

1 **Satellite solar-induced chlorophyll fluorescence and near-infrared reflectance**
2 **capture complementary aspects of dryland vegetation productivity dynamics**

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26
27 **Abstract**

28 Mounting evidence indicates dryland ecosystems play an important role in driving the interannual
29 variability and trend of the terrestrial carbon sink. Nevertheless, our understanding of the
30 seasonal dynamics of dryland ecosystem carbon uptake through photosynthesis [gross primary
31 productivity (GPP)] remains relatively limited due in part to the limited availability of long-term

32 data and unique challenges associated with satellite remote sensing across dryland ecosystems.
33 Here, we comprehensively evaluated longstanding and emerging satellite vegetation proxies in
34 their ability to capture seasonal dryland GPP dynamics. Specifically, we evaluated: 1)
35 reflectance-based proxies normalized difference vegetation index (NDVI), soil adjusted
36 vegetation index (SAVI), near infrared reflectance index (NIR_v), and kernel NDVI (kNDVI) from
37 the MODerate resolution Imaging Spectroradiometer (MODIS); and 2) newly available
38 physiologically-based proxy solar-induced chlorophyll fluorescence (SIF) from the
39 TROPOspheric Monitoring Instrument (TROPOMI). As a performance benchmark, we used GPP
40 estimates from a robust network of 21 western United States eddy covariance tower sites that
41 span representative gradients in dryland ecosystem climate and functional composition. We
42 found that NIR_v and SIF were the best performing GPP proxies and captured complementary
43 aspects of seasonal GPP dynamics across dryland ecosystem types. NIR_v offered better
44 performance than the other proxies across relatively low-productivity, sparsely non-evergreen
45 vegetated sites ($R^2 = 0.59 \pm 0.13$); whereas SIF best captured seasonal dynamics across relatively
46 high-productivity sites, including evergreen-dominated sites ($R^2 = 0.74 \pm 0.07$). Notably, across
47 grass-dominated sites, all reflectance-based proxies (NDVI, SAVI, NIR_v and kNDVI) showed
48 significant seasonal bias (hysteresis) that strengthened with the total fraction of woody vegetation
49 cover, likely due to seasonal patterns in woody vegetation reflectance that are unrelated to or
50 decoupled from GPP. Future efforts to fully integrate the complementary strengths of NIR_v and
51 SIF could significantly improve our understanding and representation of dryland GPP dynamics
52 in satellite-based models.

53 **Keywords:** gross primary productivity, dryland heterogeneity, remote sensing, near-infrared
54 reflectance, solar-induced fluorescence

55 **1. Introduction**

56 Dryland ecosystems cover more than 40% of Earth's land surface and are a major driver of the
57 interannual variability and subsequent trend in the terrestrial carbon sink (Poulter et al., 2014;
58 Ahlstrom et al, 2015; Smith et al., 2019). A better understanding of these critical ecosystems is
59 urgently needed, since aridity is projected to increase with climate change, potentially resulting in
60 the expansion of drylands to more than half of the planet's terrestrial surface by 2100 (Reynolds
61 et al., 2007; Huang et al., 2016, 2017; Yao et al., 2020). Yet, despite their emerging importance,
62 there remains limited availability of continuous, long-term measurements of carbon cycling in
63 dryland regions, resulting in data products that are poorly constrained (Biederman et al., 2017).

64 Satellite observations have provided key insights into dryland GPP dynamics over the last four
65 decades (Smith et al. 2019), and have been used as observational input in numerous satellite-
66 based GPP products (Smith et al., 2016; Bodesheim et al., 2018; Robinson et al., 2018). Yet, GPP
67 cannot be directly observed and satellite-based GPP products have been found to perform
68 relatively poorly across heterogeneous dryland regions (Verma et al., 2014; Biederman et al.,
69 2017). For example, the widely-used MODerate resolution Imaging Spectroradiometer (MODIS)
70 GPP product captured only about 30% of the interannual variation in GPP observations across a
71 network of dryland eddy covariance tower (EC) sites in United States (US) (Biederman et al.,
72 2017). This is due in large part to unique challenges associated with drylands, which represent
73 dynamic mixtures of herbaceous, woody, and bare soil components, that limit the ability of
74 common satellite vegetation proxies, such as the normalized difference vegetation index (NDVI),
75 in their ability to track vegetation function (Smith et al., 2018; Allred et al., 2020; Ma et al.,
76 2020).

77 Multiple surface reflectance-based proxies have been developed to better constrain our
78 understanding of dryland GPP including soil adjusted vegetation index (SAVI), near infrared
79 reflectance index (NIR_v), and kernel NDVI (kNDVI). SAVI and NIR_v are thought to reduce the
80 confounding effects of background soil brightness (Huete, 1988; Badgley et al., 2017), and NIR_v
81 is also thought to capture the depth distribution of canopy photosynthetic capacity, thus more
82 accurately capturing changes in structurally complex landscapes (Badgley et al., 2017, 2019;
83 Baldocchi et al., 2020). The recently developed kNDVI was designed to exploit all higher-order
84 relationships between the input surface reflectance observations, and thus better represent any
85 non-linearity in the NDVI:GPP relationship (Camps-Valls et al., 2021). Yet, to our knowledge,
86 these reflectance-based proxies, particularly NIR_v and kNDVI, have not been comprehensively
87 evaluated across dryland ecosystem types specifically.

88 Solar-induced chlorophyll fluorescence (SIF) is a promising advance in remote sensing and
89 fundamentally different from the above vegetation reflectance-based proxies. SIF is not based on
90 vegetation reflectance, but instead a measure of radiance emitted by the vegetation during the
91 light reactions of photosynthesis. SIF is unique in that it has both physical and physiological
92 controls, and thus may represent a critical advance in our ability to track GPP (Joiner et al., 2014;
93 Walther et al., 2016). Previous studies have indicated a near-linear relationship with GPP, but the
94 extent to which factors such as environmental condition and spatial heterogeneity mediate the
95 SIF:GPP relationship remains unresolved and an area of active research (Guanter et al., 2007,
96 2014; Sun et al., 2017; Smith et al., 2018). A particular advantage of SIF is its apparent ability to
97 track changes in GPP even in the absence of changes in spectral reflectance (Smith et al., 2018;
98 Zuromski et al., 2018; Magney et al., 2019). Surface reflectance-based proxies often overestimate
99 GPP when plants are green but photosynthetically inactive (e.g., evergreen species during periods

100 of severe drought), which can result in seasonal hysteretic relationships between these proxies
101 and GPP (Gitelson et al., 2014; Flanagan et al., 2015; Nestola et al., 2016; Wang et al., 2020a).
102 The recently launched TROPospheric Monitoring Instrument (TROPOMI) represents a
103 revolutionary advance in SIF observation by providing daily global coverage and relatively high
104 spatial resolution (Köhler et al., 2018). With TROPOMI SIF estimates, we have a first-time
105 opportunity to fully evaluate at relatively high spatiotemporal resolution the potential of SIF for
106 improving seasonal GPP estimates across dryland ecosystems.

107 Drylands have long been recognized as useful test-beds for evaluating satellite proxies in their
108 ability to track different aspects of vegetation dynamics (Tucker, 1979; Smith et al., 2019).
109 However, it remains unclear how factors like functional and structural diversity, such as bare
110 ground exposure, impact these evaluations (Smith et al., 2019). Dryland vegetation communities
111 are characterized by a seasonally dynamic mixture of annual and perennial grasses, shrubs and
112 trees growing individually or in small groups, and exposed bare ground (Brandt et al., 2020). Yet,
113 these heterogeneous mixtures of functional types are classified as simply grass- or shrub-
114 dominated in widely used land cover classifications including the National Land Cover Database
115 (NLCD) (Dewitz 2019). As a result, the importance of functional and structural diversity in
116 influencing the relationship between satellite proxies and ecosystem function has been
117 underexplored (Smith et al. 2019). The availability of new fractional vegetation cover products
118 that quantify heterogeneity in annual and perennial grasses, shrubs, trees, and bare ground at the
119 sub-pixel level provide a first opportunity to address this knowledge gap and quantify the
120 mediating role of functional and structural diversity in influencing the relationship between the
121 above identified satellite proxies and ecosystem GPP (Jones et al., 2018; Allred et al., 2020).

122 Here, we comprehensively evaluate NDVI, kNDVI, NIR_v, SAVI, and SIF in their ability to
123 capture seasonal dryland GPP dynamics. As a performance benchmark, we used GPP estimates
124 from a robust network of 21 western US eddy covariance tower sites that span representative
125 gradients in dryland ecosystem climate and functional composition. We further integrate a new
126 fractional vegetation cover product, the Rangeland Analysis Product (Allred et al., 2020), to fully
127 classify each site by its functional type heterogeneity. Our over-arching research questions are: (1)
128 how well do reflectance-based proxies (NDVI, SAVI, NIR_v, and kNDVI) and a physiologically-
129 based proxy (SIF) track seasonal GPP dynamics across dryland ecosystems? (2) how does
130 functional heterogeneity of dryland ecosystems mediate the performance of these two distinct
131 classes of vegetation GPP proxies? (3) can a simply integrated proxy perform best by combining
132 the independent strengths of reflectance-based and physiologically-based proxies?

133 **2. Methods**

134 *2.1 Study area and land cover classification*

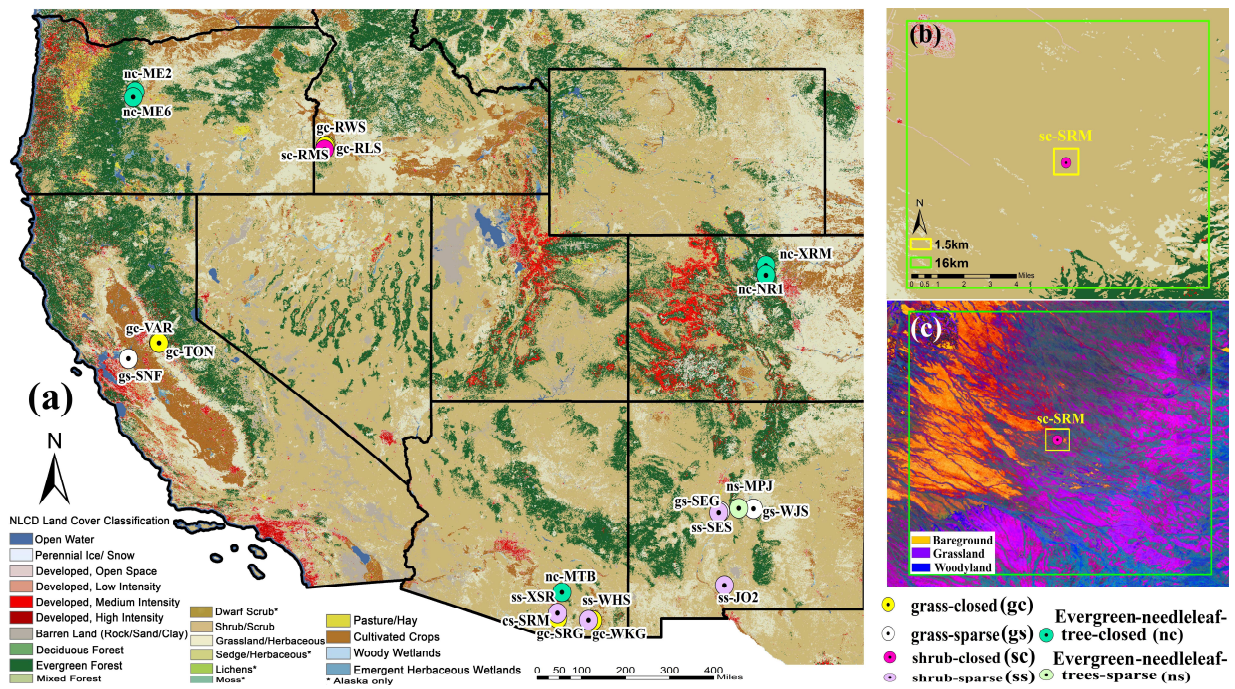
135 We used GPP estimates from 21 EC sites distributed across ecological sub-regions of the western
136 US (Fig. 1, Table S1) (Biederman et al., 2017; Smith et al., 2018). Daily GPP was averaged from
137 half-hour observations of GPP at each EC site between April 2018 and December 2019 (Fig. 1,
138 Table S1; ameriflux.lbl.gov). The widely used partitioning algorithm (Reichstein 2005; Lasslop
139 et al., 2010) was used to partition the net ecosystem exchange of CO₂ flux into GPP and
140 ecosystem respiration.

141 EC sites were classified as shrub-sparse, shrub-closed, grass-sparse, grass-closed, evergreen-
142 needleleaf-tree-sparse, and evergreen-needleleaf-tree-closed using 30-meter fractional land cover
143 estimates for the year 2019 from the Rangeland Cover V2.0 data product (Allred et al., 2020)

144 (Fig. 1, Table S1). We first calculated the mean fractional cover of bare ground, grass, shrub, and
145 tree within a 1.5-km² and 16-km² area around each EC site. We note that sites may have some
146 photosynthetic soil cover (i.e., biological soil crusts; Ferrenberg et al., 2017), and this cover was
147 categorized as bare ground. While incorporating biological soil crust cover may help improve
148 dryland GPP assessments in the future (Smith et al., 2019), those data do not currently exist and
149 are beyond the scope of this study.

150 Sites with tree cover as the largest fractional cover within 1.5-km² of the EC tower were classified
151 as tree-dominant sites. We used the ratio between woody (tree+shrub) and total vegetation
152 coverage (Ratio1) to define shrub-dominant and grass-dominant sites. Sites that were not tree-
153 dominated and Ratio1 > 0.4 were classified as shrub-dominated; otherwise, sites that were not
154 tree-dominated and with Ratio1 ≤ 0.4 were classified as grass-dominated. We then used the ratio
155 between bare ground and vegetation (Ratio2) to define the vegetation openness. If Ratio2 > 0.2,
156 the site was considered densely vegetated (“-closed”); otherwise the site was considered to be
157 sparsely vegetated (“-sparse”). The sparse site with mean GPP < 1 μmol CO₂ m⁻² s⁻¹ in the period
158 of April 2018 to December 2019 was further considered as low-productivity, otherwise, the site
159 was considered as high-productivity. In total, this resulted in 9 grass and 6 shrub sites with 8
160 closed and 7 sparse, as well as 7 low-productivity and 8 high-productivity for a well-balanced
161 classification of the available sites (Fig. 1, Table S2). Based on the above criteria, we classified
162 EC sites RWS, WKG, SRG, TON, VAR, and RLS as grass-closed (gc); SEG, SNF, and WJS as
163 grass-sparse (gs); RMS and SRM as shrub-closed (sc); SES, JO2, XSR, and WHS as shrub-sparse
164 (ss); ME2, MTB, ME6, XRM, and NR1 as evergreen-needleleaf-tree-closed (nc); as well as MPJ
165 as evergreen-needleleaf-tree-sparse (ns) (Fig. 1, Table S2). Hereafter, we also included the prefix
166 of the vegetation classification abbreviations in the site name to be more informative. We note

167 that nc-ME6 and gs-WJS are not typically classified as closed and grass-sparse ecosystems,
 168 respectively (Law and Waring, 2011), which could result either from our classification method or
 169 classification errors in the Rangeland Cover V2.0 product. In addition, the Rangeland cover
 170 product does not include evergreen vegetation cover, which we instead obtained from the 2016
 171 NLCD product (<https://www.mrlc.gov/data/nlcd-2016-land-cover-conus>).



173 **Figure 1.** (a) Locations of the 21 eddy covariance tower (EC) sites distributed across the western
 174 US. Abbreviations correspond to the AmeriFlux network site codes and symbol colors denote
 175 prevailing ecosystem type. Base map shows the 2016 Landsat-based National Land Cover
 176 Database (NLCD). (b) The sc-SRM EC site with 1.5-km² and 16-km² buffers against the 2016
 177 NLCD data product background. (c) The sc-SRM EC site with 1.5-km² and 16-km² buffers
 178 against an RGB composite background showing the three major vegetation covers: 11% bare
 179 ground, 28% grass, and 29% woody (shrub+tree) within a 1.5-km² grid and 14% bare ground, 26%
 180 grass, and 30% woody covers within a 16-km² grid.

181
 182 *2.2 MODIS vegetation reflectance-based proxies*

183 We calculated the NDVI, kNDVI, NIR_v, and SAVI from daily 500-m resolution nadir BRDF-
 184 adjusted reflectance (MCD43A4) data, which we refer to reflectance-based proxies in this study.

185 The data quality was controlled by removing all pixels that were not classified as either “good”
186 (QA=0) or “marginal” (QA=1) quality in the QA layer. NDVI is defined as the normalized
187 difference between near-infrared (NIR: MCD43A4 Band 2) and red (Red: MCD43A4 Band 1)
188 reflectance (Sellers et al., 1992):

$$189 \quad NDVI = \frac{(NIR-Red)}{(NIR+Red)} \quad (1)$$

190 kNDVI is a nonlinear generalization of NDVI (Camps-Valls et al., 2021):

$$191 \quad kNDVI = \tanh\left(\left(\frac{NIR-Red}{2\sigma}\right)^2\right) \quad (2)$$

192 where σ is a tunable length-scale parameter intended to capture non-linear sensitivity of NDVI to
193 vegetation density. Following Camps-Valls et al. (2021), we use the generalization $\sigma = 0.5(NIR +$
194 $red)$, which simplifies Equation (2) to $kNDVI = \tanh((NDVI)^2)$.

195 NIR_v is designed to better represent the depth-distribution of canopy photosynthetic capacity and
196 is calculated as the product of the NDVI and NIR bands (Badgley et al., 2017):

$$197 \quad NIR_v = (NDVI - 0.08) \times NIR \quad (3)$$

198 SAVI accounts for differential red and near-infrared extinction through the vegetation canopy
199 and utilizes a transformation technique to minimize soil brightness influences (Huete, 1998):

$$200 \quad SAVI = \frac{(1+L) \times (NIR-Red)}{(NIR+Red+L)} \quad (4)$$

201 where L is a soil brightness correction factor. The value of L varies with the amount or cover of
202 green vegetation: in very high vegetation regions, $L = 0$, and in areas with no green vegetation, L
203 $= 1$. Here, we used $L = 0.5$ (the default value), which works well in most situations (Richardson
204 and Everitt, 1992).

205 *2.3 TROPOMI SIF data*

206 We obtained daily TROPOMI SIF data between April 2018 and December 2019 from
207 <ftp://fluo.gps.caltech.edu/data/tropomi/ungridded/>. TROPOMI ground pixels are 5.6-km along-
208 track and vary from 3.5-14km across-track (nadir to 60° viewing angle). The data were pre-
209 processed using a data-driven approach to separate SIF emissions from the reflected solar
210 radiation. The data were subsequently restricted to 20–200 mW m⁻²sr⁻¹nm⁻¹ and filtered for
211 pixels with visible infrared imaging radiometer suite cloud fractions larger than 0.8 and view
212 zenith angles (VZAs) lower than 60 degrees. We converted instantaneous SIF observations to
213 integrated daily SIF estimates by multiplying instantaneous SIF by the daily correction factor that
214 was made available with the data product (Köhler et al., 2018; Doughty et al., 2019). We
215 averaged all available SIF retrievals within a 16-km² grid centered on each EC site (Fig. 1).

216 *2.4 SIF downscaling*

217 While the spatial resolution of TROPOMI SIF (5.6-km × 3.5-km at nadir) is relatively fine
218 compared to retrievals from previous sensors (e.g. GOME-2), there remains a considerable spatial
219 mismatch between TROPOMI and the average footprint of EC sites (Turner et al., 2020; Chu et
220 al., 2021). To enhance the spatial resolution of TROPOMI SIF observations, Turner et al. (2020)
221 developed a spatial downscaling approach that statistically integrated TROPOMI SIF and
222 MODIS NIR_v observations to generate a 500-m SIF product (SIF_NIR_v). SIF_NIR_v was
223 demonstrated to be an effective GPP proxy capable of accurately detecting interannual GPP
224 variability across the conterminous US (Turner et al., 2020, 2021). We therefore also combined
225 the SIF and NIR_v signals into a downscaled SIF product at 500m resolution that we refer as
226 SIF_NIR_v (Turner et al., 2020):

227 $(SIF_NIR_v)_{i,j} = SIF \times ((NIR_v)_{i,j} \div NIR_v)$ (5)

228 where i and j indicate the row and column number of each pixel in the 16-km² grid, $(SIF_NIR_v)_{i,j}$
229 and $(NIR_v)_{i,j}$ are SIF_NIR_v and NIR_v for each MODIS cell, and SIF and NIR_v are means of
230 $(SIF_NIR_v)_{i,j}$ and $(NIR_v)_{i,j}$ within the 16 km² footprint. For each EC tower, $(SIF_NIR_v)_{i,j}$ was
231 calculated for each MODIS cell within the 16-km² grid, then the mean SIF_NIR_v was calculated
232 as the average $(SIF_NIR_v)_{i,j}$ within a 1.5-km² area centered on the flux tower, in order to most
233 accurately match the tower flux footprint (Fig. S1; Chu et al., 2021).

234 *2.5 Comparison of satellite-based proxies with eddy covariance GPP*

235 We calculated the daily averages of NDVI, kNDVI, NIR_v , SAVI and SIF_NIR_v in a 1.5-km² area
236 centered on each EC site, and TROPOMI SIF across the corresponding 16-km² grid. Within the
237 larger grid, we only retained TROPOMI SIF retrievals with the same vegetation cover as the
238 target EC site. Averages of all vegetation covers were calculated in the 1.5-km² and 16-km² grid
239 using 30-m rangeland cover v2.0 data (<https://rangelands.app/>).

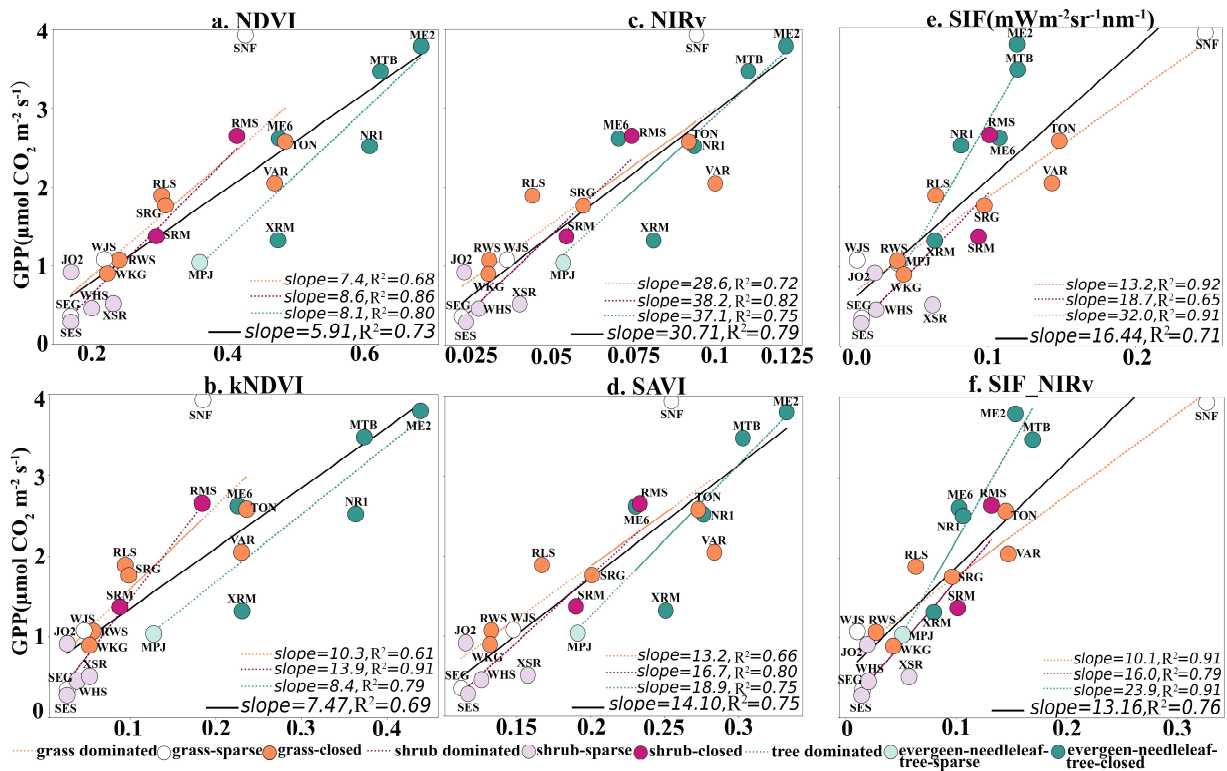
240 For the across-site analyses, we used daily GPP, NDVI, kNDVI, NIR_v , SAVI, SIF, and SIF_NIR_v
241 at each site over the 21-month (full record) study period from April 2018 to December 2019 as
242 inputs into vegetation class-specific linear regression models. To investigate seasonal dynamics,
243 we calculated approximately monthly (± 14 -days moving window) mean time series as inputs to
244 EC site-specific linear regression models. To better evaluate proxy performance at EC sites,
245 across which all proxies were relatively well correlated with seasonal GPP dynamics ($R^2 > 0.4$),
246 we further analyzed the seasonal bias in the proxy: GPP relationship. We focused six grass-
247 dominant sites plus sc-SRM with mixed grass and woody covers (grass > 28%, Ratio 1 < 50%,
248 Table S2), we reported the seasonal bias for each proxy as the average of residuals from each
249 linear fit during pre- and post-peak periods that were defined by the timing of peak GPP during

250 the major growing season in both 2018 and 2019. Note sc-SRM is here defined as shrub-closed
251 (similar to other sources) but has nearly identical grass and woody covers (~28% each), which
252 justifies its inclusion in this analysis. We then plotted the absolute values of the seasonal biases
253 for two categories of woody vegetation (shrub+tree) fraction ((0, 30%), (30%,50%)) to
254 understand the influence of woody cover on seasonal biases across grass-dominant sites. We
255 additionally calculated these seasonal biases at weekly (± 4 -days moving window) and biweekly
256 (± 7 -days moving window) scales to assess their sensitivity to temporal scale. We excluded shrub-
257 and tree-dominated sites from this hysteresis analysis since only two sites from each category
258 satisfied criteria for these analyses.

259 **3. Results**

260 *3.1 Across-site spatial evaluation of dryland GPP proxies*

261 All proxies captured the across-site spatial variability of GPP, with R^2 values ranging from 0.69
262 for kNDVI to 0.79 for NIR_v (Fig. 2). NDVI and kNDVI (Figs. 2a, 2b), NIR_v and SAVI (Figs. 2c,
263 2d), SIF and NIR_v -downscaled SIF (SIF_ NIR_v) (Figs. 2e, 2f) performed strikingly similarly in
264 their ability to capture GPP across ecosystem types. Notably, compared to the four spectral
265 reflectance-based indices, the SIF:GPP and SIF_ NIR_v :GPP relationship appeared to vary as a
266 function of ecosystem type. For example, for SIF:GPP, the slope of the relationship across
267 evergreen-needleleaf-tree-dominant sites ($32.0 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}/\text{mW m}^{-2} \text{ nm}^{-1} \text{ sr}^{-1}$) was more
268 than double the slope of the relationship across grass-dominated sites ($13.2 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}/\text{mW}$
269 $\text{m}^{-2} \text{ nm}^{-1} \text{ sr}^{-1}$) (Figs. 2e, 2f).



270

271 **Figure 2.** A comparison of mean EC GPP and (a) mean NDVI (1.5-km²), (b) mean kNDVI (1.5-
 272 km²), (c) mean NIR_v (1.5-km²), (d) mean SAVI (1.5-km²), (e) mean SIF (16-km²), and (f) mean
 273 SIF downscaled by NIR_v (SIF_NIR_v; 1.5-km²) over the study period from April 2018 to
 274 December 2019 across shrub-dominated, grass-dominated, and evergreen-needleleaf- tree-
 275 dominated sites.

276 3.2 Within-site seasonal evaluation of dryland GPP proxies

277 3.2.1 Reflectance-based proxies: NDVI, kNDVI, NIR_v, and SAVI

278 The monthly NDVI:GPP relationship across all sites (mean $R^2 = 0.41$) was the lowest of all the
 279 reflectance-based indices (Fig. 3a). Specifically, NDVI had the lowest average R^2 in grass-
 280 dominated (mean $R^2 = 0.60$), shrub-dominated (mean $R^2 = 0.31$), and evergreen-needleleaf-tree-
 281 dominated (mean $R^2 = 0.22$) sites. The correlation between NDVI and GPP was insensitive to
 282 temporal aggregation at weekly, biweekly, and monthly time scales (Figs. 3a, S2a, S2b). The
 283 monthly kNDVI:GPP relationship across all sites (mean $R^2 = 0.43$) was strikingly similar to the
 284 NDVI:GPP relationship (Fig. 3b), though R^2 at grass- and shrub- dominated sites was slightly

285 higher for kNDVI (mean $R^2 = 0.64$ and 0.37 , respectively) relative to NDVI (mean $R^2 = 0.60$ and
286 0.31 , respectively).

287 The monthly NIR_v :GPP relationship across all sites (mean $R^2 = 0.57$) was improved across all
288 vegetation classes relative to NDVI and kNDVI (Fig. 3c). Notably, the NIR_v :GPP relationship
289 was relatively low across the evergreen-needleleaf-tree- and evergreen-shrub-dominated sites (ns-
290 MPJ, nc-VCP, nc-ME6, nc-MTB, nc-ME2, and ss-Jo2), with R^2 ranging from 0.02 to 0.35 (Fig.
291 3c). Excluding evergreen-dominated sites, the NIR_v :GPP relationship increased significantly
292 (mean $R^2 = 0.71$). NIR_v performed best as an indicator of monthly GPP in grass-closed and
293 shrub-closed sites (mean $R^2 = 0.79$). The correlation between NIR_v and GPP was insensitive to
294 temporal aggregation at weekly, biweekly, and monthly time scales (Figs. 3c, S2c, S2d). The
295 monthly SAVI:GPP relationship across all sites (mean $R^2 = 0.57$) was strikingly similar to the
296 NIR_v :GPP relationship (Fig. 3d).

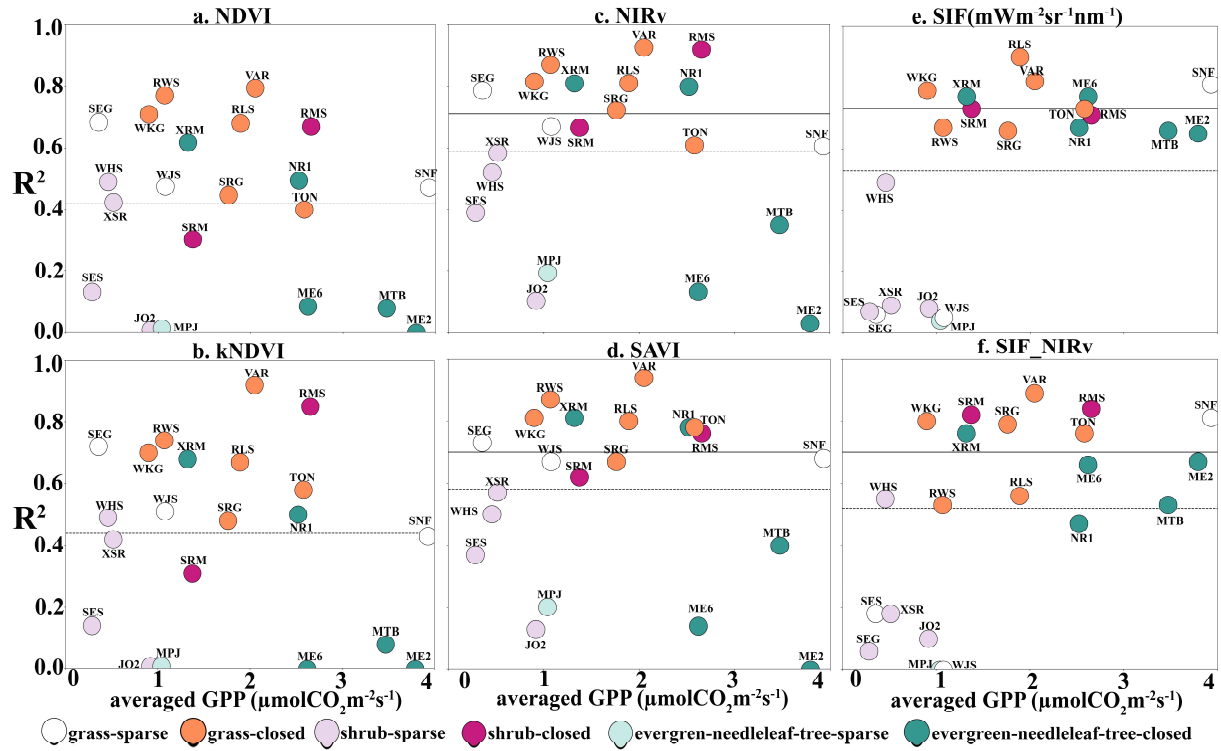
297 Across seven grass-closed sites, the relationship between reflectance-based proxies (i.e., NDVI,
298 kNDVI, NIR_v , and SAVI) and GPP was characterized by significant seasonal hysteresis, which
299 we defined as the averaged bias in the relationships during the pre- and post-peak periods (Table
300 S3). For example, pre-peak biases were -0.50 and $-1.18 \text{ umol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ and post-peak biases
301 were 0.36 , and $0.75 \text{ umol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ in the monthly NDVI:GPP relationship at gc-WKG (Fig.4)
302 and sc-SRM (Fig. 5), respectively. These seasonal hysteresis effects were significantly larger at
303 sites with 30-50% fractional woody plant cover (e.g., sc-SRM) than at sites with 0-30% fractional
304 woody plant cover (e.g., gs-WKG) (Figs. 4-6). Across all sites, the seasonal hysteresis of the
305 NDVI:GPP relationship (mean bias= 0.27 and $0.97 \text{ umol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for 0-30% and 30%-50%
306 woody cover, respectively) was again similar to the kNDVI:GPP relationship (mean bias= 0.24
307 and $0.81 \text{ umol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for 0-30% and 30%-50% woody cover, respectively) (Fig. 6). Notably,

308 the seasonal NIR_v:GPP and SAVI:GPP hysteresis estimates were also strikingly similar and less
309 pronounced than the NDVI:GPP and kNDVI:GPP hysteresis estimates(Figs. 4-6). For example,
310 pre-peak biases were reduced to -0.29 and -0.76 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ and post-peak biases were
311 reduced to 0.25 and 0.58 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ in the monthly NIR_v: GPP relationship at gc-WKG
312 (Fig.4) and sc-SRM (Fig. 5), respectively. The NDVI:GPP, kNDVI:GPP, NIR_v:GPP, and
313 SAVI:GPP relationships and their seasonal hysteresis characters were insensitive to temporal
314 aggregation at weekly, biweekly, and monthly time scales (Figs. 3, 6, S2, S3).

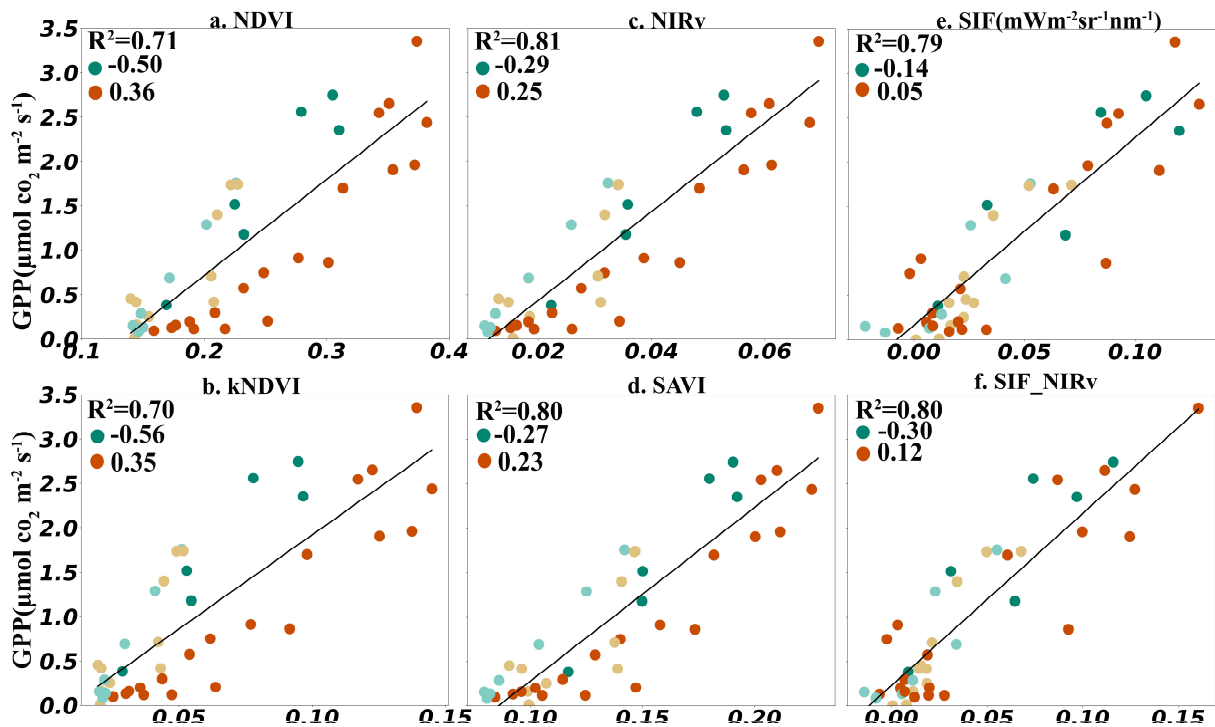
315 3.2.2 SIF and SIF_NIR_v

316 The correlation between monthly SIF and monthly GPP across most sites (mean $R^2 = 0.53$) and
317 all vegetation classes was improved relative to NDVI (Fig. 3e). Notably, the seasonal relationship
318 between SIF and GPP was relatively weak (mean $R^2 = 0.13$) across low-productivity sites
319 characterized by average GPP < 1 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ and significant bare ground relative to
320 vegetation coverage (Ratio 2 > 0.2), including gs-SEG, gs-WJS, ss-WHS, ss-JO2, ss-SES, ss-
321 XSR and ns-MPJ (Fig. 3e, Table S2). Also, unlike the four reflectance-based proxies, temporal
322 aggregation had a large effect on the strength of the SIF:GPP relationship, with correlation values
323 increasing from weekly to monthly time scales (Figs. 3e, S2e, S2f). After excluding sparse sites,
324 the monthly SIF:GPP relationship increased (mean $R^2 = 0.74$). The SIF:GPP relationship had low
325 seasonal hysteresis, with pre-peak biases of -0.14 and -0.20 and post-peak biases of 0.05 and 0.11
326 ($\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) for gs-WKG and sc-SRM respectively (Figs. 4-5). Notably, seasonal biases
327 were not significantly different between the two classes of woody (tree+shrub) fractional cover
328 and these findings were preserved at monthly, biweekly, and weekly timescale (Figs. 6, S3).
329 Compared to the SIF:GPP relationship, the SIF_NIR_v:GPP seasonal R^2 values were lower across

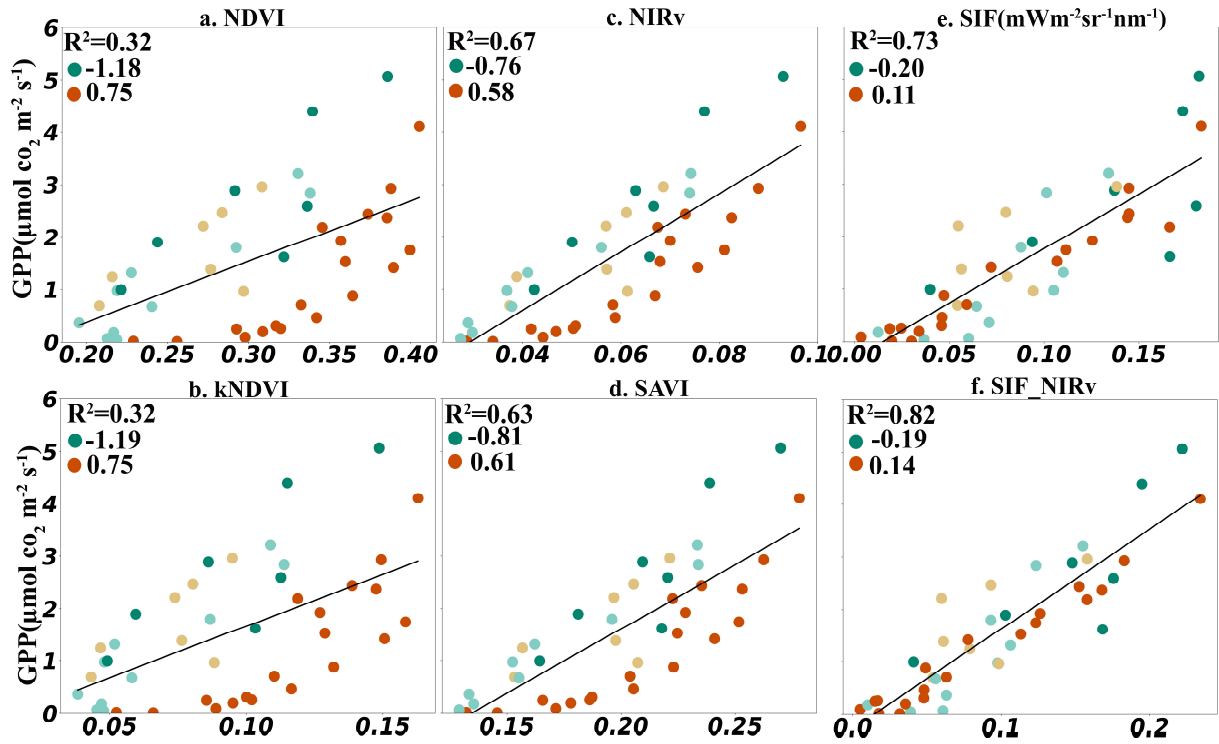
330 all evergreen-needleleaf-tree sites but slightly higher and with reduced seasonal biases at six out
 331 of eight non-evergreen, high-productivity sites (Figs. 3-6, S3).



333 **Figure 3.** The coefficient of determination (R^2) for 21 dryland EC sites derived from monthly
 334 GPP and monthly satellite proxies plotted against the mean GPP measured at the site over the full
 335 study period. Comparison of EC GPP and (a) NDVI (1.5-km²), (b) kNDVI (1.5-km²), (c) NIR_v
 336 (1.5-km²), (d) SAVI (1.5-km²), (e) SIF (16-km²) and (f) SIF_NIR_v (1.5-km²) are color coded by
 337 grass-sparse, grass-closed, shrub-sparse, shrub-closed, evergreen-needleleaf-tree-sparse, and
 338 evergreen-needleleaf-tree-dominated sites. The dashed lines denote the means of the R^2 values of
 339 all sites. The solid lines denote the mean R^2 when excluding all evergreen-needleleaf-tree sites
 340 for NIR_v and SAVI, and all sparse sites for SIF and SIF_NIR_v.



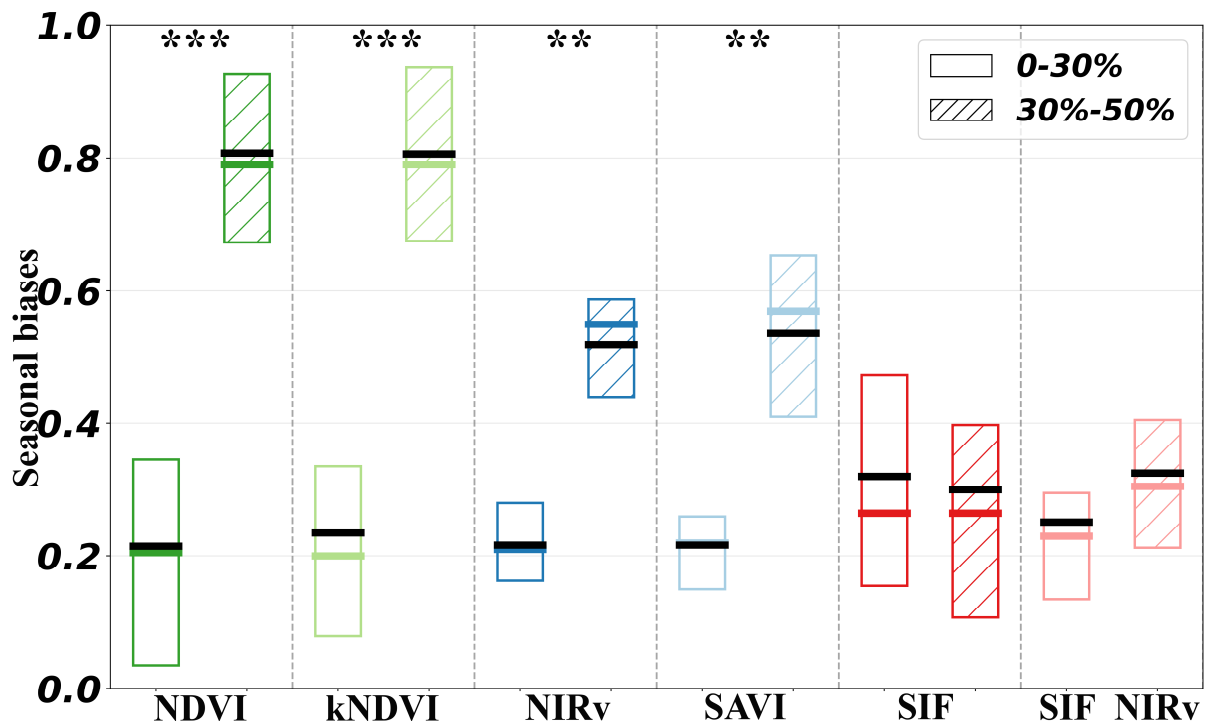
341
 342 **Figure 4.** The relationship between EC GPP and satellite NDVI(a), kNDVI(b), NIR_v(c), SAVI(d),
 343 SIF(e), and SIF_NIR_v(f) for a representative closed grassland site (gc-WKG; 23% fractional
 344 woody cover). Green and brown circles represent the pre-peak and post-peak periods,
 345 respectively. Seasonal biases were calculated as the mean of the residuals from each linear fit
 346 during pre-peak and post-peak periods as reported in the legend. Gc-WKG was characterized by
 347 bi-modal growing seasons during both 2018 and 2019 and pre-peak and post-peak bias is
 348 reported for both periods; light green and light brown dots represent the first growing season and
 349 dark green and orange dots represent the second growing season.



350
 351 **Figure 5.** The relationship between EC GPP and satellite NDVI(a), kNDVI(b), NIR_v(c), SAVI(d),
 352 SIF(e), and SIF_NIR_v(f) for a representative closed shrubland site (sc-SRM; 50% fractional
 353 woody cover). Green and brown circles represent the pre-peak and post-peak periods,
 354 respectively. Seasonal biases were calculated as the mean of the residuals from each linear fit
 355 during pre-peak and post-peak periods as reported in the legend. Sc-SRM was characterized by
 356 bi-modal growing seasons during both 2018 and 2019 and pre-peak and post-peak bias is
 357 reported for both periods; light green and light brown dots represent the first growing season and
 358 dark green and orange dots represent the second growing season.

359

360



361
 362 **Figure 6.** Seasonal biases in the monthly NDVI:GPP, kNDVI:GPP, NIR_v:GPP, SAVI:GPP,
 363 SIF:GPP and SIF_NIR_v:GPP relationships across two fractional woody (shrub+tree) cover
 364 classes (0-30% and 30%-50%). Seasonal biases represent the absolute value of averaged residuals
 365 of the linear regression fit during pre- and post-season (Table S3). Fractional shrub and tree cover
 366 is defined according to Ratio 1 [(shrub+tree) / total vegetation] (Table S2). "****" denotes
 367 significant differences between paired distributions based on a standard ANOVA test with $p <$
 368 0.001. "***" denotes significant differences between paired distributions based on a standard
 369 ANOVA test with $p <$ 0.01. The bars represent the 25th-50th -75th percentile of the data points in
 370 each relationship; the black lines represent the means of seasonal bias in each relationship.

371

372 4. Discussion

373 4.1 Across-site spatial evaluation of dryland GPP proxies

374 All satellite-based proxies captured across-site spatial variability of GPP reasonably well across
 375 sites classified by dominant functional types (Fig. 2). NDVI and kNDVI performed strikingly
 376 similarly, as did NIR_v and SAVI, in their ability to capture GPP across sites. We note, however,
 377 that we applied the default nonlinear sensitivity factor (σ) and soil brightness correction factor (L)
 378 for kNDVI and SAVI, respectively. There remains a need for future research that explores the

379 sensitivity of these indices to their respective correction factors (Zhao et al., 2018; Camps-Valls
380 et al., 2021).

381 We found that the SIF:GPP relationship was relatively sensitive to ecosystem type, with
382 improved correlation when ecosystem types were separated (Fig. 2). Two potential explanations
383 for the observed ecosystem-specific sensitivities of SIF:GPP include: 1) differential SIF emission
384 and scattering due to differences in canopy structures (Migliavacca et al., 2017; Qiu et al., 2019;
385 Dechant et al., 2020); and 2) differential SIF emission per unit CO₂ uptake driven by variation in
386 the dominant photosynthetic pathways (C3 vs. C4) (He et al., 2020). Notable, the C3-grass-
387 dominated sites had relatively lower SIF:GPP slopes compared to all other sites, possibly due to
388 higher rates of photorespiration in the C3 photosynthetic pathway (Fig. S4; Turner et al., 2021).
389 However, the C3-evergreen-needleleaf-tree-dominated sites had the highest SIF:GPP slopes,
390 which suggests that structure and re-absorption of SIF emissions may also play a dominant role in
391 SIF:GPP relationship (Fig. S4). Future research is needed to more thoroughly understand and
392 separate the physiological and structural factors that mediate the SIF:GPP relationship.

393 *4.2 Within-site seasonal evaluation of dryland GPP proxies*

394 NIR_v and SIF outperformed all other proxies in their ability to track seasonal GPP dynamics (Fig.
395 3), consistent with previous studies focused on the western US (Smith et al., 2018; Zuromski et
396 al., 2018), Australian drylands (Wang et al., 2019), and African savannas (Mengistu et al., 2020).
397 At low-productivity sites, e.g., grass-sparse and shrub-sparse sites (gs-WJS, gs-SEG, ss-WHS,
398 and ss-XSR), NIR_v consistently outperformed SIF at capturing seasonal GPP variation (Fig. 3,
399 Table S4). NIR_v was largely successful in isolating the vegetation signal from the confounding
400 effects of soil brightness (Templeton et al., 2014; Badgley et al., 2019; Baldocchi et al., 2020) as
401 demonstrated by its close correlation with SAVI (Fig. 3). SIF performed relatively poorly at

402 capturing seasonal GPP dynamics at low-productivity sites, likely due to the low signal-to-noise
403 ratio of SIF retrievals (Guanter et al., 2015; Köhler et al., 2021). In other words, at low
404 productivity sites, the true SIF signal is likely too weak to overcome the inherent noise associated
405 with the SIF retrieval (Köhler et al., 2021), resulting in very low correlation with seasonal GPP
406 dynamics. We also found evidence of low SIF signal-to-noise ratios during non-growing season
407 periods at high-productivity sites (gc-SRG, gc-VAR, gc-TON, gc-RWS, and sc-RMS), resulting
408 in large apparent pre- and post-peak biases (Fig. S5). Further sensitivity analysis in which the
409 temporal resolution was increased from monthly to weekly revealed that SIF:GPP R^2 values were
410 generally reduced (Figs. 3e, S2e, S2f) and SIF:GPP seasonal biases were general increased
411 relative to the reflectance-based proxies (Figs. 6, S3). This relatively high sensitivity to temporal
412 scale suggests that satellite SIF observations are more susceptible to noise at these higher
413 frequencies than reflectance-based proxies. Other factors including SIF signal re-absorption and
414 sun-sensor geometry could also drive noise in the SIF:GPP relationship and should be the focus
415 of future research efforts (Romero et al., 2018; Hao et al., 2020, 2021).

416 At closed evergreen-needleleaf-tree-dominated sites, SIF captured more seasonal GPP variability
417 ($R^2 = 0.70$) than NIR_v ($R^2 = 0.42$) (Fig. 3, Table S4). This is consistent with previous research
418 finding that SIF captures more than just seasonal changes in vegetation greenness and potentially
419 contains more information related to physiological function (e.g., photosynthetic rate, stomatal
420 regulation, photochemical quenching, etc.) (Joiner et al., 2014; Walther et al., 2016; Magney et
421 al., 2019). These relatively high correlations persist despite the coarser spatial resolution of
422 TROPOMI SIF observations, which is an order of magnitude larger than typical EC tower
423 footprints (Chu et al., 2021). While, we minimized the impact of this spatial mismatch by
424 filtering TROPOMI SIF observations to include only the vegetation cover types associated with a

425 given EC tower, it is likely that this simple filtering was inadequate in some cases, especially for
426 sites with relatively high heterogeneity around the EC tower site. Downscaling SIF by NIR_v
427 observations was attempted to better account for this spatial mismatch, but instead seemed to
428 transfer the existing limitations of NIR_v for tracking GPP seasonal variability (Fig. 3), and thus
429 SIF alone outperformed $\text{SIF}_{\text{NIR}_v}$ at evergreen-needleleaf-tree sites (Fig. 3, Table S4). More
430 accurate monitoring of heterogeneous dryland regions will require higher resolution SIF
431 observations that better isolate the signals of mixed vegetation functional types (Smith et al.,
432 2019). NIR_v did perform relatively well and similar to SIF at two neighboring evergreen forest
433 sites: Niwot Ridge Forest and Rocky Mountain National Park (nc-NR1 and nc-XRM, Fig. 3).
434 This finding is potentially explained by previous work demonstrating that seasonal changes in
435 canopy color, which can be tracked with reflectance-based proxies, correlate with seasonal GPP
436 for these ecosystems (Wu et al., 2014; Seyednasrollah et al., 2020).

437 SIF downscaled by NIR_v ($\text{SIF}_{\text{NIR}_v}$) slightly improved seasonal GPP estimates and reduced
438 seasonal bias relative to SIF alone at most grass-closed and shrub-closed sites (Figs. 3, 6, Tables
439 S3, S4). At these sites, the higher spatial resolution of NIR_v observations likely helped to reduce
440 noise inherent in the SIF signal and associated with footprint mismatch (Turner et al., 2020).
441 However, at low productivity sites (mean GPP < $\sim 1 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) with significant bare
442 ground cover, the SIF signal was too weak to overcome random noise in the retrieval, resulting in
443 the NIR_v observations outperforming $\text{SIF}_{\text{NIR}_v}$ (Fig. 3, Table S4). This finding is consistent with
444 recent work showing that $\text{SIF}_{\text{NIR}_v}$ explained less than 10% of GPP variability at shrub and
445 savanna EC tower sites (Turner et al., 2021). Here, by incorporating fractional land cover data
446 within a 1.5-km² and 16-km² grid at each EC site, we provide new evidence that the signal-to-
447 noise of the SIF retrieval is sensitive to fractional bare ground cover when the surface is brighter

448 than vegetation (Guanter et al., 2015), which is a common issue across drylands (Huete 1988;
449 Gholami Baghi and Oldeland, 2019). Other methods, such as machine learning (Gentine and
450 Alemohammad, 2018) or a semi-empirical downscaling based on light use efficiency (Duveiller
451 et al., 2020), should be explored in the future to build upon the respective strengths of NIR_v and
452 SIF while limiting their weaknesses.

453 *4.3 Reflectance-based GPP proxies are prone to seasonal hysteresis*

454 While NIR_v and SIF accurately captured seasonal GPP across grass-closed sites with mixed grass
455 and woody cover, we found significant patterns of seasonal hysteresis in the correlations between
456 GPP and reflectance-based vegetation proxies (NDVI, kNDVI, NIR_v, and SAVI) from weekly to
457 monthly scale (Figs. 4-6, S3, Table S3). These patterns differed depending on the proportion of
458 evergreen shrubs and trees in the tower footprint (Figs. 6, S3), suggesting that reflectance-based
459 proxies are seasonally decoupled from GPP for these vegetation types. In other words, evergreen
460 vegetation maintained the appearance of high photosynthetic capacity (i.e., remains green) even
461 during periods of low GPP (Smith et al., 2018; Yan et al., 2019; Knowles et al., 2020). Thus, it is
462 important to account for spatially heterogeneous mixtures of these functional types when using
463 reflectance-based proxies as GPP proxies or inputs of GPP models. SIF was much less impacted
464 by this type of seasonal bias and thus represents an improved proxy for seasonal variability of
465 GPP in many dryland ecosystems, especially at the end of the growing season (Figs. 6, S3; Wang
466 et al., 2020b).

467 Seasonal hysteresis in the relationships between reflectance-based proxies and GPP could also be
468 impacted by changing soil brightness during pre- and post- peak periods (Gitelson et al., 2014;
469 Flanagan et al., 2015; Nestola et al., 2016; Peng et al., 2017). For example, at low productivity
470 sites like gs-SEG and gs-WJS with more than 30% bare ground coverage (Fig S6, Table S3-1),

471 the bare ground fraction likely changes throughout the growing season as bare ground fills in
472 with annual cover species and this could also contribute to seasonal hysteresis effects. However,
473 across closed canopy sites with high productivity and less than 20% bare ground coverage, we
474 find evidence that seasonal hysteresis with GPP is driven by differences in the herbaceous and
475 woody vegetation fractions (Figs. 6, S3). This idea is supported by our findings at gc-WKG and
476 sc-SRM sites with potentially two growing seasons, where the seasonal hysteresis with GPP
477 mostly occurs during the second (dominant) growing season when bare ground coverage is
478 minimized (Figs. 4-5). Similarly, changes in nitrogen content, chlorophyll content, and vegetation
479 structure between the pre- and post-peak periods might also contribute to seasonal hysteresis
480 (Gitelson et al., 2014; Flanagan et al., 2015; Nestola et al., 2016; Peng et al., 2017). Notably, the
481 NIR_v:GPP relationship showed less seasonal hysteresis compared to NDVI:GPP likely due to its
482 reduced sensitivity to background soil brightness (Huete, 1988; Badgley et al., 2017), and
483 increased sensitivity to ecosystem structure (Table S3; Peng et al., 2017).

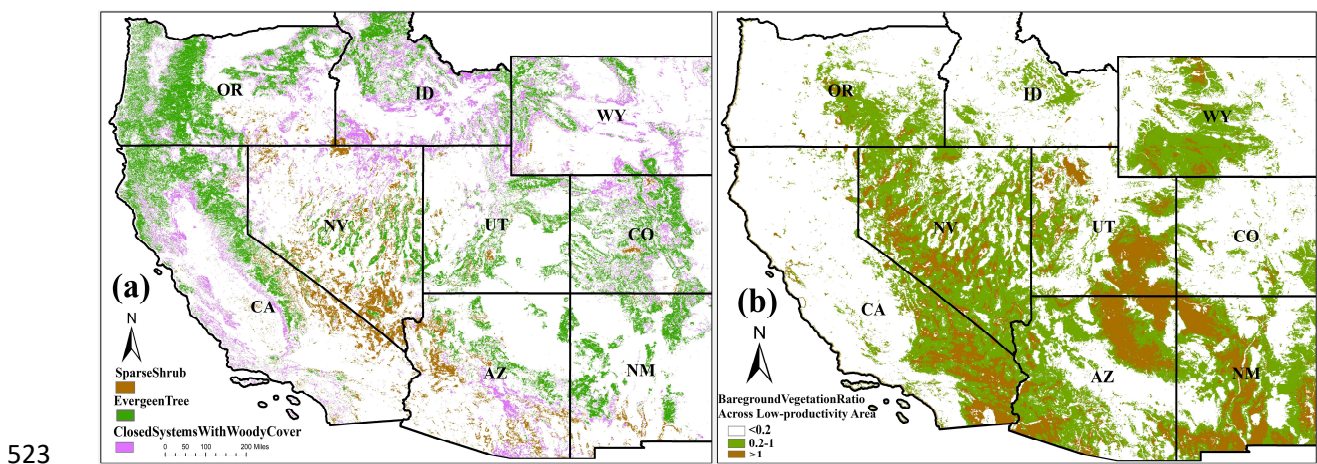
484 Seasonal hysteresis between reflectance-based proxies and GPP also differed across climate
485 zones. At the Mediterranean California sites (gc-TON and gs-SNF), reflectance-based vegetation
486 proxies (NDVI, kNDVI, NIR_v, and SAVI) overestimated GPP prior to the annual peak but
487 underestimated it thereafter (Table S3-1). However, at the North American Monsoon-affected
488 sites (the summer rainfall-dominated sites of gs-WJS, gs-SES, gc-WKG, gc-SRG, and sc-SRM),
489 the hysteresis pattern between reflectance-based vegetation proxies and GPP was reversed (Table
490 S3-1). At the Mediterranean-climate sites in California, the understory grasses distinctly green up
491 during the cool, wet season from October to April, but energy limits GPP during the winter,
492 resulting in greenness increasing ahead of GPP. In late spring, the shallow soil dries, understory
493 grasses brown, and greenness declines, but the overstory oaks thrive on deeper soil moisture such

494 that GPP stays elevated (Bartolome, 1979; Xu and Baldocchi, 2003; Ma et al., 2007; Liu et al.,
495 2017). In contrast, at Monsoon-affected sites, both grasses and shrubs green up and
496 photosynthesize during spring using stored winter moisture. Following the second, dominant
497 growing season driven by summer rainfall, overstory shrubs and trees can remain green for weeks
498 to months after soil moisture is depleted and GPP falls (Scott et al., 2009; Barron-Gafford et al.,
499 2017; Yan et al., 2019).

500 *4.4 Complementary application of reflectance-based proxies and SIF observations across* 501 *dryland ecosystems*

502 The reflectance-based proxies considered here estimated seasonal GPP dynamics relatively well
503 in homogeneous and non-evergreen areas but poorly in heterogeneous and/or evergreen regions.
504 On the contrary, SIF performed well in high-productivity regions and poorly in low-productivity
505 regions. These differences in skill are often complementary and can be potentially leveraged to
506 more accurately characterize GPP dynamics across much of the western US. Based on the current
507 analysis, we estimate that 25% of the semi-arid western US land area corresponds to
508 heterogeneous or evergreen regions (Fig. 7a), which will induce bias when using reflectance-
509 based proxies to estimate seasonal GPP. This area specifically includes 13% evergreen-
510 dominated regions, where reflectance-based proxies failed to capture seasonal GPP dynamics; 8%
511 grass-closed regions with more than 30% woody cover, where reflectance-based proxies had
512 seasonally hysteretic relationships to GPP; and 4% sparse shrub regions, where NIR_v only
513 captured around 50% of the variance in seasonal GPP dynamics (Fig. 7a). In contrast, TROPOMI
514 SIF failed to accurately estimate seasonal GPP across low productivity areas that account for 39%
515 of the semi-arid western US (Fig. 7b). These results underscore the importance of functional
516 heterogeneity in dryland ecosystems and demonstrate that generalizing across regions using a

517 single vegetation proxy, whether with a process-based or empirical model, will likely result in
 518 inaccurate and/or biased GPP estimates (Smith et al. 2019). Taken together, our findings indicate
 519 that different vegetation proxies are better suited for different dryland ecosystem types, and
 520 suggest that data integration approaches, particularly those focused on integrating SIF and NIR_v
 521 observations, are critical to improved performance of satellite-based GPP models across drylands
 522 from the region to the globe.



524 **Figure 7.** Areas in the semi-arid western US where (a) TROPOMI SIF would be expected to
 525 outperform MODIS NDVI/NIR_v and (b) MODIS NDVI/NIR_v would be expected to outperform
 526 TROPOMI SIF at estimating GPP seasonal dynamics. (a) Purple pixels are grass-closed regions
 527 with significant tree covers (30% to 50%), green pixels are dominated by evergreen trees, and
 528 grey pixels are dominated by sparse shrubs. (b) The low-productivity regions (colored pixels) are
 529 defined as average GPP $< \sim 1 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ (or averaged $\text{NIR}_v < 0.04$), where the ratio
 530 between bare ground and vegetation coverage is greater than 0.2.

531

532 5. Conclusions

533 Accurate detection of seasonal to interannual variability of GPP in drylands is complicated in part
 534 by the highly heterogeneous mixtures of bare ground, grass, shrubs, and trees characteristic of
 535 dryland ecosystems. Here, we evaluated the skill of NDVI , kNDVI , NIR_v , SAVI , SIF , and
 536 $\text{SIF}_{\text{NIR}_v}$ to predict GPP dynamics as measured by 21 eddy covariance tower sites across six

537 major dryland classes: grass-sparse, grass-closed, shrub-sparse, shrub-closed, evergreen-
538 needleleaf-tree-sparse, and evergreen-needleleaf-tree-closed of the western US. NIR_v and SIF
539 were found to perform best in capturing both spatial patterns and seasonal dynamics of GPP, and
540 further captured complementary aspects of seasonal GPP dynamics: NIR_v was the best GPP
541 proxy across non-evergreen vegetation-sparse sites, while SIF was the best GPP proxy across
542 evergreen and closed-canopy sites. We also found significant seasonal bias and hysteresis in the
543 relationships between the reflectance-based proxies considered here and GPP as a function of
544 increasing fractional shrub and tree coverage across grass-dominated sites. Based on our findings,
545 NIR_v likely has significant limitations for approximating GPP across 25% of the western US
546 (high heterogeneity and/or evergreen vegetation), while SIF has significant limitations for
547 approximating GPP across 39% of the western US (low productivity sites with a significant
548 portion of bare ground). Our research indicates that use of a single proxy or method to
549 characterize vegetation dynamics across dryland areas will likely result in biased estimates of
550 GPP in at least one part of the growing season. We suggest careful consideration of vegetation
551 heterogeneity when integrating different proxies or methods for improved representation of
552 vegetation dynamics across dryland regions.

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561 available via <ftp://fluo.gps.caltech.edu/data/tropomi/ungridded/>. Eddy covariance half-hour
562 observations of the net ecosystem exchange of CO₂ are available via ameriflux.lbl.gov through
563 funding from the U.S. Department of Energy's Office of Science. Vegetation fractional cover
564 data are available via <https://rangelands.app/>. Any use of trade, firm, or product names is for
565 descriptive purposes only and does not imply endorsement by the U.S. Government.

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838 **List of Figure Captions:**

839 **Figure 1.** (a) Locations of the 21 eddy covariance tower (EC) sites distributed across the western
840 United States. Abbreviations correspond to the AmeriFlux network site codes and symbol colors
841 denote prevailing ecosystem type. Base map shows the 2016 Landsat-based National Land Cover
842 Database (NLCD). (b) The sc-SRM EC site with 1.5-km² and 16-km² buffers against the 2016
843 NLCD data product background. (c) The sc-SRM EC site with 1.5-km² and 16-km² buffers
844 against an RGB composite background showing the three major vegetation covers: 11% bare
845 ground, 28% grass, and 29% woody (shrub+tree) within a 1.5-km² grid and 14% bare ground, 26%
846 grass, and 30% woody covers within a 16-km² grid.

847

848 **Figure 2.** A comparison of mean EC GPP and (a) mean NDVI (1.5-km²), (b) mean kNDVI (1.5-
849 km²), (c) mean NIR_v (1.5-km²), (d) mean SAVI (1.5-km²), (e) mean SIF (16-km²), and (f) mean
850 SIF downscaled by NIR_v (SIF_NIR_v; 1.5-km²) over the study period from April 2018 to
851 December 2019 across shrub-dominated, grass-dominated, and evergreen-needleleaf- tree-
852 dominated sites.

853

854 **Figure 3.** The coefficient of determination (R^2) for 21 dryland EC sites derived from monthly
855 GPP and monthly satellite proxies plotted against the mean GPP measured at the site over the full
856 study period. Comparison of EC GPP and (a) NDVI (1.5-km²), (b) kNDVI (1.5-km²), (c) NIR_v
857 (1.5-km²), (d) SAVI (1.5-km²), (e) SIF (16-km²) and (f) SIF_NIR_v (1.5-km²) are color coded by
858 grass-sparse, grass-closed, shrub-sparse, shrub-closed, evergreen-needleleaf-tree-sparse, and
859 evergreen-needleleaf-tree-dominated sites. The dashed lines denote the means of the R^2 values of
860 all sites. The solid lines denote the mean R^2 when excluding all evergreen-needleleaf-tree sites
861 for NIR_v and SAVI, and all sparse sites for SIF and SIF_NIR_v.

862

863 **Figure 4.** The relationship between EC GPP and satellite NDVI(a), kNDVI(b), NIR_v(c), SAVI(d),
864 SIF(e), and SIF_NIR_v(f) for a representative closed grassland site (gc-WKG; 23% fractional
865 woody cover). Green and brown circles represent the pre-peak and post-peak periods,
866 respectively. Seasonal biases were calculated as the mean of the residuals from each linear fit
867 during pre-peak and post-peak periods as reported in the legend. Gc-WKG was characterized by
868 bi-modal growing seasons during both 2018 and 2019 and pre-peak and post-peak bias is
869 reported for both periods; light green and light brown dots represent the first growing season and
870 dark green and orange dots represent the second growing season.

871

872 **Figure 5.** The relationship between EC GPP and satellite NDVI(a), kNDVI(b), NIR_v(c), SAVI(d),
873 SIF(e), and SIF_NIR_v(f) for a representative closed shrubland site (sc-SRM; 50% fractional
874 woody cover). Green and brown circles represent the pre-peak and post-peak periods,
875 respectively. Seasonal biases were calculated as the mean of the residuals from each linear fit
876 during pre-peak and post-peak periods as reported in the legend. Sc-SRM was characterized by

877 bi-modal growing seasons during both 2018 and 2019 and pre-peak and post-peak bias is
878 reported for both periods; light green and light brown dots represent the first growing season and
879 dark green and orange dots represent the second growing season.

880

881 **Figure 6.** Seasonal biases in the monthly NDVI:GPP, kNDVI:GPP, NIR_v:GPP, SAVI:GPP,
882 SIF:GPP and SIF_NIR_v:GPP relationships across two fractional woody (shrub+tree) cover
883 classes (0-30% and 30%-50%). Seasonal biases represent the absolute value of averaged residuals
884 of the linear regression fit during pre- and post-season (Table S3). Fractional shrub and tree cover
885 is defined according to Ratio 1 [(shrub+tree) / total vegetation] (Table S2). "***" denotes
886 significant differences between paired distributions based on a standard ANOVA test with $p <$
887 0.001. "**" denotes significant differences between paired distributions based on a standard
888 ANOVA test with $p <$ 0.01. The bars represent the 25th-50th -75th percentile of the data points in
889 each relationship; the black lines represent the means of seasonal bias in each relationship.

890

891 **Figure 7.** Areas in the semi-arid western US where (a) TROPOMI SIF would be expected to
892 outperform MODIS NDVI/NIR_v and (b) MODIS NDVI/NIR_v would be expected to outperform
893 TROPOMI SIF at estimating GPP seasonal dynamics. (a) Purple pixels are grass-closed regions
894 with significant tree covers (30% to 50%), green pixels are dominated by evergreen trees, and
895 grey pixels are dominated by sparse shrubs. (b) The low-productivity regions (colored pixels) are
896 defined as average GPP $< \sim 1 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ (or averaged NIR_v $<$ 0.04,) where the ratio
897 between bare ground and vegetation coverage is greater than 0.2.

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