-4 A	"	60	80.9	1
256-5 A	"	120	85.8	1
256-6 A	"	210	90.6	1
256-1 B	"	10	0.41	2
256-2 B	"	20	1.17	2
256-3 B	"	45	2.50	2
256-4 B	"	65	2.97	2
256-5 B	"	120	4.75	2
256-6 B	"	360	7.30	2
256-7 B	"	510	7.47	2
256-1 C	"	30	0.29	3
256-2 C	"	75	0.55	3
256-3 C	"	135	0.87	3
256-4 C	"	315	1.06	3
256-5 C	"	435	1.09	3
256-6 C	"	675	1.31	3
256-7 C	"	675	1.39	3
256-1 D	"	20	0.029	4
256-2 D	"	45	0.063	4
256-3 D	"	60	0.079	4
256-4 D	"	135	0.139	4
256-5 D	"	330	0.213	4
256-6 D	"	645	0.222	4
256-7 D	"	645	0.228	4
256-1 E	"	15	0.011	5
256-2 E	"	30	0.012	5
256-3 E	"	60	0.015	5
256-4 E	"	135	0.023	5
256-5 E	"	195	0.025	5
256-6 E	"	465	0.029	5
256-7 E	"	600	0.031	5
256-8 E	"	780	0.032	5
256-9 E	"	780	0.025	5
256-1 F	12N HNO <sub>3</sub> -0.1N HF	30	0.016	6
256-2 F	"	60	0.030	6
256-3 F	"	240	0.047	6
256-4 F	"	585	0.079	6
256-5 F	"	945	0.075	6

<sup>a</sup>CAN is an acronym for (NH<sub>4</sub>)<sub>2</sub>Ce(NO<sub>3</sub>)<sub>6</sub>.

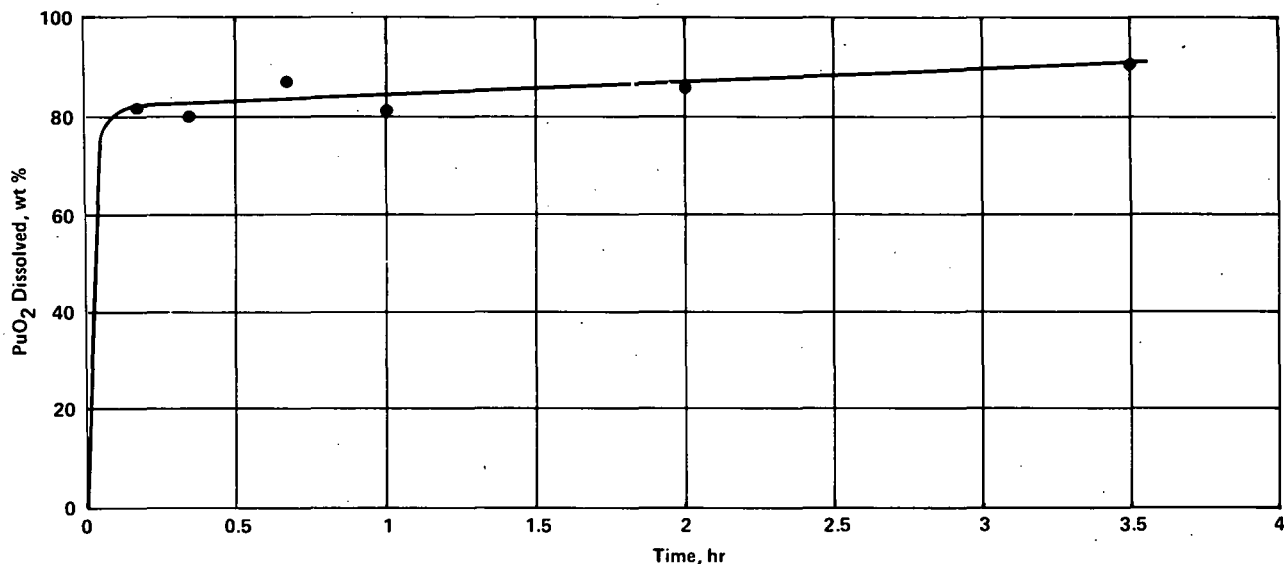


FIGURE 6 - Dissolution rate of  $\text{PuO}_2$  from filter media (1st stage).

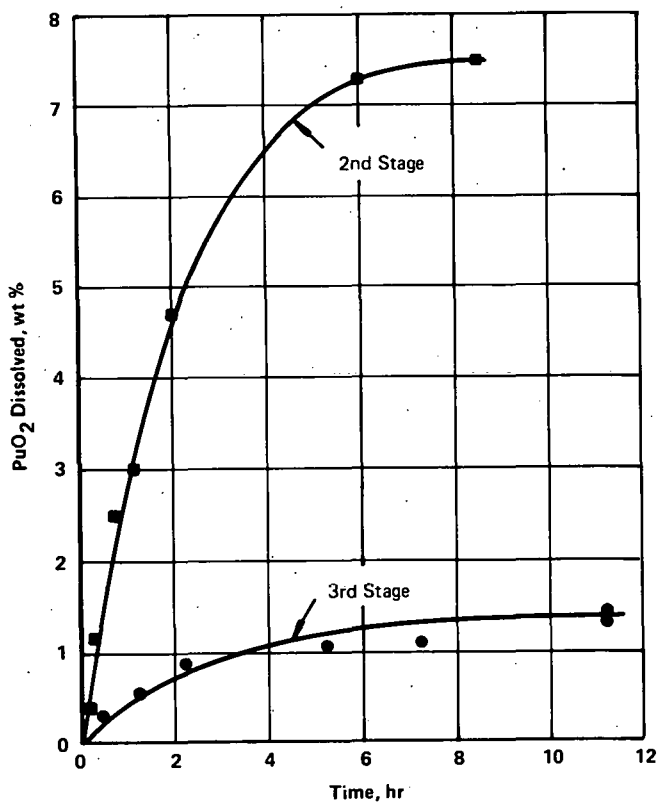


FIGURE 7 - Dissolution rate of  $\text{PuO}_2$  from filter media (2nd and 3rd stage).

was 417. A logical question would be: Why was the decontamination factor less when more ceric ammonium nitrate per mole of  $\text{PuO}_2$  was used? It is known that the 25.6-g sample of filter media contained a greater percentage of pack-to-frame sealant\* than the 9.5-g sample. It is also known (from the color change of the leaching solution) that the  $\text{Ce}^{+4}$  ions were reduced much more quickly during the dissolution of  $\text{PuO}_2$  from the 25.6-g filter media sample. Therefore, it is surmised that the reason for the smaller decontamination factor (smaller percent dissolved) is that the  $\text{Ce}^{+4}$  ions were mainly used to oxidize the sealant thus leaving

\*This is the material that is used to seal the asbestos-glass filter pack to the wooden frame. It is a fire retardant polyurethane foam and a rubber base adhesive.

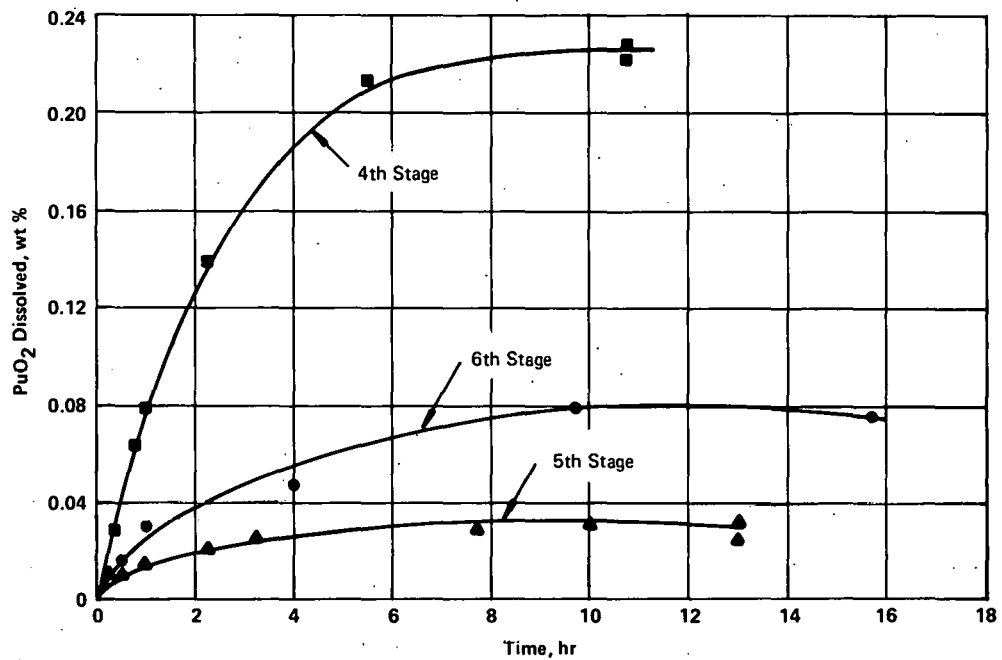


FIGURE 8 - Dissolution of  $\text{PuO}_2$  from filter media (4th, 5th, and 6th stages).

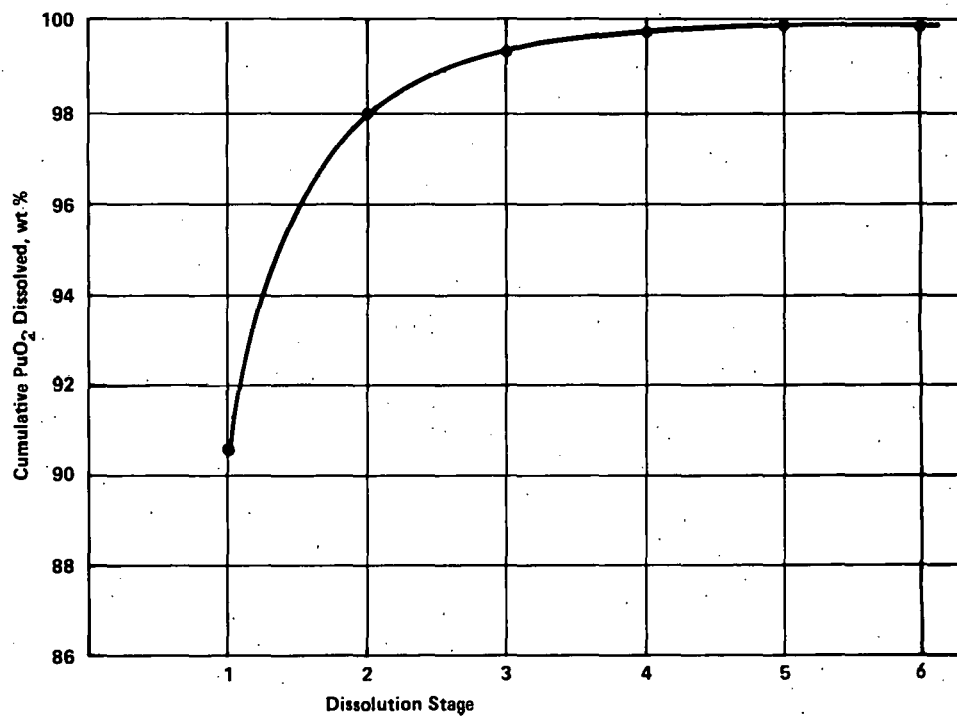


FIGURE 9 - Cumulative weight %  $\text{PuO}_2$  dissolved during six stage process.

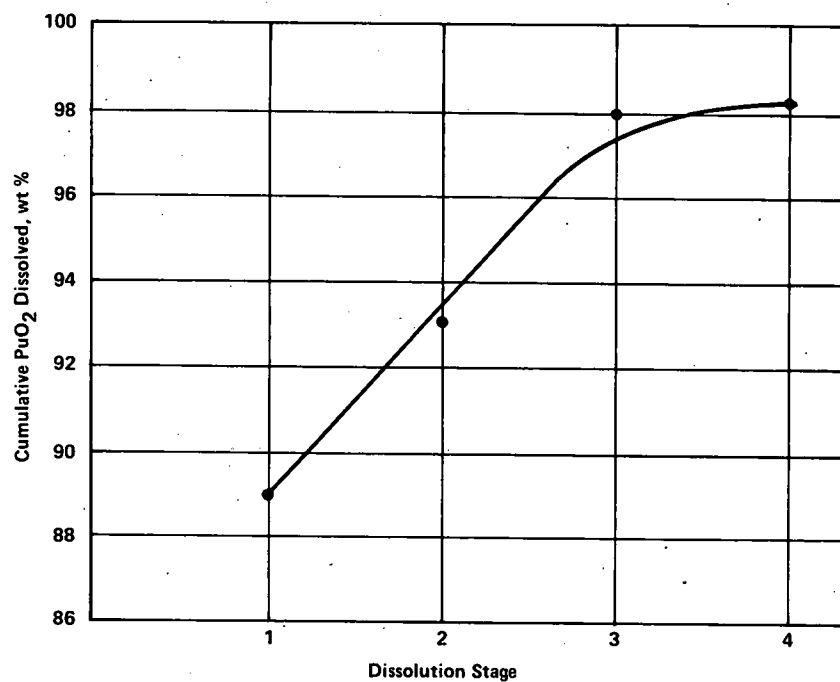


FIGURE 10 - Cumulative weight % PuO<sub>2</sub> dissolved during four stage process.

Table 8 - PuO<sub>2</sub> DISSOLUTION DATA (25.6-g sample)

Acid Test Number	Reagent Composition	Stage	Time Heated (hr)	PuO <sub>2</sub> Dissolved (wt%)	Cumulative PuO <sub>2</sub> Dissolved (wt%)
12-1	4N HNO <sub>3</sub> -0.3M CAN <sup>a</sup>	1	6 1/2	88.40	88.40
12-2	"	2	15	4.04	92.44
12-3	"	3	7 1/4	4.91	97.35
12-4	"	4	7	0.27	97.62 <sup>b</sup>

<sup>a</sup>CAN is an acronym for ceric ammonium nitrate.

<sup>b</sup>Gamma ray analysis of remaining filter media indicated that 98.24% of the PuO<sub>2</sub> had been dissolved.

fewer  $\text{Ce}^{+4}$  ions to oxidize to  $\text{Pu}^{+4}$ . A simple solution to this problem might be to increase the amount of ceric ammonium nitrate when larger amounts of sealant are present with the filter media.

#### EXPERIMENT C

A 30-g sample of filter media containing 0.33 g  $\text{PuO}_2$  was dissolved in a four-stage dissolution process as shown on the flow-sheet submitted last quarter [1]. The amount of ceric ammonium nitrate required for each step was based on the assumption that 0.01 mole  $\text{Ce}^{+4}$  would oxidize 1 Ci (0.058 g) of plutonium-238 oxide. Verification of this assumption was the purpose of this experiment. A 90% dissolution of the  $\text{PuO}_2$  present was to be achieved in each of the four steps. The results of this experiment can be seen in Figure 11. Only 93.96% of the  $\text{PuO}_2$  was dissolved in the four-stage process (a decontamination

factor of only 16.55). This experiment certainly showed that the desired decontamination factor of 10,000 cannot be achieved using the above mentioned basis for determining the amount of  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  required. As found in Experiment B, the amount of ceric ammonium nitrate must be based on the total amount of material ( $\text{PuO}_2$ , sealant, and other organics) present. This must be determined for each dissolution since the amount to be reacted varies with each batch.

## Summary

Filter media which were contaminated with americium-241 and plutonium-239 oxides were treated with three different leaching reagents to determine the effectiveness of each reagent in dissolving the actinide oxides. The three reagents used were 8N  $\text{HNO}_3$ , 4N  $\text{HNO}_3$ -0.1M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$ , and 12N  $\text{HNO}_3$ -0.05N  $\text{HF}$ -0.01N  $\text{H}_2\text{SO}_4$ . It was

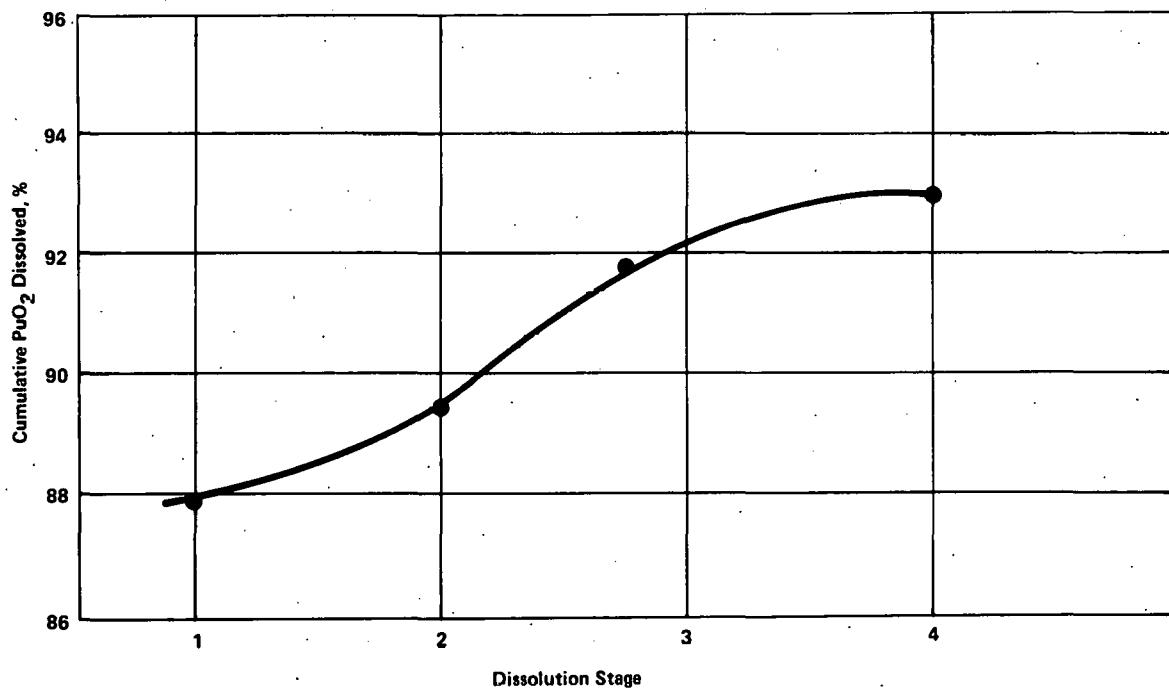


FIGURE 11 - Cumulative weight %  $\text{PuO}_2$  dissolved during four stage process.

found that the best leaching agent was the 4N  $\text{HNO}_3$ -0.1M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  which dissolved greater than 90% of the  $\text{AmO}_2$  and the  $\text{PuO}_2$ .

Another series of tests was conducted to determine the best leaching agent for uranium-233 oxide ( $\text{U}_3\text{O}_8$ ). Simulated samples of contaminated filter media were prepared by mixing 0.2 g  $\text{U}_3\text{O}_8$  powder with 3 g of shredded filter media. The reagents used were the same three as used for the  $\text{AmO}_2$ - $\text{PuO}_2$  studies. It was found that all three reagents were successful in dissolving greater than 90% of the  $\text{U}_3\text{O}_8$  in a very short time (15 min).

Fusion studies were conducted using  $\text{Na}_2\text{CO}_3$  and  $\text{PuO}_2$ . The purpose of these experiments was to determine the effect of temperature on fusion efficiency. It was found that there was not a significant difference in percent  $\text{PuO}_2$  solubilized (4.6%) within the temperature range under investigation (1200-1300°C).

Experiments were conducted using filter media contaminated with neptunium-237 oxide. The purpose of the tests was to determine which of the three reagents would be successful in dissolving the  $\text{NpO}_2$ . Results indicate that both the 4N  $\text{HNO}_3$ -0.1M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  and the 12N  $\text{HNO}_3$ -0.05N HF-0.01N  $\text{H}_2\text{SO}_4$  were successful in dissolving greater than 90% of the  $\text{NpO}_2$  in 0.5 hr. The 8N  $\text{HNO}_3$  on the other hand solubilized only 32% of the  $\text{NpO}_2$  in 8 hr and therefore would not be recommended as a leaching agent for  $\text{NpO}_2$ .

In other series of experiments, filter media contaminated with curium-244 oxide were dissolved in leaching tests using both 4N  $\text{HNO}_3$ -0.1M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  and

12N  $\text{HNO}_3$ -0.05N HF. Both reagents dissolved essentially 100% of the  $\text{CmO}_2$  in approximately 10 min and therefore would be recommended for dissolving  $\text{CmO}_2$  from filter media.

Filter media contaminated with thorium-232 oxide were leached with a 4N  $\text{HNO}_3$ -0.1M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  solution in an attempt to dissolve the  $\text{ThO}_2$ . The rate of dissolution was slow at the beginning (0-7 hr), rapid from 7 to 9 hr, and slow again at the end (9-20 hr). Complete dissolution was achieved in 20 hr, considerably longer than for the other actinide oxides previously tested. A point of interest is that in a similar experiment using only 4N  $\text{HNO}_3$ -0.1M ceric ammonium nitrate and  $\text{ThO}_2$  (no filter media), the amount dissolved was only 25% after 20 hr. The reason for this difference is not understood at this time. Therefore, it is concluded that although the dissolution time was much longer for the  $\text{ThO}_2$ , it is still an acceptable leaching reagent.

Several experiments were completed using  $\text{PuO}_2$  contaminated filter media from spent glovebox filters. These tests were made in order to verify the flowsheet given last quarter [1] and to determine how much  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  is required for dissolution of  $\text{PuO}_2$  from such filters. In one experiment, a 9.5-g sample of filter media was processed through a six-stage dissolution process. In the initial five stages, 4N  $\text{HNO}_3$ -0.23M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  was used and in the sixth stage 12N  $\text{HNO}_3$ -0.1N HF was used. The amount of  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  used in each stage was 60 times the initial moles of  $\text{PuO}_2$  present in the filter media. The overall decontamination factor obtained for the six-stage process was 833. In a second experiment, a 25.6-g sample of filter media was processed through a four-stage dissolution procedure. The leaching



solution was 4N  $\text{HNO}_3$ -0.30M  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$ . The amount of ceric ammonium nitrate used in each stage was 120 times the initial moles of  $\text{PuO}_2$  present in the filter media. The overall decontamination factor achieved in the four-stage process was 55.8. In a third experiment, a 30-g sample of filter media was dissolved in a four-stage dissolution process. The amount of  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  used in each stage of this process was based upon the assumption that 0.01 mole of  $\text{Ce}^{+4}$  would oxidize 1 Ci (0.058 g) of plutonium-238. The overall decontamination factor achieved for the four-stage procedure was 16.55. The flow-sheet given last quarter [1] was based on the above assumption (0.01 mole  $\text{Ce}^{+4}$ /curie Pu-238) and an overall decontamination factor of 10,000. This degree of decontamination was certainly not achieved and hence the flowsheet is not valid.

The amount of ceric ammonium nitrate required cannot be based exclusively upon the amount of actinide present in the filter media. The reason for this is

that other reactants such as pack-to-frame sealant\* and other organic materials are present in the spent filter media. A high percentage of sealant was present in the 25.6-g sample and this is perhaps why less plutonium was dissolved. In other words, even though more ceric ammonium nitrate was used per mole of plutonium present, a lower percentage of the plutonium dissolved because a large percentage of the  $\text{Ce}^{+4}$  ions were depleted by reaction with organics present. A formula for calculating the  $(\text{NH}_4)_2\text{Ce}(\text{NO}_3)_6$  required per gram of filter media has not been developed. More studies must be done before the relationship can be determined.

\*This is a fire retardant polyurethane and a rubber base adhesive.

## Reference

1. D. F. Luthy and E. L. Lewis, Decontamination of HEPA Filters: April-June 1977, MLM-2445 (September 22, 1977), Figure 6, p. 12.

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