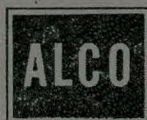


RADIOCHEMICAL ANALYSIS OF CRUD FROM THE ARMY PACKAGE POWER REACTOR



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**RADIOCHEMICAL ANALYSIS OF CRUD
FROM THE ARMY PACKAGE POWER REACTOR**

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By

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February 15, 1958

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ABSTRACT

A study has been made of the radiochemical composition and the specific activity of insoluble corrosion products (crud) removed from the primary system of the APPR-1. This report presents the results of analysis of twelve crud samples collected during the interval from September 3, 1957 to December 1, 1957.

The samples were radiochemically analyzed for long-lived gamma emitting nuclides only. Data are presented on the measured values of the specific activity of crud, the ratios of the nuclide specific activities, and the concentration of crud (crud level) in the circulating primary water. Also included is data, based on the analysis of a single sample, comparing the specific activity of the deposited and circulating corrosion products.

I. INTRODUCTION

An experimental program is in progress to obtain information on the buildup of long-lived radioactivity on the primary surfaces of the APPR-1. Quantitative measurements of the long-lived nuclides circulating in the coolant were first made during the 700 hour continuous performance test of the reactor (1). The experimental methods employed for that study measured the total activity of the coolant sample, with the results reported on a volume basis (dpm/ml of coolant). Consequently, the results of these initial measurements of the coolant activity were subject to the variation of the coolant solids level (crud level). Crud levels were not determined during the 700 hour test since crud sampling equipment was not installed in the system.

Investigations at WAPD have indicated that the major portion of the long-lived activity is associated with insoluble material (crud) suspended in the coolant (2). In order to obtain direct measurements of the activity of this insoluble material in the coolant of APPR-1, a suitable crud sampling device was installed in the primary purification system blowdown line. This report presents the data obtained from the radiochemical analysis of twelve crud samples withdrawn over the period from September 3, 1957 to December 1, 1957.

II. CONCLUSIONS

1. The insoluble corrosion products (crud) obtained from the primary system of the APPR-1 contain the following nuclides: Co^{60} , Co^{58} , Fe^{59} , Mn^{54} , and Cr^{51} . These nuclides account for all the long-lived gamma activity associated with the crud.
2. The specific activity (dpm/mg) of these nuclides appears to be increasing with reactor operating time.
3. The specific activity of Co^{60} is accumulating faster than that of Co^{58} in the circulating crud. Mn^{54} specific activity is accumulating at about the same rate as that of Co^{60} .
4. Uncertainties in the experimental data preclude defining a relationship between crud specific activity and crud level.

III. RECOMMENDATIONS

1. Further investigation is necessary to establish the mechanism by which activity accumulates on primary system components, and to define the relationship between circulating activity and deposited activity.
2. It should be experimentally determined if a relationship exists between crud specific activity and crud level.
3. It is necessary to establish improved methods to accurately measure the concentration (crud level) of primary system insoluble material.
4. Measurements of the $\text{Co}^{58}/\text{Co}^{60}$ ratio should continue in order to establish the relative importance of cobalt impurities with respect to the nickel content of stainless steel. The possibility of Co^{60} arising from an n-p reaction on Ni^{60} deserves consideration.

IV. EXPERIMENTAL METHODS

A. SAMPLING PROCEDURE

Figure 1 is a simplified flow diagram of the APPR-1 primary system showing the location of the crud sampling device. The sampling unit is a 4-1/2 inch long by 1-1/2 inch diameter micrometallic "PSS" type 316 stainless steel, hollow cylindrical filter element, mounted within a 2 inch diameter stainless steel housing. The element is attached to the filter housing by means of a screw type insertion. Flow passes from the outside surface through the porous stainless steel into the hollow center, and out through the threaded connection. The element is rated to retain any particle larger than 2 microns. In collecting the first five crud samples, the filter was precoated with a fine spectrographic grade of graphite to increase filtering efficiency. However, difficulty was encountered in removing the graphite during subsequent treatment of the filters, and the pre-coating was eliminated when collecting the last seven samples. Flow through the filter was regulated by means of three needle valves. A positive displacement flowmeter with an accuracy of 2% was used to measure flow.

Sampling times were chosen to give a minimum of activity yet provide sufficient sample to permit accurate analysis. Approximately 65-70 gallons of water were passed through each filter over a period of several days. Upon removal of the filter element from the housing, the filter plus the water and crud contained in the housing were placed in a polyethylene container and shipped to Schenectady for radiochemical analysis.

B. CRUD REMOVAL AND DISSOLUTION TECHNIQUES

In preparation for radiochemical analysis of the crud, an effort was

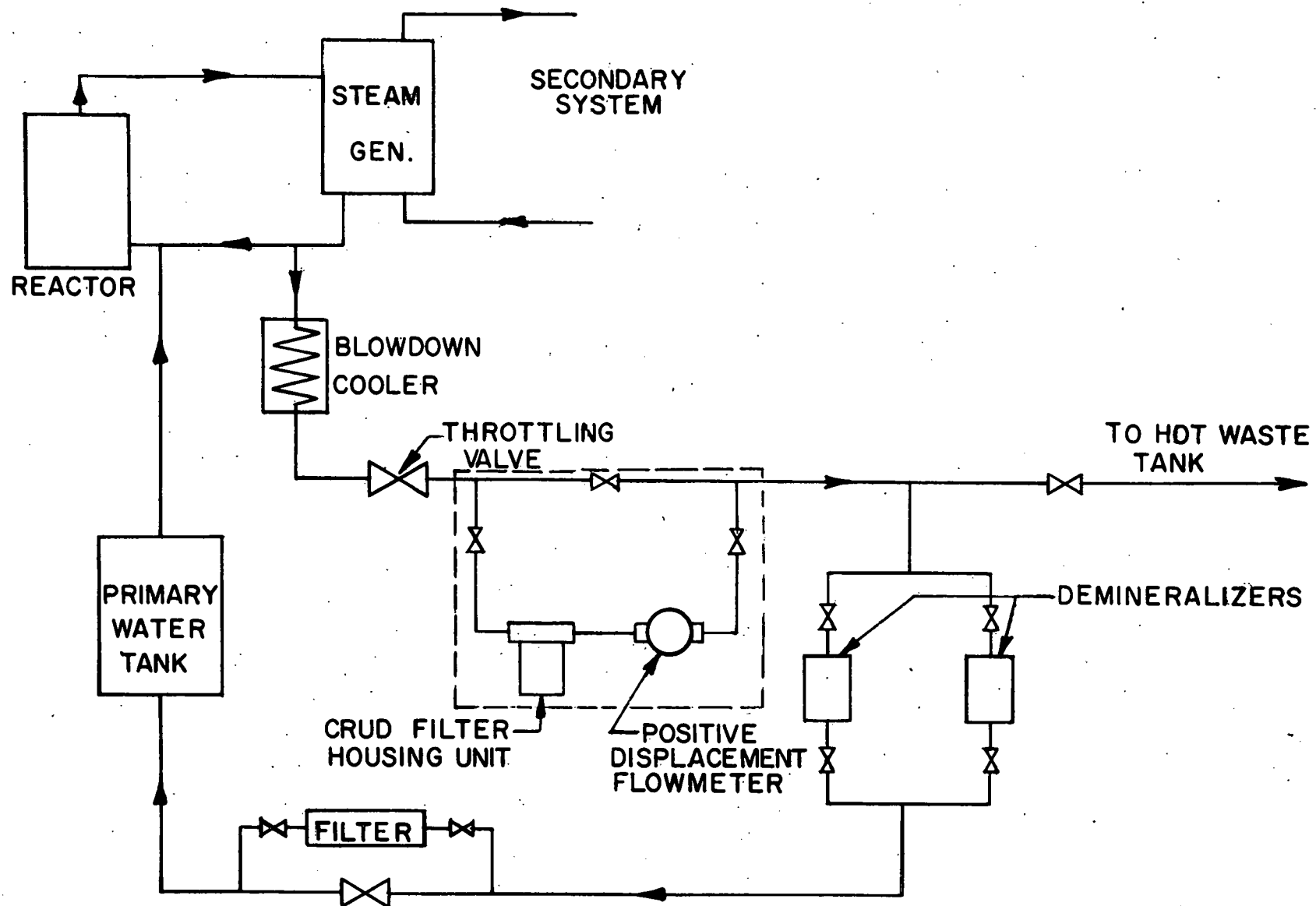


FIG. 1 SAMPLING DIAGRAM

made to remove all crud and activity from the filter elements by means of a chemical decontaminating solution. The decontaminating solutions used actually removed about 80 percent of the activity from the filter. The solution was then evaporated to dryness, ignited to remove volatile and organic matter, and the oxide residue dissolved and radiochemically analyzed. It was expected that the weight of the oxide residue would be an accurate measure of the total weight of crud collected. This general procedure was used on the first nine filters. However, since it was found that this technique did not permit a direct determination of the weight of crud collected, an alternate procedure was used for samples 10-12. For these samples, the crud was removed directly from the filter by backwashing, weighed, and brought into solution for analysis. A limitation of this procedure was that only about 25 percent of the activity was removed by backwashing.

The detailed procedures followed for removal and dissolution of the crud are as follows:

1. EDTA-Sulfamic Acid Method (Samples 1 - 5).

The filter element was placed in a solution containing 13 grams EDTA (ethylenediaminetetraacetic acid disodium salt) and 12 grams of sulfamic acid in 600 ml. of distilled water at 150-160°F for one hour. The filter was removed from the decontaminating solution, and the graphite and insoluble material filtered off and ignited over a Meeker burner. The filtrate was treated with 25 ml. each of concentrated nitric acid and concentrated sulfuric acid and evaporated to dryness. The residue was combined with the ignited oxides, and mixed with 20 parts by weight of a fusion mixture (50-50 Na_2CO_3 - K_2CO_3). The

mixture was fused at 1300-1400°F for one hour, cooled, and dissolved slowly in 3 N hydrochloric acid. An acid insoluble salt residue was removed by centrifugation and discarded, since it contained no measurable radioactivity. The supernate was then made up to a definite volume for analysis.

2. EDTA-Oxalic Acid Mixture (Samples 6, 7)

The decontaminating solution was essentially the same as the above, except oxalic acid was substituted for sulfamic acid. This solution and the insoluble material removed from the filter was treated successively with sulfuric, nitric, and perchloric acids. After evaporating to fumes of perchloric acid to destroy organic matter, the solution was centrifuged to remove insoluble material. The insoluble material was fused as above, and dissolved in 3 N HCl. As with samples 1-5, there was an acid insoluble, non-radioactive residue which was discarded. The two solutions were combined and made up to volume for analysis.

3. EDTA-Citric Acid Method (Samples 8, 9)

The unit was decontaminated in a solution consisting of equal weights of EDTA and citric acid (1 gram) in 600 ml. of distilled water at 160°-180°F for 1 - 2 hours. The solution was then treated in the same manner as the EDTA-oxalic acid method.

4. Filter Backflushing Method

The filter unit was backflushed with air and water to remove as much activity as possible. A rubber policeman was used to remove any loosely adherent material. The water and crud solution was then evaporated to dryness in a tared platinum dish, ignited, and weighed to determine the amount of insoluble material recovered. This residue was then successively treated with a mixture of nitric, perchloric, and hydrofluoric acids until dissolution was attained. The

solution was then made up to a known volume for analysis.

C. DETERMINATION OF CRUD WEIGHTS

The crud removal and dissolution procedures followed for samples 1 - 9 did not permit a direct determination of the weight of insoluble material (crud) removed from the filter. This became evident when analyzing aliquots of the final solutions. Metal hydroxides which were precipitated from an aliquot by sodium hydroxide, washed, and ignited, yielded oxide weights considerably higher than would be normally expected. Crud weights for samples 1 - 9 were therefore calculated indirectly from chemical analysis of the dissolved crud solutions. These calculated weights of crud are given in Table I. Crud weights for samples 10, 11, 12 were measured directly by weighing the insoluble material recovered by backflushing the filter element. These weights are also included in Table I.

Chemical analyses were made by standard colorimetric methods using a Beckman DU spectrophotometer. Aliquots of the solutions were analyzed for iron, nickel, chromium, manganese, and cobalt. It was assumed that the oxides of these elements accounted for 100 percent of the weight removed from the filter. Results of the chemical analyses are reported in Table II. The variation in results indicates that the decontaminating solutions had a definite effect on the calculated weight of crud. Also, each of the three chemical solutions apparently attacked the stainless steel of the filter element in a different manner.

TABLE I
SAMPLE CRUD WEIGHTS

Sample No.	Crud Filter Decontaminating Solution	Crud Weight * mg.
1	EDTA + Sulfamic Acid	174
2	EDTA + Sulfamic Acid	89
3	EDTA + Sulfamic Acid	79
4	EDTA + Sulfamic Acid	398
5	EDTA + Sulfamic Acid	123
6	EDTA + Oxalic Acid	170
7	EDTA + Oxalic Acid	298
8	EDTA + Citric Acid	52
9	EDTA + Citric Acid	40
10	Filter Backwashed	50**
11	Filter Backwashed	24
12	Filter Backwashed	64

* Calculated as the total metal oxides: Fe_3O_4 , NiO , MnO_2 , Cr_2O_3 .

**Direct weight of recovered material.

TABLE II

CHEMICAL COMPOSITION OF CRUD
EXPRESSED AS PERCENT OF TOTAL METAL

Sample No.	Chemical Decontamination Solution	% Fe	% Ni	% Cr	% Mn	% Co
1	EDTA + Sulfamic Acid	64.0	18.0	3.3	14.7	0
2	EDTA + Sulfamic Acid	79.0	4.9	4.35	13.0	0
3	EDTA + Sulfamic Acid	71.5	13.6	3.6	11.2	0
4	EDTA + Sulfamic Acid	69.1	15.7	0.6	14.7	0
5	EDTA + Sulfamic Acid	86.0	8.7	5.8	0	0
6	EDTA + Oxalic Acid	48.5	53.0	0	0	0
7	EDTA + Oxalic Acid	68.5	31.0	0	0.6	0
8	EDTA + Citric Acid	88.0	12.0	0	0	0
9	EDTA + Citric Acid	96.0	4.0	0	0	0

D. RADIOCHEMICAL PROCEDURES

The activity of the dissolved crud solutions was sufficiently high so that it was not necessary to perform a general chemical separation of the elements. Instead small aliquots of the solution were analyzed directly for cobalt, iron, manganese, and chromium according to the procedures outlined in the Appendix. Analyses were corrected for the amount of iron present in the crud. Corrections for the other elements were not required, since the aliquots contained only negligible amounts of those metals.

Samples were counted for Co^{60} , Co^{58} , Fe^{59} , Mn^{54} , and Cr^{51} by a modified method of Connalley and Leboeuf (1,3) using the photopeak height r_{max} and the width of the photopeak at $1/2 r_{\text{max}}$. The counts were corrected for decay back to the midpoint of the sample collection period. Fe^{55} could not be counted by this technique. In our previous measurements (1), Fe^{55} had been reported to have a total disintegration energy of 0.21 Mev. This energy is actually emitted in a continuous spectrum. A check of this energy peak indicated a decay with a half life closely approximating that of Fe^{59} (45 days). It therefore could only be concluded that the peak in question was in reality due to Compton scattering from Fe^{59} . Work is still continuing to find a suitable quantitative method for detecting Fe^{55} . The manganese K-Xray emitted as a result of the electron capture of Fe^{55} cannot be observed with the present equipment.

It is possible that the presence of Co^{58} can lead to an error when measuring Co^{60} at the 1.33 Mev peak with the gamma spectrometer. On a theoretical basis, the 0.51 Mev positron annihilation radiation and the 0.81 Mev gamma ray from Co^{58} may combine to give 1.32 Mev gamma energy. Therefore,

the possibility exists that the counts measured at the 1.33 Mev Co^{60} peak may partially be due to radiation "scattering up" from Co^{58} . Since the present method of analyses for Co^{60} consists of measuring the photopeak heights at 1.17 and 1.33 Mev, the results reported from the 1.33 Mev peak could be in error. The effect of this phenomena on our results is discussed in Section V A.

III. DISCUSSION OF RESULTS

During the period covered by this report, the APPR-1 was operating on a schedule of twenty days at continuous power level followed by approximately nine days at intermittent low power and shutdowns for training purposes. The effects of this erratic type of operation must be considered when analyzing the data. Figure 2 is a plot of the APPR-1 $^{\circ}\text{F days}^*$ as a function of time. The slope of the line is a direct indication of the power level. The reduced slopes (dotted lines) represent periods of shutdown and operation at low power levels. Since steady state conditions are not reached under these conditions, nuclide activation becomes complex. The nuclides Fe^{59} and Cr^{51} will decay more rapidly, leaving the bulk of the activity associated with the long lived nuclides Mn^{54} and Co^{60} .

A. SPECIFIC ACTIVITY OF CRUD AS A FUNCTION OF EQUIVALENT FULL POWER HOURS

The variation in the nuclide specific activities with increasing reactor operating time is summarized in Table III. For the first nine samples, the actual specific activities of the nuclides may be different than the results reported in this table. The possibility of error arises from two sources:

$$\begin{aligned} *1^{\circ}\text{F day} &= 1.2 \text{ equivalent full power hours} \\ &= 0.00137 \text{ MWYR} \end{aligned}$$

1. The acid-complexing agent used to decontaminate the crud filter may actually have removed activity associated with colloidal and soluble material, in addition to the activity associated with the insoluble crud. It has been reported ⁽⁴⁾ that porous stainless steel filters will retain activity associated with colloidal and soluble material.

2. The calculated weights of crud, from which the specific activity is found, may be high due to attack on the steel filter by the chemical decontaminating solutions. It is suspected from the chemical analysis results, that some attack did in fact occur, particularly with Samples 1-7.

The specific activity results of Samples 10, 11, and 12 can be considered the most reliable, since the activity was removed by backflushing with water. These values can be assigned with confidence to insoluble material which is large enough to be retained by the stainless steel filter (2 microns or greater). However, it should be pointed out that all of the insoluble material may not have been removed by the backwashing procedure. If the activity associated with insoluble corrosion products is independent of particle size, the specific activity values are not affected. The specific activity of crud probably is independent of particle size ⁽⁵⁾. Caution must be exercised in comparing the results of Samples 10-12 with Samples 1-9 because of the difference in treatment of the collected crud.

Examination of the data of Table III reveals the specific activities of cobalt on a volume basis do not change appreciably throughout the interval covered in this study except in the case of Sample 12. However, the cobalt results expressed on a weight basis (dpm/mg) show larger variations. The overall trend appears to be an increase in the weight specific activity with reactor operating time, with the volume specific activity remaining relatively constant.

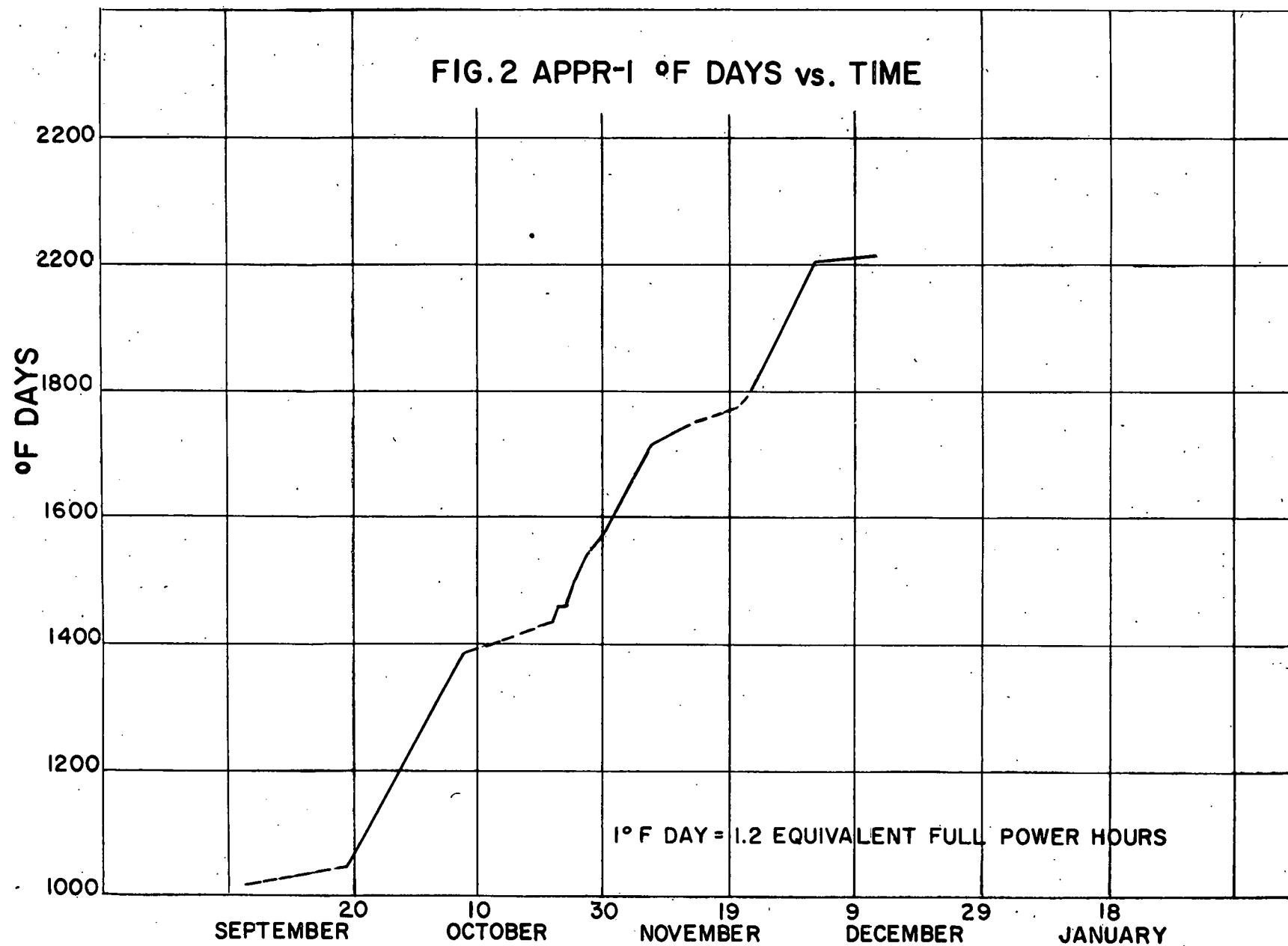


TABLE III
SPECIFIC ACTIVITY OF CRUD VS. EQUIVALENT FULL POWER HOURS

Sample No.	efph	⁶⁰ Co				⁵⁸ Co		⁵⁹ Fe		⁵⁴ Mn		⁵¹ Cr	
		1.17 Mev. Peak		1.33 Mev. Peak									
		dpm/mg	dpm/ml	dpm/mg	dpm/ml	dpm/mg	dpm/ml	dpm/mg	dpm/ml	dpm/mg	dpm/ml	dpm/mg	dpm/ml
1	1210	0.73x10 ⁵	11.8	0.81x10 ⁵	13.1	4.4x10 ⁵	73.9	0.67x10 ⁵	11.3	0.77x10 ⁵	13.0	0.86x10 ⁵	14.4
2*	1265	2.9 x10 ⁵	---*	2.8 x10 ⁵	---*	14 x10 ⁵	---*	1.6 x10 ⁵	---*	.90x10 ⁵	---*	3.7 x10 ⁵	---*
3	1300	4.4 x10 ⁵	340	4.1 x10 ⁵	322	26 x10 ⁵	2020	4.7 x10 ⁵	376	1.6 x10 ⁵	129	5.0 x10 ⁵	395
4	1445	2.1 x10 ⁵	404	2.3 x10 ⁵	447	10 x10 ⁵	1940	1.4 x10 ⁵	269	0.58x10 ⁵	110	1.3 x10 ⁵	250
5	1470	4.9 x10 ⁵	242	5.2 x10 ⁵	265	27 x10 ⁵	1295	8.8 x10 ⁵	430	2.0 x10 ⁵	96.8	8.4 x10 ⁵	412
6	1744	5.8 x10 ⁵	395	5.8 x10 ⁵	395	28 x10 ⁵	1949	5.8 x10 ⁵	394	2.0 x10 ⁵	135	15 x10 ⁵	1038
7	1802	3.7 x10 ⁵	447	3.6 x10 ⁵	434	16 x10 ⁵	1948	3.5 x10 ⁵	419	1.0 x10 ⁵	121	4.0 x10 ⁵	483
8	1869	15 x10 ⁵	320	15 x10 ⁵	309	66 x10 ⁵	1369	64 x10 ⁵	1334	7.1 x10 ⁵	147	22 x10 ⁵	463
9	1934	14 x10 ⁵	222	13 x10 ⁵	206	65 x10 ⁵	934	60 x10 ⁵	961	7.8 x10 ⁵	126	43 x10 ⁵	680
10	2185	15 x10 ⁵	302	14 x10 ⁵	281	60 x10 ⁵	2019	11 x10 ⁵	352	5.0 x10 ⁵	99.3	29 x10 ⁵	571
11	2265	11 x10 ⁵	361	10 x10 ⁵	332	43 x10 ⁵	1380	7.9 x10 ⁵	254	3.0 x10 ⁵	58	19 x10 ⁵	360
12	2385	30 x10 ⁵	770	33 x10 ⁵	835	131 x10 ⁵	3340	24 x10 ⁵	625	12 x10 ⁵	296	58 x10 ⁵	1500

* No Flow Data

Since the volume specific activity (dpm/ml) is dependent on crud level, a relationship may possibly exist between the crud level and weight specific activity. More data is required to establish if a definite relationship does exist.

The large increase in the specific activity on both the volume and weight basis observed in Sample 12 is difficult to explain. The difference is real since the duplicate samples agreed well. The time at power between Sample 11 and 12 is not long enough to account for this large variation. Since both volume and weight values increased the deviation from expected behavior is perhaps attributable to a non-representative sample.

Samples 8 and 9, in which citric acid - EDTA was the complexing agent, show a considerably higher weight specific activity than the previous samples. This could possibly be attributed to the greater complexing ability or the less corrosive attack of this solution on the structural material of the filter element. Probably both mechanisms are simultaneously involved resulting in the large increase. The volume specific activities of these two samples are comparable to those for other samples, except in the case of Fe^{59} . This might indicate that the citric acid-EDTA-iron complex is stronger than the other metal ion complexes, so that the iron activity is preferentially removed.

From this data in Table III, the effect of Co^{58} scattering on the Co^{60} results can be obtained. If this additive scattering were a significant factor, it would be expected that the results from the 1.33 Mev peak of Co^{60} would be consistently higher than the results from the 1.17 Mev peak of Co^{60} . An examination of the data from the two Co^{60} peaks shows a maximum difference between two values of about 10 percent, with the average variation about five percent. In approximately half of the cases where there is a difference, the 1.17 peak is

in fact higher than the 1.33 peak. On the basis of these observations the differences observed are attributed to experimental error. Further observations on the possibility of this additive scattering will be made by measuring the decay rate of the samples at the 1.33 Mev peak.

The cobalt, iron, manganese, and chromium nuclides accounted for all of the long-lived gamma activity in the samples. In several cases the material balance on the separated fractions apparently accounted for more than 100%. The paradox is explained by the very high counting rates of the gross samples using our method of assay ⁽¹⁾. The counting rates are sufficiently high to have appreciable coincidence loss, even when employing scintillation counting techniques.

B. NUCLIDE RATIO DATA

Table IV presents the ratios of the specific activity of each nuclide to the specific activity of Co^{60} . The ratios were calculated to Co^{60} since this nuclide is of principal interest in activity build-up problems because of its nuclear properties -- long half-life and two cascading hard gammas per disintegration. The use of ratios eliminates some of the errors previously mentioned that could be present in the absolute values of the specific activity. For example, if the specific activity is in error primarily due to material being dissolved from the filter, this error will be canceled in the ratio. In addition, in the case of $\text{Co}^{58}/\text{Co}^{60}$ ratios, chemical errors will also be eliminated.

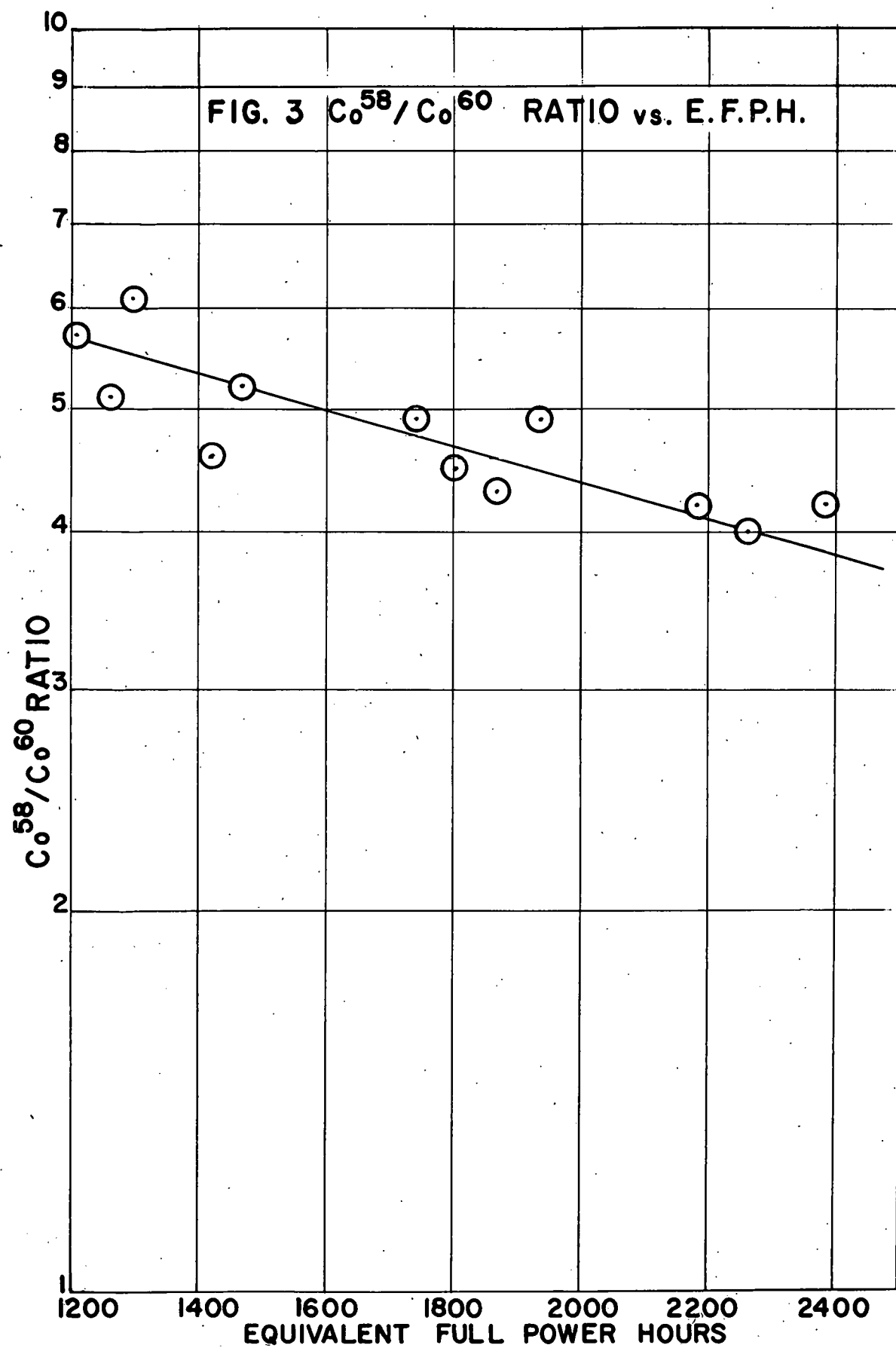
The ratios of $\text{Co}^{58}/\text{Co}^{60}$ and $\text{Mn}^{54}/\text{Co}^{60}$ are plotted as a function of equivalent full power hours in Figure 3 and Figure 4 respectively. The curves were determined by the method of least squares. The decreasing $\text{Co}^{58}/\text{Co}^{60}$ ratio indicates that the specific activity of Co^{60} is accumulating at a faster rate

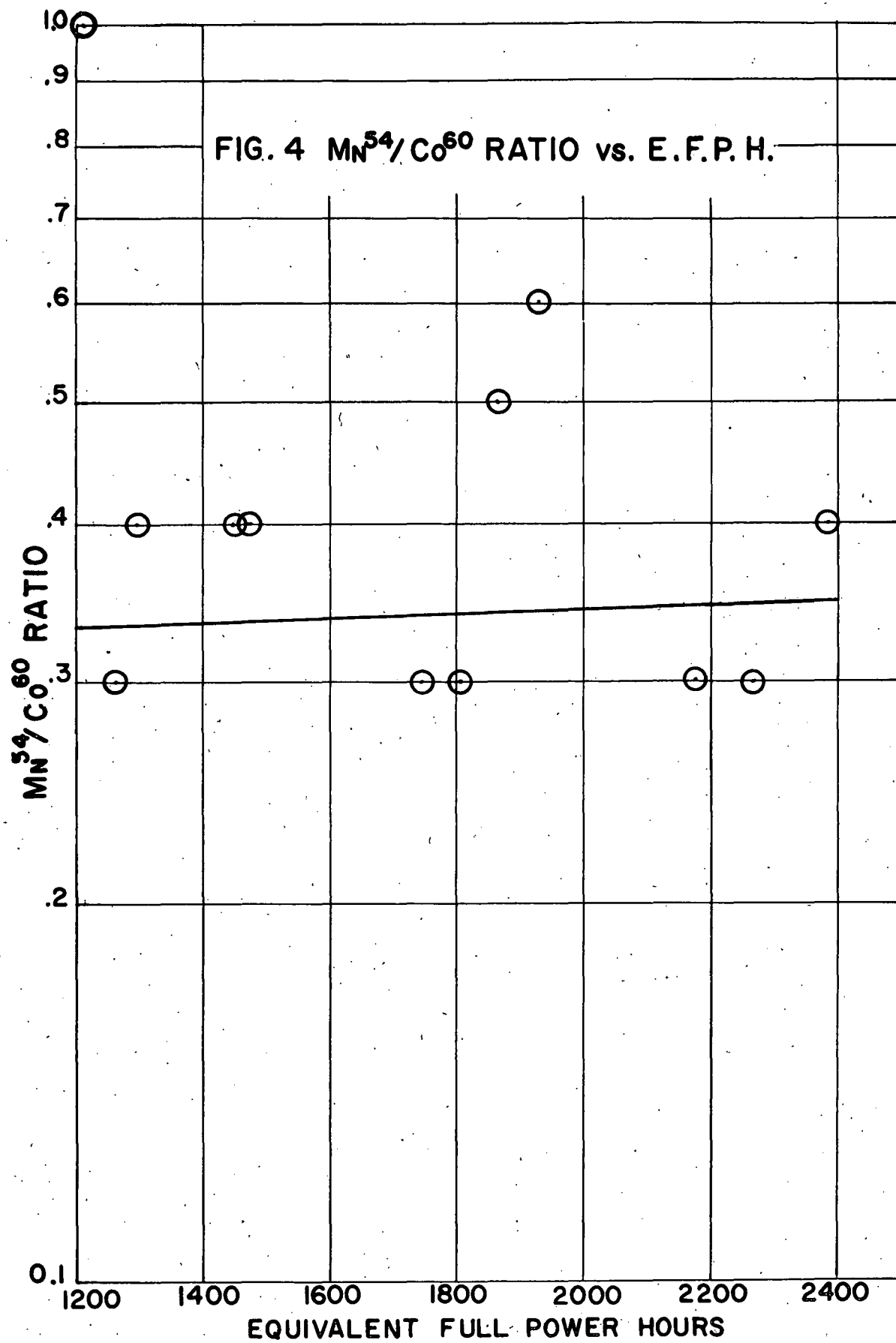
than that of Co^{58} . It is still too early to predict what the equilibrium value for this ratio will be, or whether the Co^{60} specific activity will eventually overtake the Co^{58} specific activity. Figure 4 indicates that Mn^{54} specific activity is accumulating at about the same rate as that of Co^{60} .

It is interesting to note that the value of the $\text{Co}^{58}/\text{Co}^{60}$ ratios for these crud samples is nearly the same as the $\text{Co}^{58}/\text{Co}^{60}$ ratio for the total activity measured in the water samples at the end of the 700-hour test. On sample 711, the last water sample of the 700-hour test, the $\text{Co}^{58}/\text{Co}^{60}$ was 4.5. The fact that the values in Table III were calculated from the weight specific activities, while the 700-hour water sample ratios were from volume specific activities is of no consequence. The ratios are the same whether volume or weight specific activities are used.

TABLE IV. NUCLIDE RATIO DATA

Sample	$\text{Co}^{58}/\text{Co}^{60}$	$\text{Fe}^{59}/\text{Co}^{60}$	$\text{Mn}^{54}/\text{Co}^{60}$	$\text{Cr}^{51}/\text{Co}^{60}$
1	5.7	0.9	1.0	1.1
2	5.1	0.6	0.3	1.3
3	6.1	1.1	0.4	1.2
4	4.6	0.6	0.4	0.6
5	5.2	1.7	0.4	1.7
6	4.9	1.0	0.3	1.1
7	4.5	1.0	0.3	1.1
8	4.3	4.2	0.5	1.5
9	4.9	4.5	0.6	3.2
10	4.2	0.7	0.3	2.0
11	4.0	0.8	0.3	1.7
12	4.2	0.8	0.4	1.9





C. CRUD LEVEL AND SPECIFIC ACTIVITY

The term "crud level" has been used in the literature in several different ways. The definition depends upon such factors as the method of sampling or collecting the crud, and the methods followed to recover and calculate the amount of crud collected. Consequently, it is necessary to define the term as it appears in this report. Crud levels for Samples 1-9 were determined on the calculated weight of material, as determined by chemical analysis, removed from the porous steel filter by the decontaminating solutions. The results are expressed on the basis of the total flow through the filter unit, i.e., milligrams of crud per liter of flow (ppm). Crud levels for samples 10-12 were determined from the amount of material removed by back-flushing the filter with air and distilled water until no further activity was removed. Levels were expressed in the same manner as Samples 1-9. Crud levels for these samples would be expected to be lower than for the first nine samples, but the values of specific activity for these samples are more reliable.

The specific activity and the crud level are two of the factors that can affect the activity buildup. The values for these parameters measured during this study are given in Table V. The results for the first nine samples may be in error for the reasons previously mentioned.

WAPD has investigated ⁽⁶⁾ whether there is a relationship between crud specific activity and crud level. In view of the uncertainties in our results, we do not feel justified in drawing a conclusion from this data as to whether a relationship exists between the specific activity and crud level. However, because of the importance of these parameters, further data on their values should be obtained.

TABLE V
CRUD LEVEL VS CRUD SPECIFIC ACTIVITY

Sample	Total Specific Activity (dpm/mg) $\times 10^{+5}$	Crud Level (ppm)
1	7.49	0.17
2	----	----
3	41.4	1.9
4	15.6	0.78
5	50.6	0.49
6	56.2	0.69
7	28.5	1.2
8	174.	0.21
9	189.	0.16
10	120.	0.20
11	83.5	0.10
12	256.	0.29

D. DEPOSITED VS CIRCULATING ACTIVITY

Investigations at WAPD (7) have indicated that the specific activity of loosely bound deposited material is probably of the same order of magnitude as the specific activity of circulating crud. However, little data is available on whether adherent deposits follow this relationship. During this study, one piece of datum was collected on the relationship between deposited activity and circulating activity. The results reported below are based upon one measurement only, and this should be borne in mind when interpreting the data.

At the end of the 700-hour test, a section of the baffle plate was removed from the steam generator and sent to Schenectady for radiochemical studies. In the process of obtaining and shipping the sample, any loosely bound material would have been lost. Therefore the relationships discussed below are only applicable to tightly bound material. The baffle plate was decontaminated with the EDTA-sulfamic acid solution previously described. Table VI presents the nuclide ratio data from this sample and compares it to the nuclide ratios from the last water sample (No. 711) of the 700-hour test. Both samples had a similar reactor history.

The difference between the $\text{Co}^{58}/\text{Co}^{60}$ ratios in the two samples is larger than can be explained by experimental error. If the deposited activity arises from a strictly chemical phenomena, the $\text{Co}^{58}/\text{Co}^{60}$ ratio would be expected to be the same, since the two cobalt nuclides would behave chemically identical. If a physical phenomena, such as diffusion, plays a part in the accumulation of deposited activity, the difference in the mass of the nuclides would become a factor.

TABLE VI

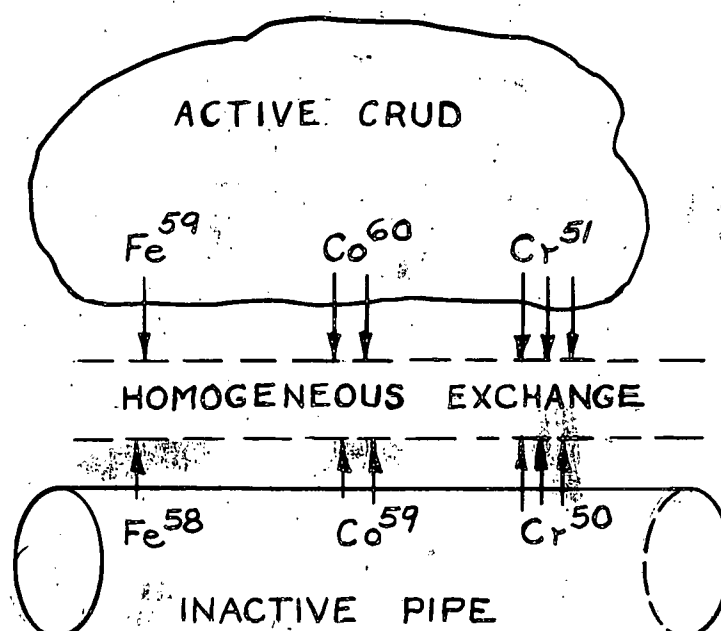
NUCLIDE RATIOS OF DEPOSITED AND CIRCULATING ACTIVITY

Sample	$\text{Co}^{58}/\text{Co}^{60}$	$\text{Fe}^{59}/\text{Co}^{60}$	$\text{Mn}^{54}/\text{Co}^{60}$	$\text{Cr}^{51}/\text{Co}^{60}$
Baffle Plate	6.0	0.08	0.20	1.3
711	4.5	0.26	0.28	0.39

One mechanism which could explain these observations could be an initial rate-determining E.M.F. reaction followed by a rapid homogeneous exchange reaction between the active and inactive nuclides. Figure 5 is a diagrammatic outline of this mechanism.

FIG. 5

POSSIBLE MECHANISM FOR ACCUMULATION OF ACTIVITY IN DEPOSITS



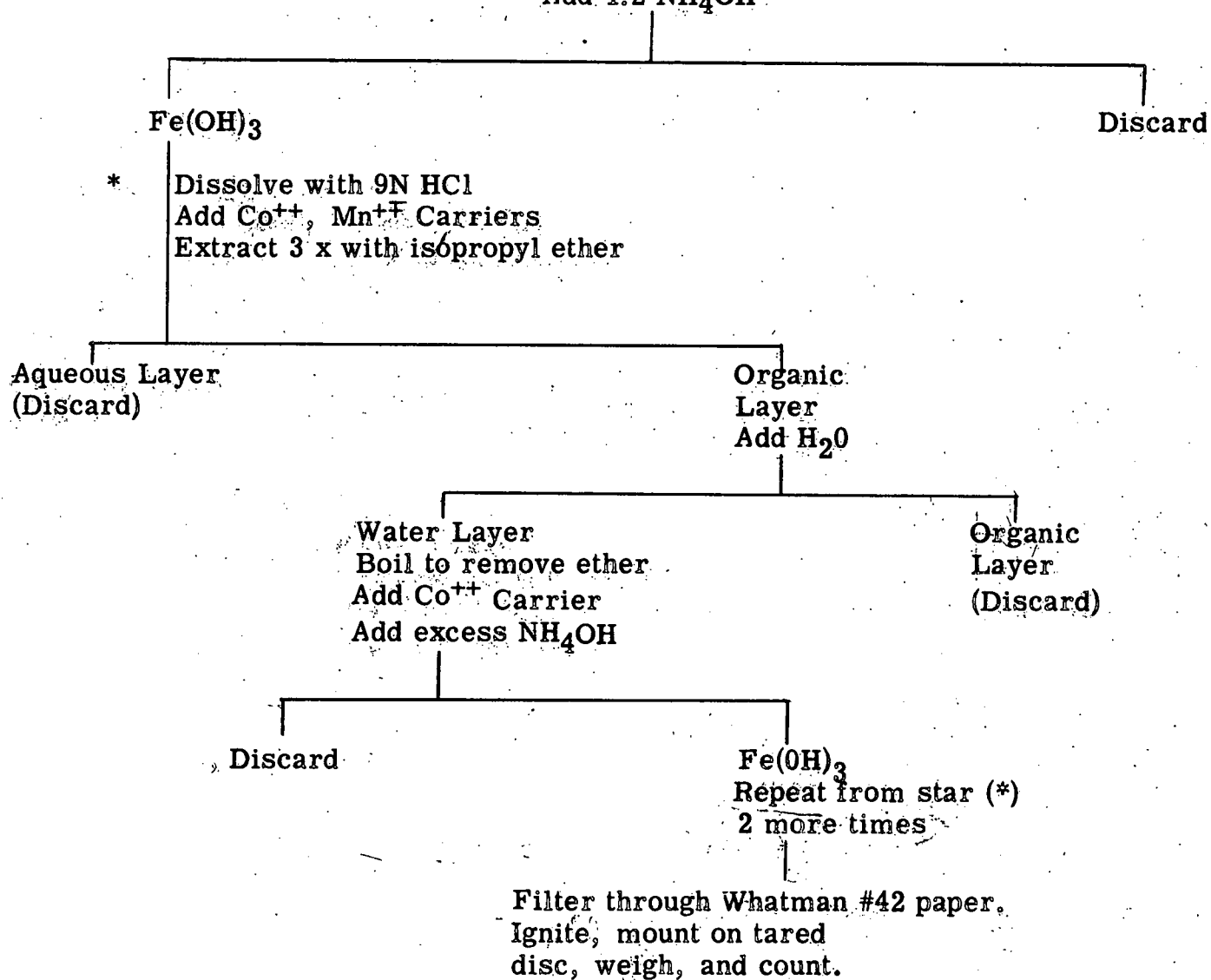
It will be assumed the majority of the radioactive iron, cobalt, and chromium is associated with insoluble material. These elements (both active and inactive species) would enter into solution relative to their E.M.F. potentials ($M^0 \rightarrow M^{n+} + ne$). Since iron is more noble than cobalt and cobalt more noble than chromium, the rates with which these elements would enter solution would be in the order chromium > cobalt > iron. Once in solution there would be rapid exchange between the active and inactive nuclides. A "plating out" could then occur with the exchanged nuclides. The rates of this reaction would be in reverse order to the dissolution step, i.e. iron > cobalt > chromium. With this mechanism, an analysis shows the rate of accumulation of chromium in the deposit is faster than cobalt, which in turn would be faster than iron.

Additional work is required to accurately establish the relationship between deposited and circulating activity, and to define the mechanism by which activity builds up on the metal surfaces of the primary system components.

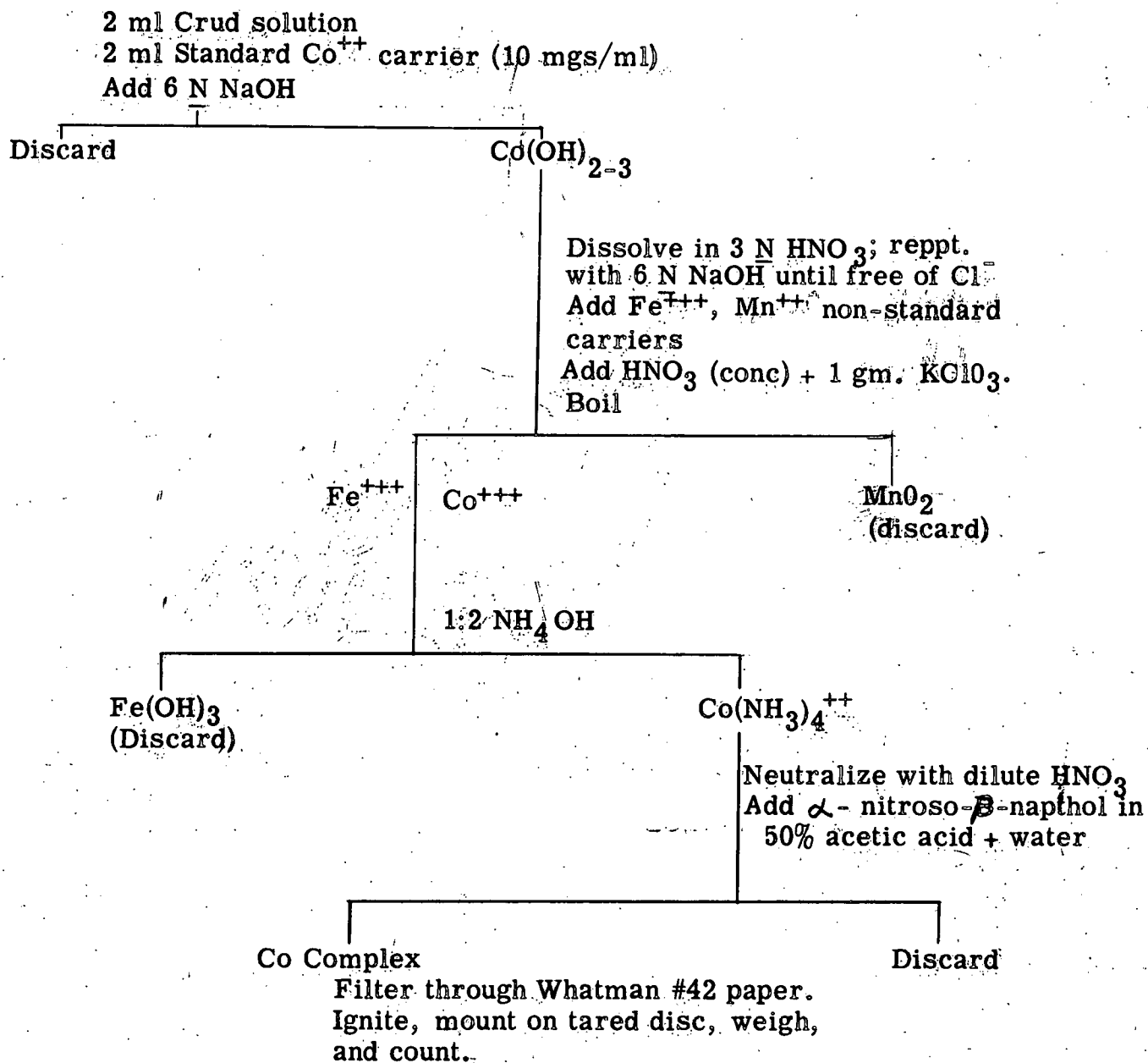
IV. APPENDIX
DECONTAMINATION SCHEME
FOR CRUD SAMPLES

IRON

2 ml Crud solution
2 ml Standard Fe^{+++} carrier (10 mgs/cc)
2 ml Non-Standard Co^{++} carrier
Add 1:2 NH_4OH



COBALT



MANGANESE

2 or 5 ml Crud Solution

5 ml 3N HCl

Boil

2 ml Standard Mn^{++} carrier (10 mgs/ml)

Add 6N NaOH

$Mn(OH)_2$

Discard

Dissolve in 3N HNO_3 ;
reppt. with 6N NaOH
until free of Cl^-

*Add Co^{++} , Fe^{+++} , Cr^{+++} carriers
Add HNO_3 (conc) 1 gm $KClO_3$

Discard

MnO_2

Wash with hot, dilute HNO_3
solution 3X

Reduce with HNO_2 ($NaNO_2$ + dil. HNO_3)

Repeat from STAR (*)

Add 6N NaOH

$Mn(OH)_2$

Discard

Dissolve in dil. HCl
Neutralize with 1:2 NH_4OH (Care)
Heat to $60^\circ C$
Add $(NH_4)_2 HPO_4$
Cool slowly

Discard

$Mn(NH_4)PO_4$

Mount on tared disc,
weigh, and count.

CHROMIUM

2 or 5 ml Crud solution

2 ml Standard Cr^{+++} carrier (10 mgs/ml)

(1) Make basic and add H_2O_2

(2) Make acid and add H_2O_2

Heat to 60°C

Add sat'd solution $\text{Ba}(\text{NO}_3)_2$

Cr^{+++}

Make basic with NaOH
Add 20 cc 30% H_2O_2
Boil gently to remove
excess H_2O_2

BaSO_4
(discard)

Discard

Ba CrO_4
or
 $\text{CrO}_4^{=}$

Neutralize with dil. HNO_3
Pass 2 x slowly over cation
resin, H^+ cycle.

$\text{H}_2\text{Cr}_2\text{O}_7$

Add Sat'd solution $\text{Ba}(\text{NO}_3)_2$
Make slightly basic

BaCrO_4

Discard

Mount on tared disc, weigh, and
count.

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