

Vadose Zone Model for SALDS for Composite Analysis

Prepared for the U.S. Department of Energy
Assistant Secretary for Environmental Management

Contractor for the U.S. Department of Energy
under Contract DE-AC06-08RL14788

CH2MHILL
Plateau Remediation Company

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Vadose Zone Model for SALDS for Composite Analysis

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Terms

CA	composite analysis
CCUc	Cold Creek unit caliche
CCUsilt	Cold Creek unit upper silt and sand
ECF	environmental calculation file
EHM	equivalent homogeneous media
ETF	Effluent Treatment Facility
eSTOMP	exascale Subsurface Transport Over Multiple Phases
GIS	geospatial information system
Hf2	Hanford formation unit 2
HSU	hydrostratigraphic unit
ICF	Integrated Computational Framework
K_d	partition coefficient
P2R	plateau to river
PA	performance assessment
PA-TCT	power-averaging tensorial connectivity-tortuosity
RET	recharge evolution tool
RTD	remove, treat, and dispose
Rtf	Ringold Formation member of Taylor Flat
Rwie	Ringold Formation member of Wooded Island – unit E
SALDS	State-Approved Land Disposal Site
SIM-v2	Hanford Soil Inventory Model
STOMP	subsurface transport over multiple phases
TCT	tensorial connectivity-tortuosity
WMA C	Waste Management Area C

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1 Purpose

The objectives of the vadose modeling for the updated Hanford Site composite analysis (CA) are to simulate the flow and transport of water and radionuclide releases from the surface to the water table and to provide radionuclide transfer rates for the plateau to river (P2R) model, version 8.3 (CP-57037, *Model Package Report: Plateau to River Groundwater Model, Version 8.3*). Water additions include natural recharge and water discharged to the ground as a result of industrial processes associated with Hanford Site operations. Contaminant sources include radionuclides in water discharged to the ground during operations and radionuclides disposed “dry” in solid waste burial grounds or other means. The following 16 radionuclides were selected for this modeling effort: carbon-14 (C-14), chlorine-36 (Cl-36), tritium (H-3), iodine-129 (I-129), neptunium-237 (Np-237), rhenium-187 (Re-187), strontium-90 (Sr-90), technetium-99 (Tc-99), uranium-232 (U-232), uranium-233 (U-233), uranium-234 (U-234), uranium-235 (U-235), uranium-236 (U-236), uranium-238 (U-238), radium-226 (Ra-226), and thorium-230 (Th-230). The simulation time starts in 1943 and ends at 12070, which is 10,000 years after assumed Hanford Site closure in 2070.

The parallel version of the Subsurface Transport Over Multiple Phases (STOMP¹) simulator, officially named the exascale Subsurface Transport Over Multiple Phases (eSTOMP) is used to simulate flow and transport for the vadose models. The documentation for the STOMP code is comprehensive. The theoretical and numerical approaches applied in the STOMP code are documented in a published theory guide (PNNL-12030, *STOMP Subsurface Transport Over Multiple Phases Version 2.0 Theory Guide*). The code has undergone a rigorous verification procedure against analytical solutions, laboratory-scale experiments, and field-scale demonstrations. The application guide (PNNL-11216, *STOMP Subsurface Transport Over Multiple Phases Application Guide*) provides instructive examples in the application of the code to classical groundwater problems. The user’s guide (PNNL-15782, *STOMP: Subsurface Transport Over Multiple Phases Version 4.0: User’s Guide*) describes the general use, input file formatting, compilation, and execution of the code. The primary output of the vadose zone modeling is radionuclide transfer rates to the groundwater for input into the P2R model. The rates will be summed over the 100 by 100 m P2R grid cells that fall within the vadose zone model source domain.

The Hanford Site Central Plateau was subdivided into 26 individual vadose zone models, with 13 in the 200 East Area and 13 in the 200 West Area. Waste sites that have a completed performance assessment (PA) or past-leak analysis were not included as sources of radionuclides. Instead the vadose zone to groundwater transfer rates of the Environmental Restoration Disposal Facility, Integrated Disposal Facility, US Ecology, and Waste Management Area C (WMA C) PAs and the past-leak analysis for WMA C were used as direct input to the P2R model. Each of the vadose zone models is documented in separate environmental calculation files (ECFs). This ECF describes the State-Approved Land Disposal Site (SALDS) model. The scope of this ECF is to document the development and results of the SALDS vadose zone model. CP-63515, *Model Package Report: Central Plateau Vadose Zone Models*, describes the approach, assumptions, process of determining the number of models required and domain of each model, input data, and processing common to all the models. Additionally, the following documents support inputs to the models:

- CP-60925, *Model Package Report: Central Plateau Vadose Zone Geoframework*, describes the hydrostratigraphic framework.

¹ STOMP is a copyright of Battelle Memorial Institute, Columbus, Ohio, and used under the Limited Government License.

- CP-61786, *Inventory Data Package for the Hanford Site Composite Analysis*, contains the solid waste inventory.
- CP-62184, *Hanford Site Composite Analysis: Radionuclide Selection for Groundwater Pathway Evaluation*, describes the selection of the 16 radionuclides used in these simulations.
- CP-62766, *Model Package Report: Composite Analysis Solid Waste Release Model (CASWR Model)*, describes the mechanisms of release of radionuclides from solid waste based on waste type.
- CP-63883, *Vadose Zone Flow and Transport Parameters Data Package for the Hanford Site Composite Analysis*, describes the process of assigning material properties to the hydrostratigraphic units (HSUs).
- ECF-HANFORD-15-0019, *Hanford Site-wide Natural Recharge Boundary Condition for Groundwater Models*, describes the recharge evolution tool (RET) used to calculate the recharge.
- ECF-HANFORD-17-0079, *Hanford Soil Inventory Model (SIM-v2) Calculated Radionuclide Inventory of Direct Liquid Discharges to Soil in the Hanford Site's 200 Areas*, describes the aqueous sources for the CA modeling effort, which uses the source inventory found in Appendix F of ECF-HANFORD-17-0079. The SALDS model only has one waste site, SALDS, or 600-211, and the inventory for that waste site is not in ECF-HANFORD-17-0079. The inventory for this model is obtained from EMDT-IN-0047², *SALDS Liquid Disposal Volumes and Tritium Inventory*. The cover sheet for EMDT-IN-0047 is shown in Appendix F of this ECF.
- ECF-HANFORD-18-0035, *Central Plateau Vadose Zone Geoframework*, describes the updates to the hydrostratigraphy surfaces defined in CP-60925, and defines the hydrostratigraphy surfaces used by this modeling effort.
- ECF-HANFORD-19-0032, *Distribution of Infiltration in the 216-U-10 and 216-B-3 Pond Systems 1944-1997*, estimates the routing of effluent and infiltration between ditches and ponds of the 216-U-10 Pond System and between the main pond and expansion lobes of the 216-B-3 Pond System.
- ECF-HANFORD-19-0094, *Calculation of Moisture-Dependent, Anisotropic Parameters Supporting the Hanford Site's Composite Analysis, Cumulative Impact Evaluation, and Performance Assessment*, describes calculations of moisture-dependent, anisotropy of hydraulic conductivity for the HSUs.
- ECF-HANFORD-19-0112, *Solid Waste Release Calculations for the Composite Analysis Baseline Assessments*, calculates the solid waste annual release rates.
- ECF-HANFORD-19-0121, *Selection of Vadose Zone Flow and Transport Properties with Gravel Fraction Corrections for the Hanford Site Composite Analysis and Cumulative Impact Evaluation*, describes the physical and chemical properties used for these models.
- ECF-HANFORD-20-0006, *Composite Analysis Solid Waste Release Data Reduction of Activity Flux from Waste Sites to the Vadose Zone*, describes the solid waste data reduction.

² EMDT-IN-0047, *SALDS Liquid Disposal Volumes and Tritium Inventory*, Rev. 0, CH2M HILL Plateau Remediation Company, Richland, Washington. Electronic model data transmittals are stored in the Environmental Model Management Archive. A copy of the cover sheet for this EMDT is provided in Appendix F.

2 Background

The SALDS model simulates releases from SALDS 600-211 north of the 200 West Area (Figure 2-1). The SALDS facility is the only disposal site in the model area. As part of the Hanford Site cleanup mission, unpermitted discharges of waste liquids to the soil column ended in 1995. To replace this practice, the Effluent Treatment Facility (ETF) was constructed in the 200 East Area to remove contaminants from wastewater streams before discharge to the environment (WHC-SD-EN-ES-036, *Site Evaluation Report, C-018H Disposal Siting Evaluation*). However, the discharged water contains H-3 because there is no cost-effective treatment technology to remove H-3 from water (DOE/RL-2014-10, *Evaluation of Tritium Removal and Mitigation Technologies for Wastewater Treatment*). Tritiated water from ETF is discharged to the environment at the SALDS facility, which is permitted by the State of Washington³. The SALDS is located north of the 200 West Area where the travel time to the Columbia River is long enough that H-3 concentrations will be reduced to safe levels by radiological decay before the plume reaches the river (WHC-SD-EN-ES-036).

SALDS is a 35 by 61 m rectangular drain field that has been in use since 1995 (RPP-RPT-61178, *Results of Tritium Tracking and Groundwater Monitoring at the Hanford Site 200 Area State Approved Land Disposal Site, Fiscal Year 2018*). Through 2017, it has received 1.2 million m³ of effluent containing 437 Ci of H-3 (RPP-CALC-61950, *Fate and Transport Analyses of Historical and Future Tritium Releases from the State Approved Land Disposal Site, FY 2018*). It is projected that SALDS will continue in use until 2065 (RPP-CALC-61876, *Estimated Tritium Discharges to the State Approved Land Disposal Site for Use in Groundwater Modeling*).

³ The U.S. Department of Energy (DOE) has taken the position that its groundwater monitoring and provision of data reporting to the State of Washington Department of Ecology (Ecology) is a matter of intergovernmental comity and cooperation, and that the Permit has no jurisdiction over radionuclides, which are regulated by DOE under *Atomic Energy Act of 1954* (AEA) authority, in the same way that permits for wastewater discharge to surface waters issued by the U.S. Environmental Protection Agency under Section 402 of the *Clean Water Act of 1977* are preempted by the AEA from regulating radionuclides. DOE shares its monitoring data with Ecology consistent with this policy of cooperation.

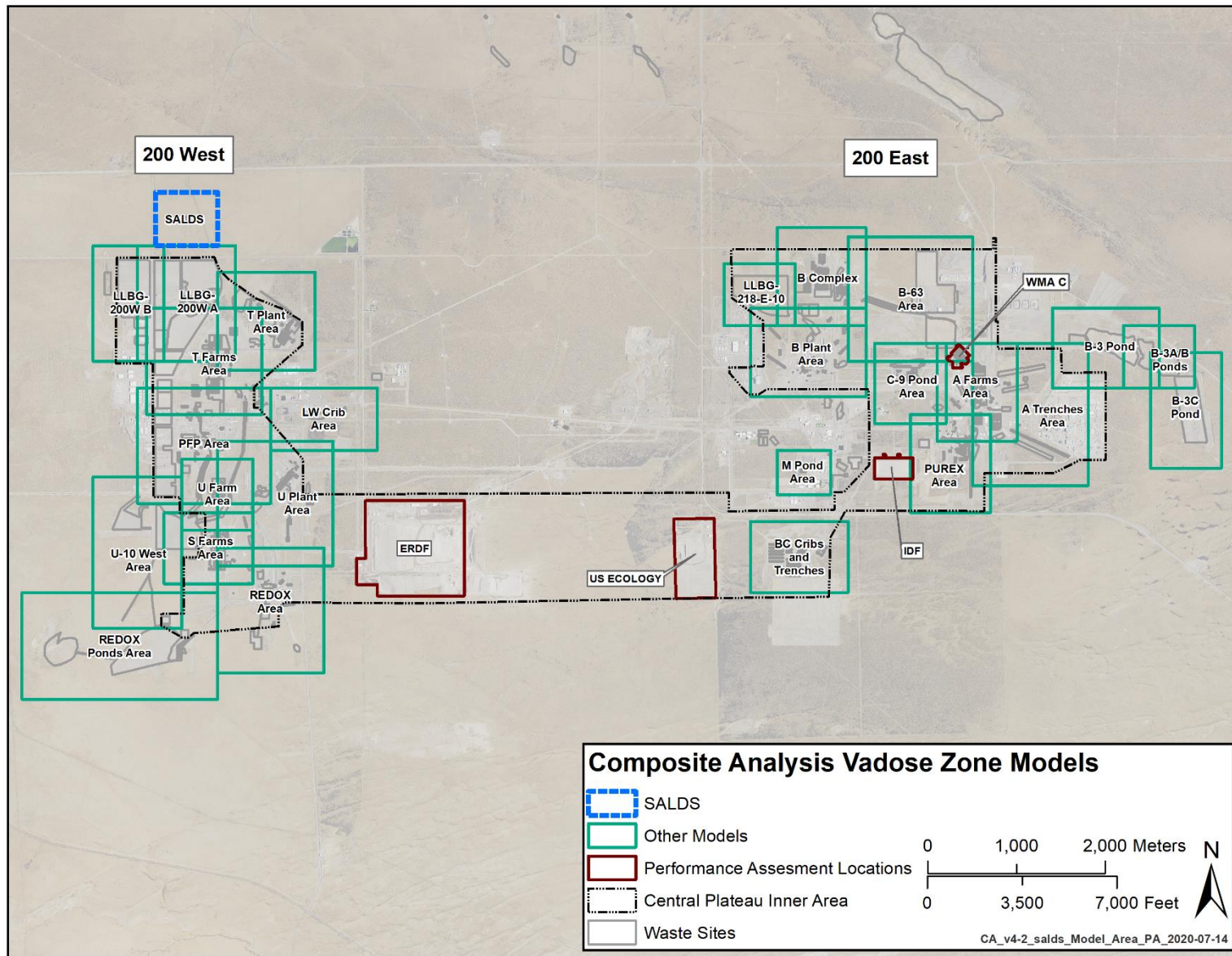


Figure 2-1. Location of the SALDS Model

3 Methodology

This chapter contains a discussion of configuration control, a brief overview of the methodology for creating the SALDS model, and a list of modifications specific to this model.

3.1 Configuration Control

A configuration control system was developed so that all vadose zone models generated for the CA would follow a consistent set of conventions and use only approved input data (e.g., geoframework, hydraulic and contaminant properties, source releases, etc.). This system was manifested as sets of qualified input data, scripts used to construct the models and post-process the results and sets of instructions for building and executing the models. Each script was reviewed, tested, and documented to qualify it for use. A list of scripts developed for the vadose zone modeling effort is found in Section 5.3 of this ECF. Each CA model used the same directory structure. A discussion of the configuration control system is found in CP-63515.

A data configuration quality-control system (hereinafter called the Integrated Computational Framework [ICF]), provides the tools necessary to verify that all model output data are correctly associated with their corresponding input data. The ICF consists of two parts: a file management system and utility scripts to support the file management system.

The ICF houses all data produced by and in support of the CA modeling effort. The ICF file management system ensures that no data can be modified, deleted, or used in a model application without being checked into the ICF, reviewed, and accepted by the ICF administrator. Separating the data flow from the modeling helps prevent accidental modification and guarantees a data review prior to acceptance of any data product into the ICF.

The utility scripts establish a pedigree for any data product stored in the ICF. The ICF allows users to ascertain all the ancestor and derivative products related to any ICF data product. By combining the file structure and software utilities, the ICF provides confidence that the CA output data are associated with a set of versioned input data.

The CA models were constructed on a central computer system, and many of the models contained over one million nodes. Along with the long time period simulated and the release of large volumes of water from liquid waste disposal sites in many of the model domains, the size of the models caused long run times. Thus, the model files were transferred to a high-power computer system, GAIA, for execution. Following completion of model runs, the input and output files were returned to the original computer system for post-processing. File fingerprinting was used to verify this transfer process and to verify that the correct input files were used for each model simulation.

3.2 Model Construction and Execution

This ECF is one of 26 similar ECFs, one for each CA vadose zone model, each of which followed the same general methodology. A detailed description of the general model construction is found in CP-63515. Adjustments are made to the methodology as needed to tailor model development to best represent the area being simulated. The steps were developed to include mass balance checks to verify model performance. A brief outline for the construction and execution of the SALDS model is as follows:

1. Construct the model grid.
2. Assign HSUs and material properties to the model grid nodes.

3. Generate the temporal-spatial recharge distributions for the model using the RET.
4. Execute the steady-state flow simulation to establish the initial conditions for the transient simulations.
5. Conduct post-processing of the steady-state simulation, including calculating the liquid volume balance.
6. Incorporate the transient RET results, radionuclide waste release, and liquid waste release data into the model input file. Generate input files for a historical simulation from 1943–2018, a forecast simulation from 2018–12070, and a simulation from 1943–12070 with no radionuclide decay which is used to check the mass balance.
7. Execute the mass balance simulation. This requires two simulations because the 16 radionuclides simulated are divided into two groups, Radionuclide Group 1 and Radionuclide Group 2, as shown in Table 3-1.
8. Conduct post-processing of the radionuclide mass balance simulations, including calculating the mass balance.
9. Execute the historical radionuclide transport simulations (1943–2018) for Radionuclide Group 1. There is no inventory for Radionuclide Group 2 for this model, so it is not simulated.
10. Execute the forecast radionuclide transport simulations from 2018–12070 for Radionuclide Group 1.
11. Conduct post-processing of the radionuclide transport simulations to generate contaminant transfer rates to groundwater for the P2R model.

**Table 3-1. List of Modeled Radionuclides in
Radionuclide Group 1 and Radionuclide Group 2**

Radionuclide Group 1	Radionuclide Group 2
C-14	U-232
Cl-36	U-233
H-3	U-234
I-129	U-235
Np-237	U-236
Re-187	U-238
Sr-90	Ra-226
Tc-99	Th-230

All model inputs were checked during production. Checking documentation is shown in Appendix A.

3.3 Model-Specific Modifications

Model-specific changes were required for some models. This model required no model-specific modifications.

4 Assumptions and Inputs

The domain and structure of the SALDS model, hydraulic properties, boundary and initial conditions, source releases, the types of simulations performed, and assumptions are described in this chapter.

4.1 Model Domain and Grid

The SALDS model was constructed to simulate radionuclide contaminant transport through the vadose zone from the waste sites at and around SALDS in the 200 West Area. The extents and grid spacing of this model are shown in Figure 4-1. A general approach to grid spacing for the CA vadose zone models, both horizontal and vertical, is discussed in CP-63515. The SALDS model grid is aligned with the P2R model grid (CP-57037) as shown in Figure 4-2. The SALDS model has 70 columns from west to east (X-nodes), 60 rows from south to north (Y-nodes), and 154 layers in the vertical dimension (Z-nodes), for a total of 646,800 nodes. The total extent of the model is 700 m in the east-west direction and 600 m in the north-south direction. The southwest corner of the domain has coordinates of 566,100 m east and 137,800 m north (Washington State Plane, South Zone [4602]). The model extends vertically from the approximate water table elevation to the ground surface. Grid spacing for each model was determined through multiple iterations based on geologic layer thickness, plume extent, waste site alignment, and mass balance considerations. Preliminary model runs were used to evaluate spatial discretization, and refinements were made as necessary (e.g., to better represent source zone geometry and plume migration). Vertical spacing is 0.5 m.

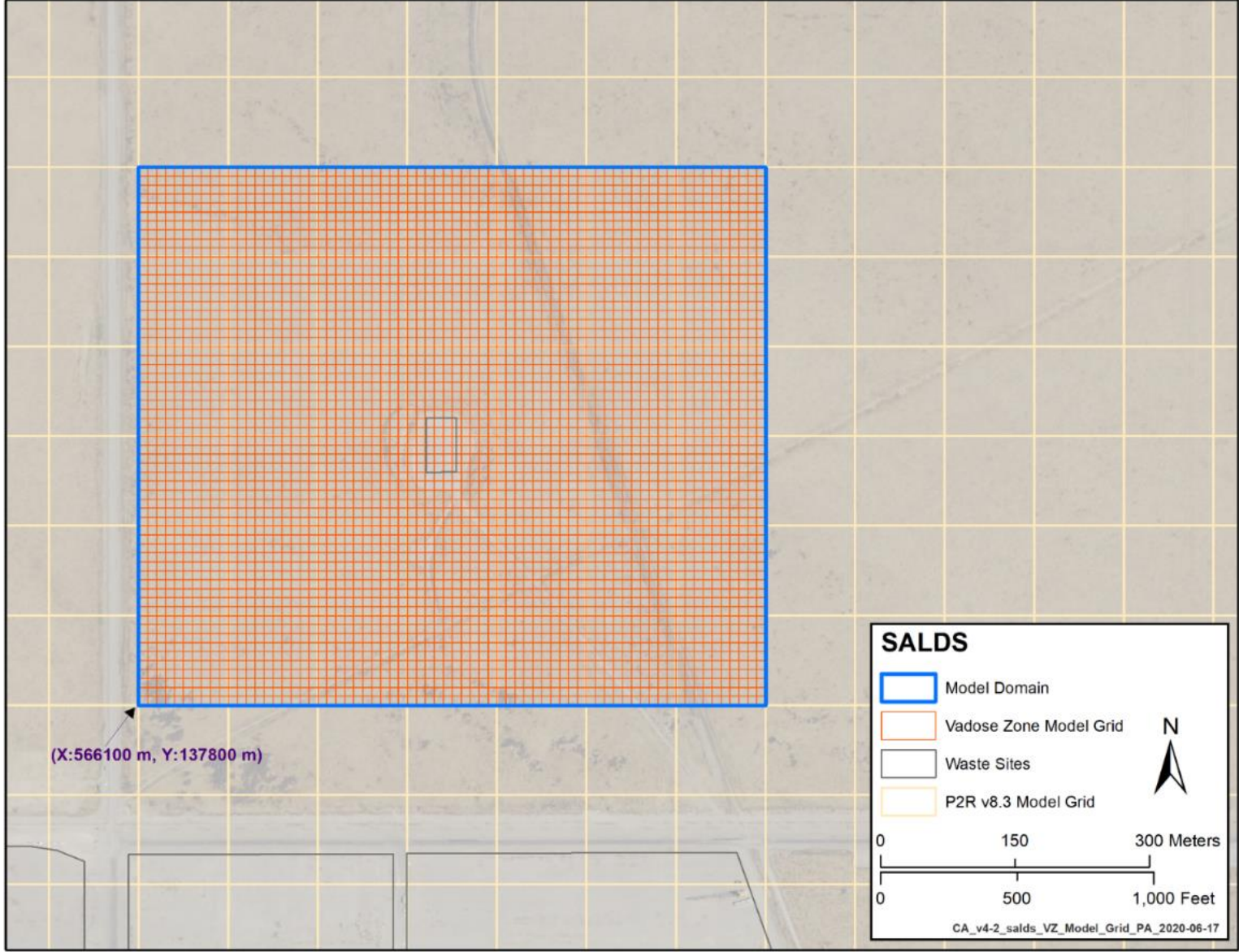


Figure 4-1. Plan View of the SALDS Model Grid Overlain on the P2R Grid Cells



Figure 4-2. Plan View of the P2R Grid Cells in the SALDS Model

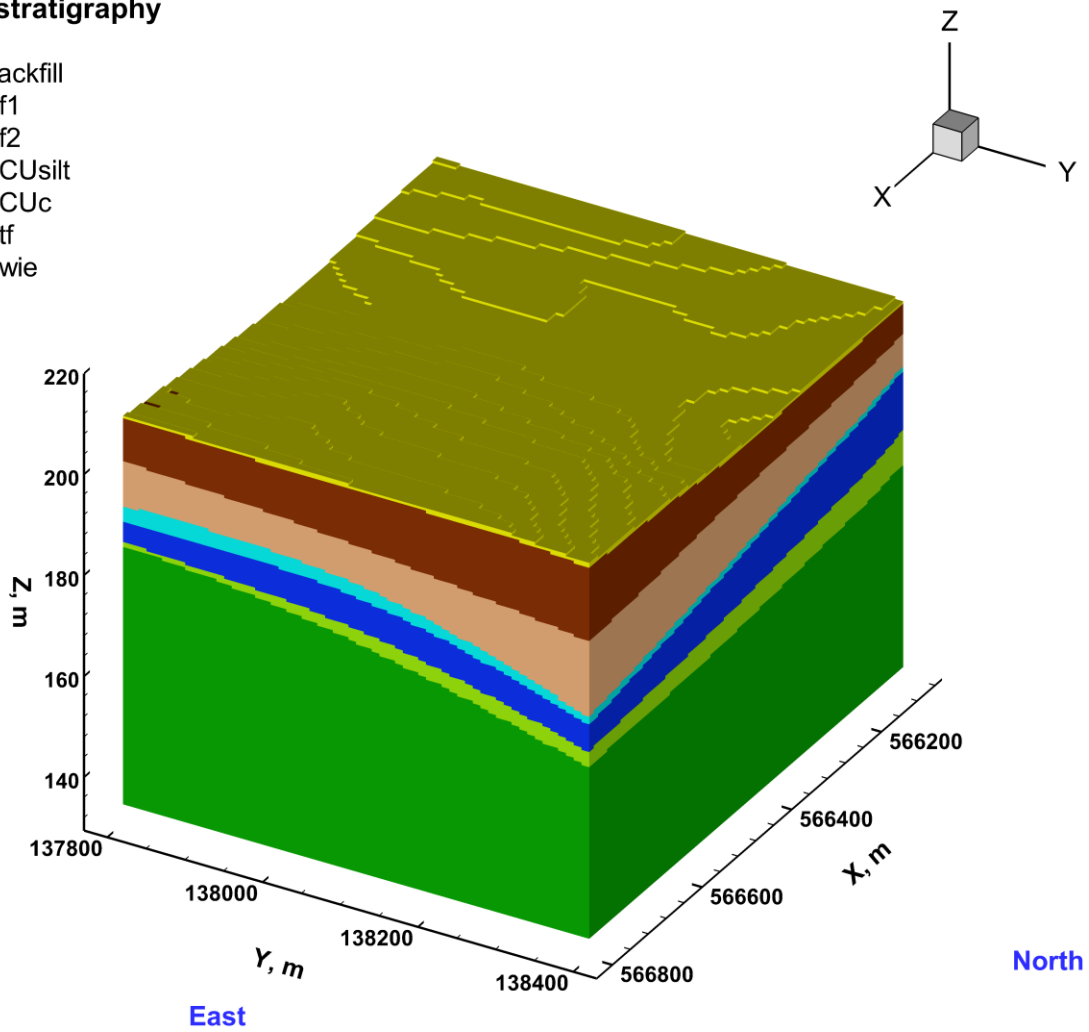
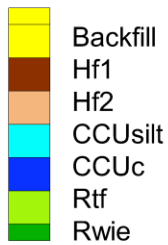
This model has a source zone but no buffer zone. These regions are distinguished by how the radionuclide inventory from waste sites is distributed. Water and radionuclide releases were simulated for waste sites in a source zone, whereas only water volume releases would have been simulated for waste sites in a buffer zone. If a buffer zone were to be present, water volume releases in the buffer zone would have been included so that their hydraulic effect on flow beneath the source zone would be accounted for. For models that have buffer zones, waste sites with radionuclide releases located in a buffer zone are included in the source zones of other models.

4.2 Model Hydrostratigraphy

The SALDS model includes seven HSUs: Backfill, Hanford formation unit 1 (Hf1), Hanford formation unit 2 (Hf2), Cold Creek unit upper silt and sand (CCUsilt), Cold Creek unit caliche (CCUc), Ringold Formation Member of Taylor Flat (Rtf), and Ringold Formation Member of Wooded Island – unit E (Rwie), in descending sequence. HSU designations were assigned to each grid node based on the surfaces in the geoframework model (ECF-HANFORD-18-0035). Properties assigned to each HSU are presented in ECF-HANFORD-19-0121 and are described in Section 4.3. For a detailed description of the hydrostratigraphy for the CA vadose zone models see CP-63515. Figures 4-3 through 4-6 show the hydrostratigraphic framework for the SALDS model from various orientations. A progression of cross-sections from west to east and south to north through the model are shown in Appendix B of this ECF.

Rwie is the thickest and oldest layer. The Hf1, Hf2, and Rtf thin towards southwest and Rtf thins towards the southeast. The CCUsilt is thinnest layer in the model. The CCUc thickens to the north. The Rwie is the thickest unit in the model and slopes to the northeast.

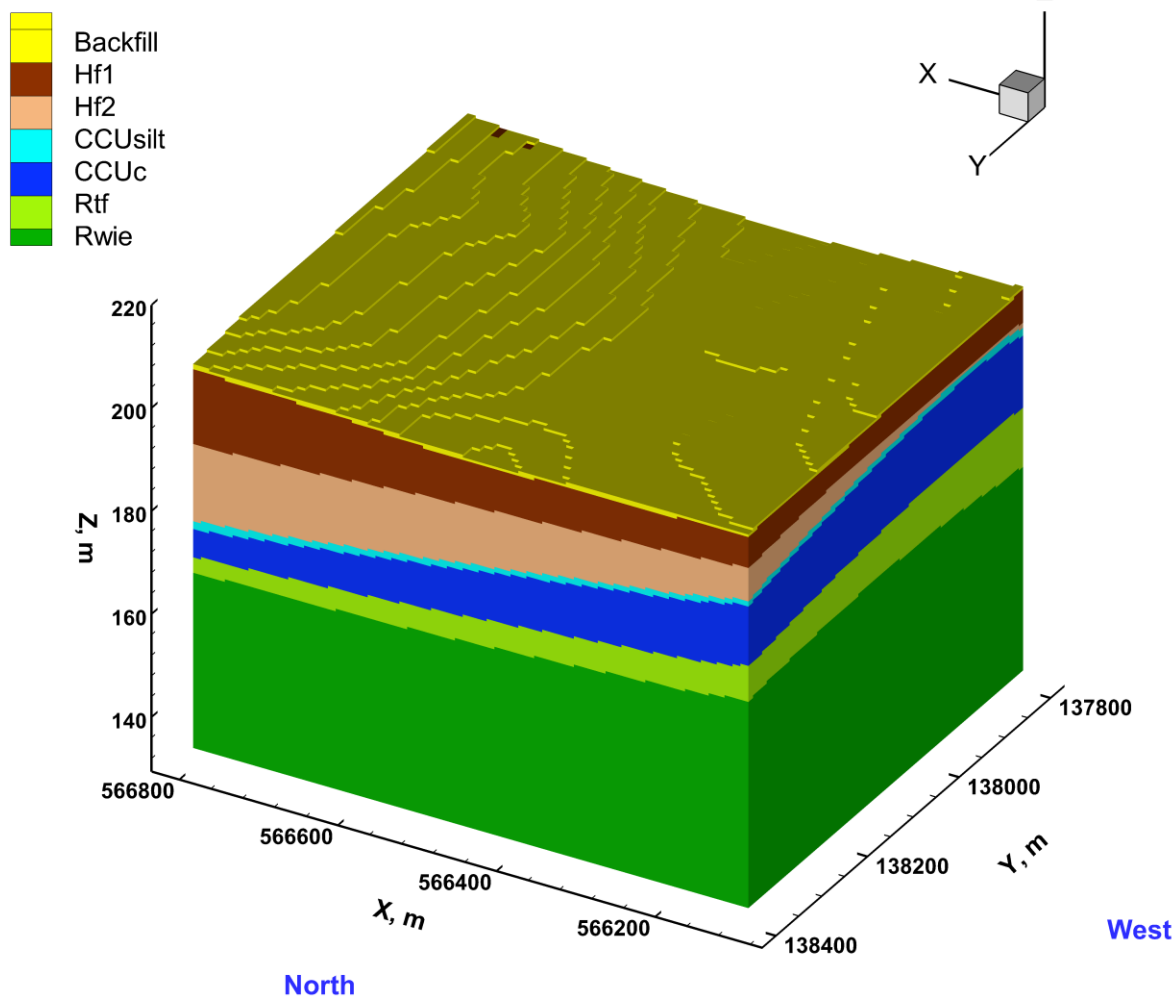
Hydrostratigraphy



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Figure 4-3. Model Hydrostratigraphy Three-Dimensional View Showing the North and East Faces

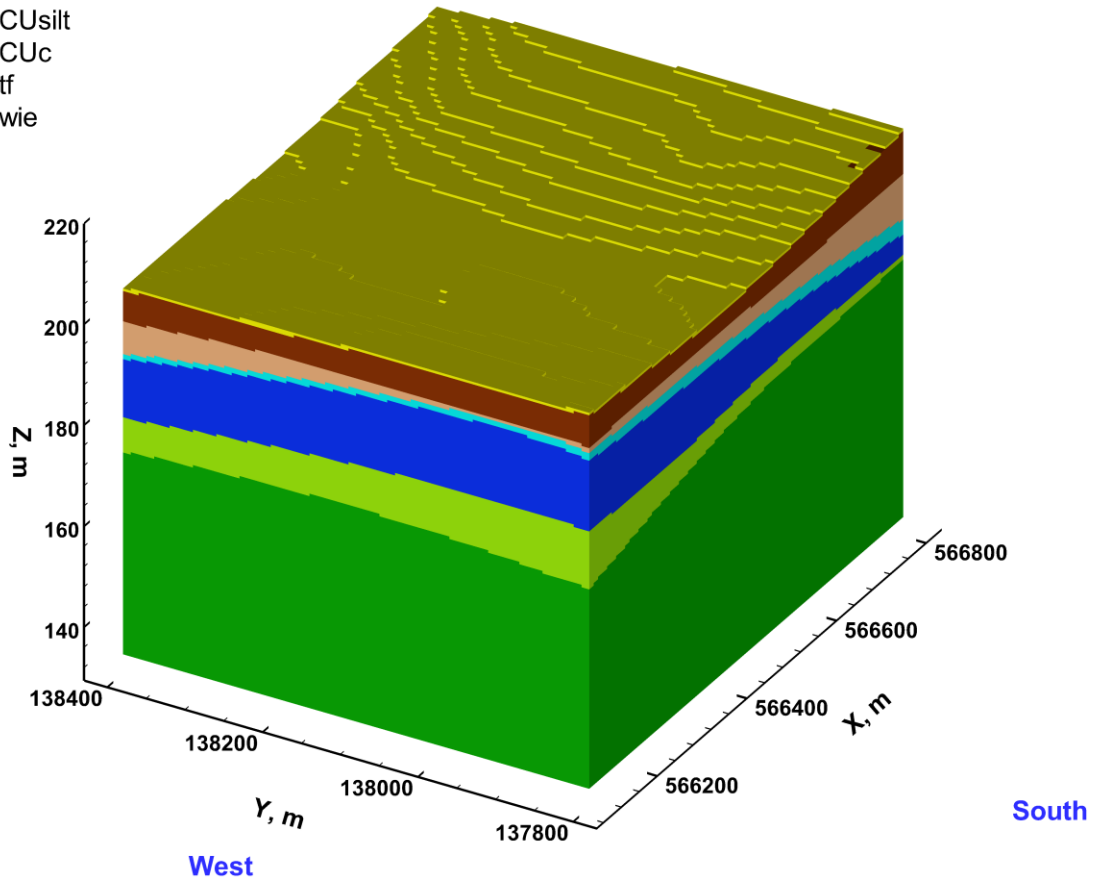
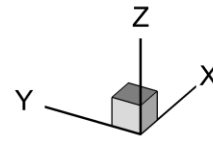
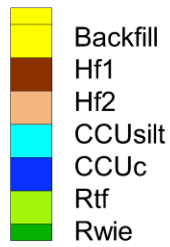
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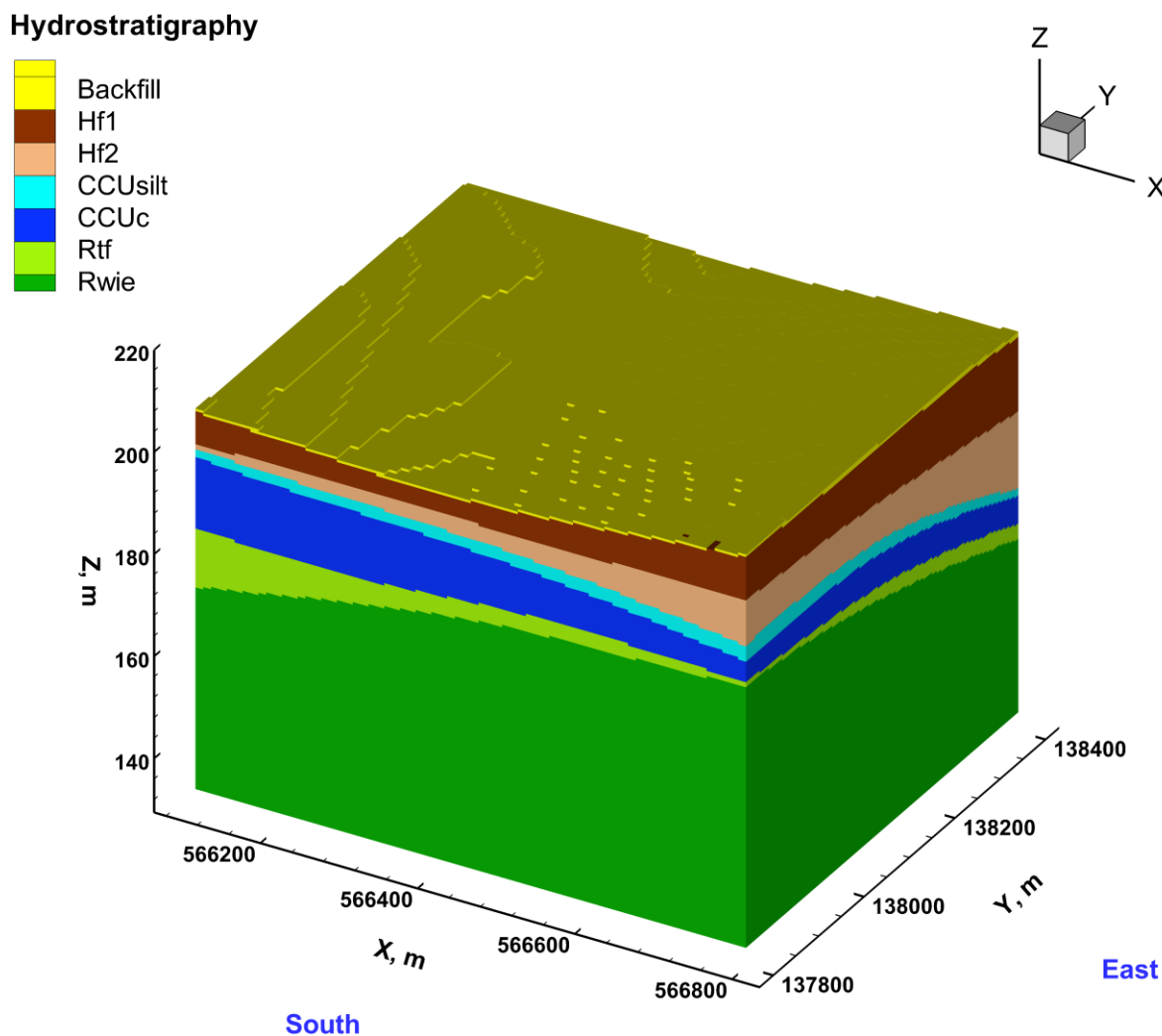
Figure 4-4. Model Hydrostratigraphy Three-Dimensional View Showing the North and West Faces

Hydrostratigraphy



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Figure 4-5. Model Hydrostratigraphy Three-Dimensional View Showing the South and West Faces



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Figure 4-6. Model Hydrostratigraphy Three-Dimensional View Showing the South and East Faces

4.3 Hydraulic Properties

Hydraulic properties for the SALDS HSUs are shown in Tables 3, 4, 6, and 7 of ECF-HANFORD-19-0121. For most of the HSUs, hydraulic property estimates in ECF-HANFORD-19-0121 were obtained from CP-63883, which contains a detailed description of the development of these parameters for the unconsolidated sediments overlying the basalt HSU in the Central Plateau. Properties for the perched zone units and the basalt HSU were obtained from other sources.

HSUs were assumed to follow the van Genuchten (van Genuchten, 1980, “A Closed-form Equation for Predicting the Hydraulic Conductivity of Unsaturated Soils”) moisture-retention constitutive relation and the Mualem-van Genuchten relative-permeability constitutive relation (Mualem, 1976, “A New Model for Predicting the Hydraulic Conductivity of Unsaturated Porous Media”), requiring values to be specified in STOMP for the following items:

- Saturated hydraulic conductivity
- Saturated moisture content
- Residual saturation, equal to the residual moisture content divided by the saturated moisture content
- van Genuchten α , proportional to the inverse of the air entry matric potential
- The dimensionless van Genuchten n fitting parameter
- The tensorial connectivity-tortuosity (TCT) parameters for moisture dependent anisotropy (discussion of the TCT parameters is in CP-63515 and ECF-HANFORD-19-0094)

4.4 Transport Parameters

In addition to the hydraulic properties discussed in Section 4.3, the transport simulations also require particle density, molecular diffusion rate, longitudinal and transverse dispersivity, solid-aqueous partition coefficient (K_d), and radionuclide half-life. Tables 5, 8, 9, 10, 13, 15, and 16 of ECF-HANFORD-19-0121 list the transport properties for the HSUs present in the modeled area. A detailed description of the transport properties used for the CA vadose zone models can be found in ECF-HANFORD-19-0121.

4.5 Source Releases

Within the source zone, the transport models consider radionuclide releases from both solid and liquid sources. Some sites within a model's source zone lack a radionuclide inventory and are also simulated as water-only releases (e.g., septic systems). This model contains no sites with solid releases. An index of waste sites contributing releases to the model are shown in Table 4-1. The waste site contributing liquid releases within this model is shown in Figure 4-7. Section 4.5.1 contains a discussion of the radionuclide inventory released from the waste site in the model; liquid waste sites are addressed in Section 4.5.1.1, and solid waste sites are addressed in Section 4.5.1.2. Section 4.5.2 addresses liquid (volume) releases from the waste site.

Table 4-1. Waste Sites Included in the SALDS Model

Source Zone – Liquid Waste Sites with Radionuclide Releases (1)
600-211
Source Zone – Liquid Waste Sites with No Radionuclide Releases (i.e., Liquid Only) (0)
None
Source Zone – Solid Waste Sites (0)
None
Buffer Zone – Waste Sites (Liquid Only) (0)
Not applicable

SALDS = State-Approved Land Disposal Site

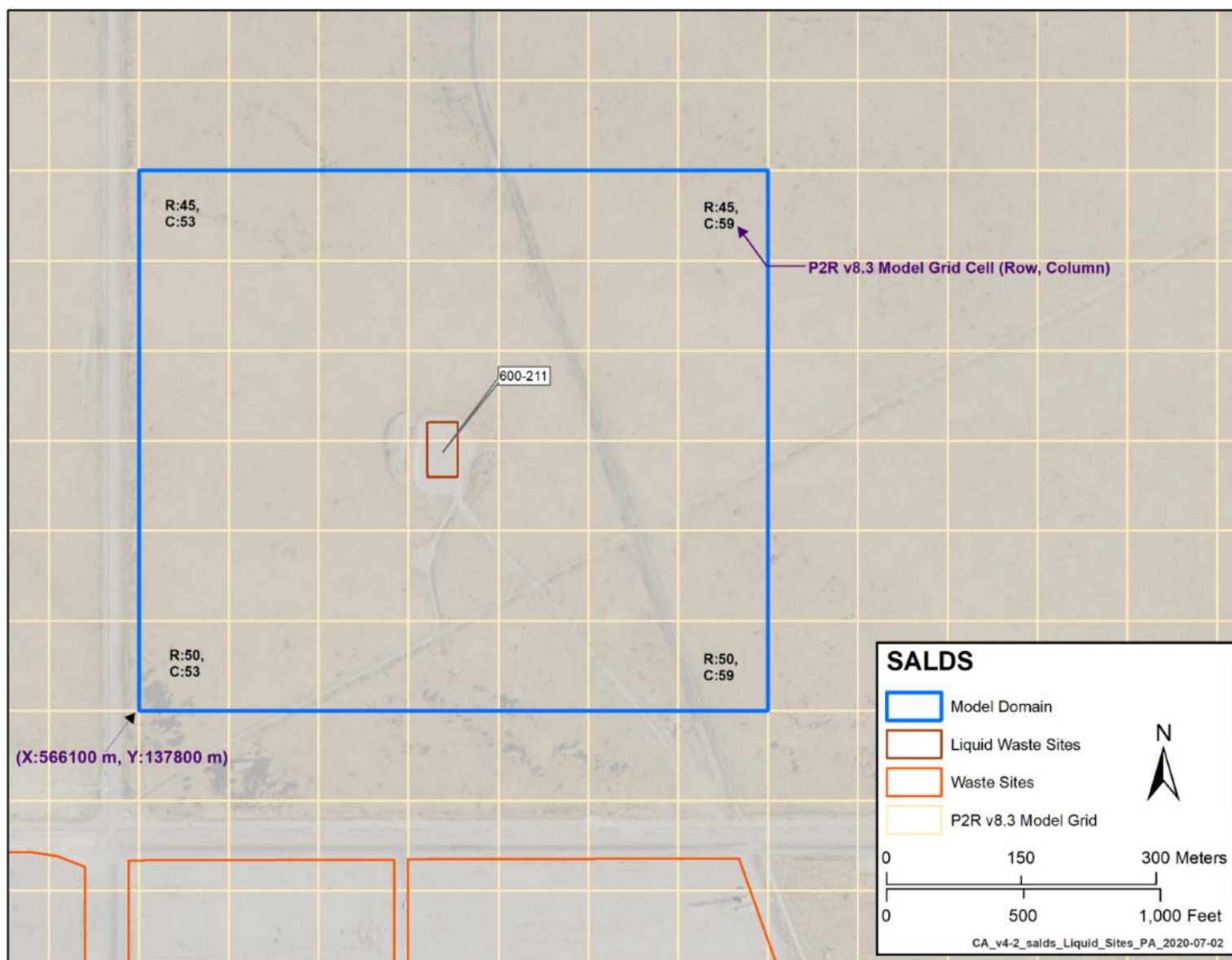


Figure 4-7. Waste Sites in the SALDS Model with Liquid Source Inventory

The radionuclides included in the CA vadose zone models were determined through a screening process based on prior modeling studies. CP-62184 discusses this screening process. This process identified 16 radionuclides for simulation. For computational reasons, transport of radionuclides for the CA vadose zone modeling effort are modeled in two separate groups, Radionuclide Group 1 and Radionuclide Group 2, as shown in Table 3-1. Transport properties and half-lives of the radionuclides are described in CP-62184. Not all 16 radionuclides are present in every model. No inventory is present at the waste sites in this model domain except for H-3; therefore, no other radionuclides were simulated and Radionuclide Group 2 was not modeled at all. Radionuclide activities released in the model (from liquid and solid waste sites separately, as well as the total) are shown in Table 4-2.

Table 4-2. Released Radionuclide Activities in the SALDS Model

Radionuclide	Total (Ci)	Liquid Waste (Ci)	Solid Waste (Ci)
Radionuclide Group 1			
C-14	0.000E+00	0.000E+00	0.000E+00
Cl-36	0.000E+00	0.000E+00	0.000E+00
H-3	1.097E+03	1.097E+03	0.000E+00
I-129	0.000E+00	0.000E+00	0.000E+00
Np-237	0.000E+00	0.000E+00	0.000E+00
Re-187	0.000E+00	0.000E+00	0.000E+00
Sr-90	0.000E+00	0.000E+00	0.000E+00
Tc-99	0.000E+00	0.000E+00	0.000E+00
Radionuclide Group 2			
U-232	0.000E+00	0.000E+00	0.000E+00
U-233	0.000E+00	0.000E+00	0.000E+00
U-234	0.000E+00	0.000E+00	0.000E+00
U-235	0.000E+00	0.000E+00	0.000E+00
U-236	0.000E+00	0.000E+00	0.000E+00
U-238	0.000E+00	0.000E+00	0.000E+00
Th-230	0.000E+00	0.000E+00	0.000E+00
Ra-226	0.000E+00	0.000E+00	0.000E+00

SALDS = State-Approved Land Disposal Site

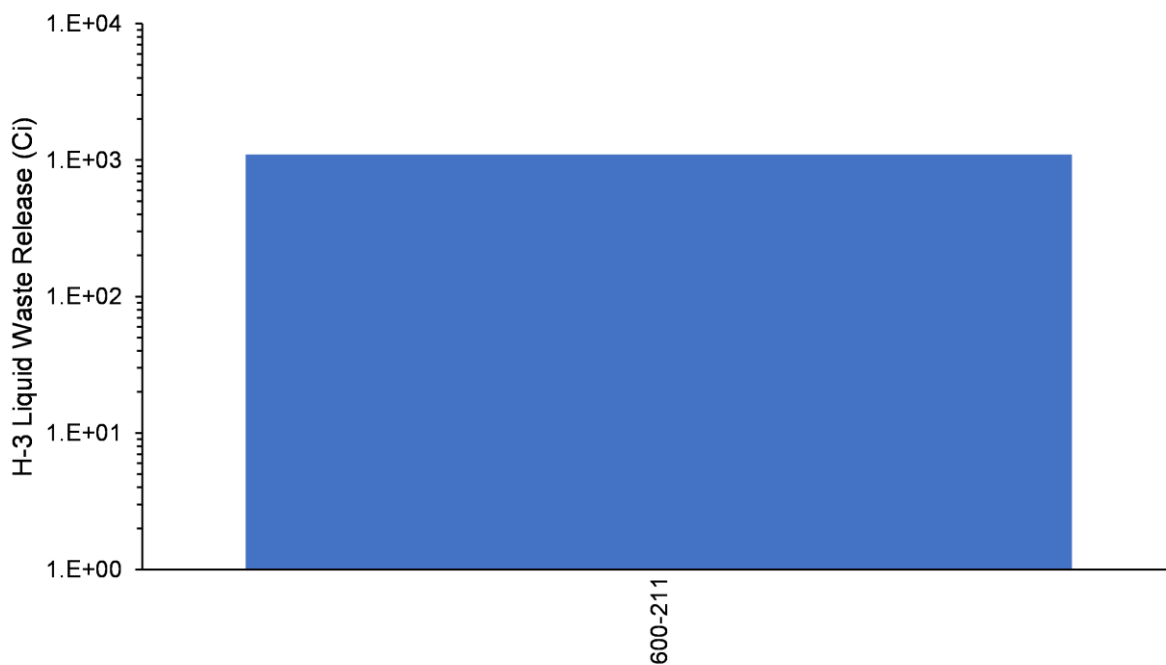
4.5.1 Contaminant (Activity) Releases

This section describes the releases of radionuclides to the subsurface included in this model. Simulations for the CA consider both liquid and solid waste sites, but only liquid waste releases are present in the

source zone of this model. These are described in Section 4.5.1.1. Releases were input to the model as annual average release rates.

4.5.1.1 Liquid Waste Site Releases

Liquid waste sites are sites where liquid wastes, often containing radionuclides, are released to the vadose zone. A map of the aqueous waste site in the SALDS model is shown in Figure 4-7. The waste site inventory was retrieved from EMDT-IN-0047. The H-3 discharged to this model from the liquid waste site is shown as a site total in Figure 4-8, and by year in Figure 4-9.



CA_v4-2_salids_H-3_liquid_release_log_pa_2020-06-30

Figure 4-8. Total H-3 Activity Released from Liquid Waste Sites in the SALDS Model

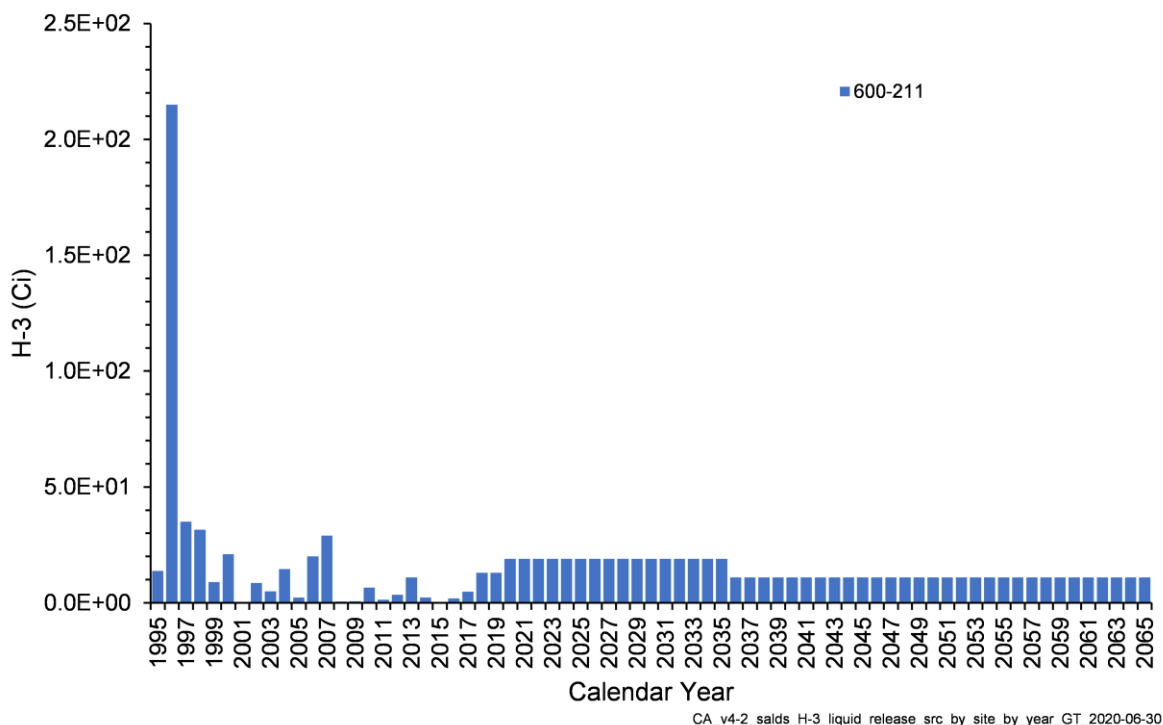


Figure 4-9. Annual H-3 Activity Released from Liquid Waste Sites in the SALDS Model

4.5.1.2 Solid Waste Site Releases

Solid wastes are contaminated materials that have the potential to release radionuclides to the vadose zone. There are no solid waste sites present in the source zone of this model.

4.5.2 Liquid (Volume) Releases

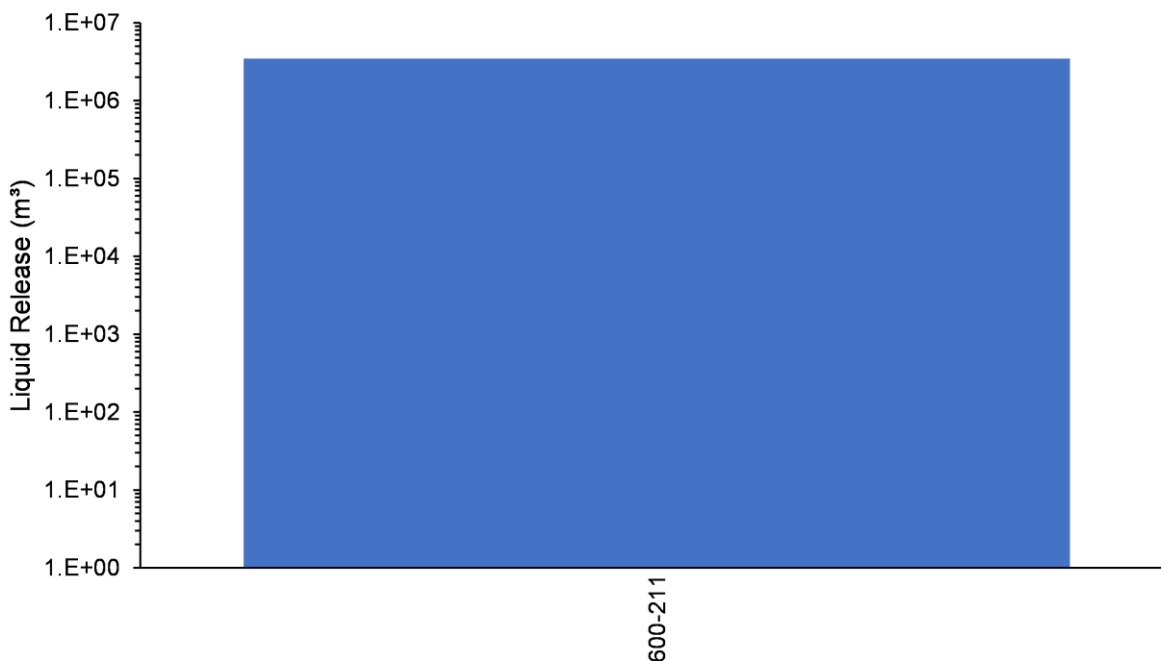
This section provides information on liquid volumes released within the domain of the SALDS model. These liquids can act as a driving force for the movement of radionuclides deeper into the subsurface. Table 4-3 shows an overview of the total liquids released in the model. Figure 4-10 shows the volume of water released within the model domain by the waste site, and Figure 4-11 shows the total volume of water released by year.

Table 4-3. Released Liquid Volumes in the SALDS Model

Total	Source Zone	Buffer Zone
3,470,760	3,470,760	0.000E+00

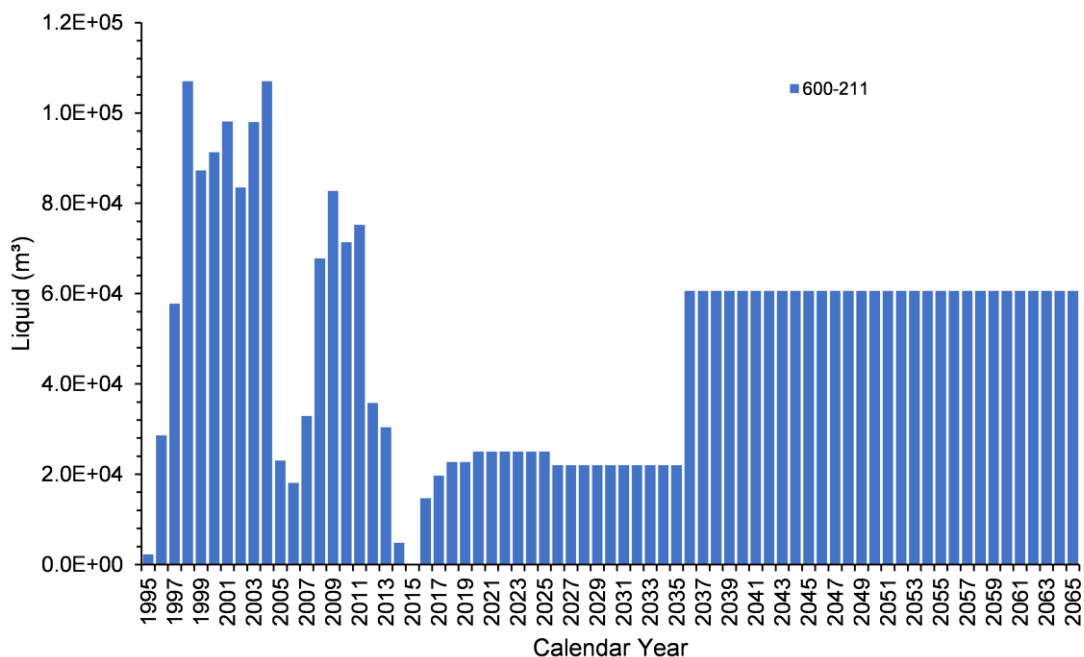
Note: All values reported in m³.

SALDS = State-Approved Land Disposal Site



CA_v4-2_salds_liquid_liquid_release_log_pa_2020-06-30

Figure 4-10. Total Volume of Water Released from Liquid Waste Sites in the SALDS Model



CA_v4-2_salds_Aqueous Volume_liquid_release_src_by_site_by_year_GT_2020-06-30

Figure 4-11. Total Volume of Water Released by Year from Liquid Waste Sites in the SALDS Model

4.5.2.1 Liquid Release Modifications

For some models, modifications to liquid release volumes were needed to help with convergence of the numerical solution or to provide for more representative transport through the vadose zone. However, no modifications to liquid releases were needed for the SALDS model.

4.6 Simulations

Three different types of simulations were performed. Constant recharge conditions were used in a flow-only simulation to set the initial aqueous pressure conditions in the model. A mass balance simulation was conducted to evaluate model performance, and transport simulations were performed to estimate radionuclide activity entering the saturated zone. These are discussed in the following sections.

4.6.1 Flow-Only (Steady-State) Simulation

The flow-only simulation was performed using recharge estimated for 1943, which was prior to the start of Hanford Site operations. This was a transient simulation, but it is referred to hereinafter as the steady-state simulation because recharge was held constant at the 1943 values and the simulation was run for 10,000 years to ensure steady-state conditions were achieved within the model domain. The results were used as the initial aqueous pressure conditions for the radionuclide transport simulations starting in 1943.

4.6.2 Mass/Activity Balance Simulation

A mass/activity balance simulation was conducted to evaluate model performance. This simulation was run for 10,000 years using the source releases described in Section 4.5 and the initial aqueous pressure conditions from the steady-state simulation, but radionuclide half-lives were set to $1.0\text{E}+20$ years to eliminate radiological decay and allow for the mass/activity balance to be evaluated directly. The mass/activity of each constituent leaving the model over 10,000 years and the mass/activity present in the model at the end of the simulation were summed, and the results were compared to the mass/activity released from the sources.

4.6.3 Transport Simulations

Transport simulations were performed to estimate the radionuclide activity entering the saturated zone. These were done in stages. The time period for the CA evaluation is 2018 to 12070. To set the initial radionuclide concentrations in the model domain for simulations of that time period (i.e., forecast period), a historical simulation of radionuclide releases was performed from 1943 up to but not including 2018. The radionuclide distribution in the model domain at the end of this simulation became the starting concentrations for the forecast runs.

The forecast simulations were performed for 2018 to 12070. The forecast simulation was performed in a single stage because this model contains no waste sites with a disposition of remove, treat, and dispose (RTD). If it had contained such sites, the forecast period would have been simulated in two stages. After starting in 2018, execution of the model would have been stopped at the year RTD was planned to reset concentrations in the model to zero at the RTD locations, and then the model would have been restarted from that year.

4.7 Initial Conditions

The simulations performed for the SALDS model require that initial aqueous pressure conditions and radionuclide concentrations in the model domain be specified, depending on the simulation. Initial aqueous pressure conditions for the steady-state, flow-only simulation are based on hydrostatic conditions

assuming that the base of the model is at the water table. This is input to STOMP as an aqueous pressure of 101,325 Pa at the water table and a z-direction gradient of -9,793.52 Pa/m.

For the historical transient simulations (i.e., 1943 to 2018), initial aqueous pressure conditions are the steady-state conditions taken from the end of the steady-state simulation. Since the purpose of the historical simulations was to define the starting radionuclide concentrations and aqueous pressure conditions for the forecast runs by simulating source release during the entirety of Hanford Site operations, the initial radionuclide concentrations were zero.

Aqueous pressure conditions and radionuclide concentration results of the historical simulation were used as the initial conditions for the forecast simulations. This model does not contain any RTD sites, so the forecast simulation was performed as a single run. If this model did have an RTD site, this would have been simulated by stopping model execution at the year designated for the RTD action, concentrations in the model where RTD would have occurred would have been set to zero, and then model execution resumed.

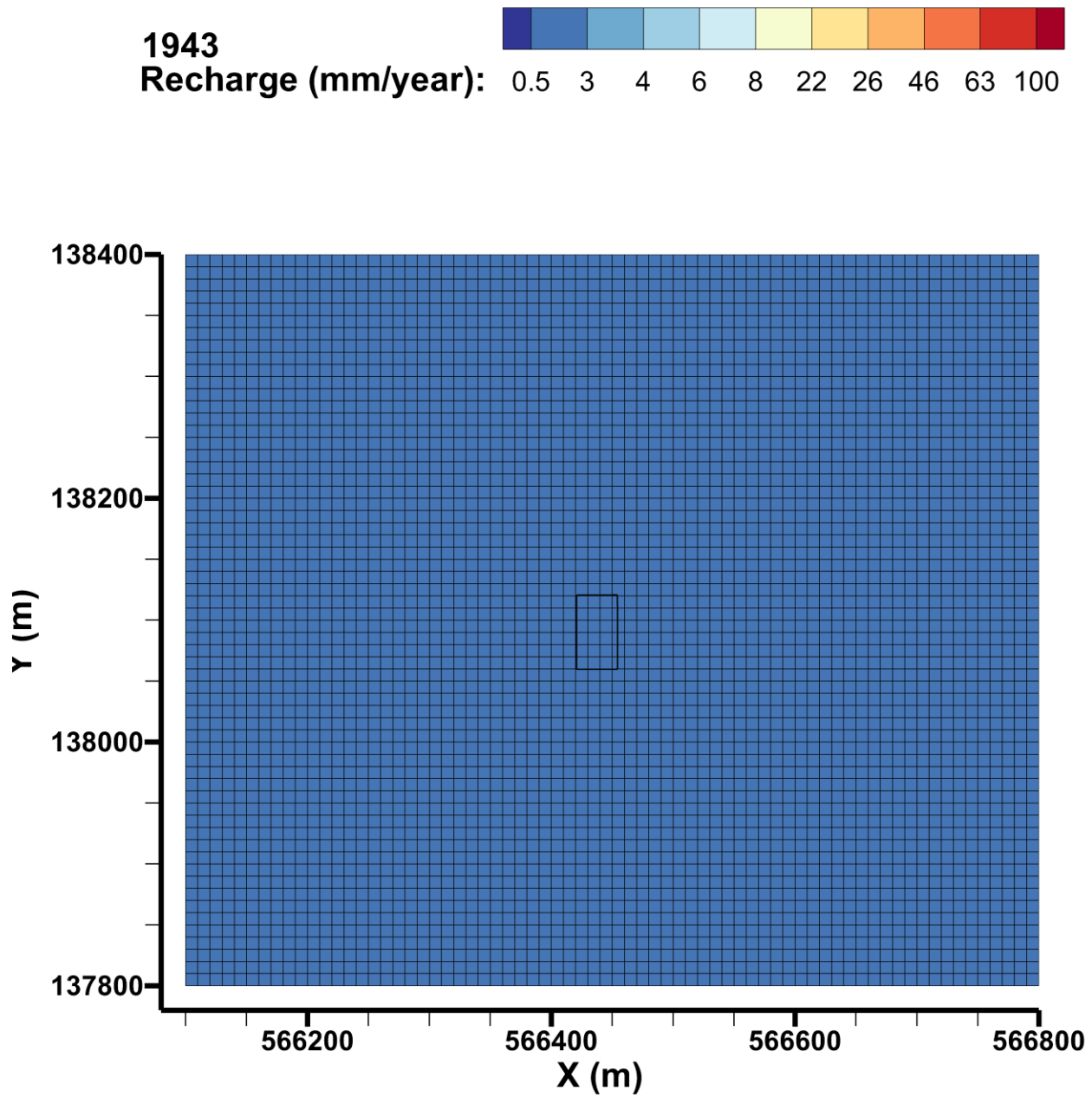
4.8 Boundary Conditions

Boundary conditions for the SALDS model include recharge to the top of the model, water table conditions at the base of the model, and no-flow conditions along the sides of the model. The boundary conditions are described in further detail in the rest of this section.

4.8.1 Natural Recharge – Top Boundary Condition

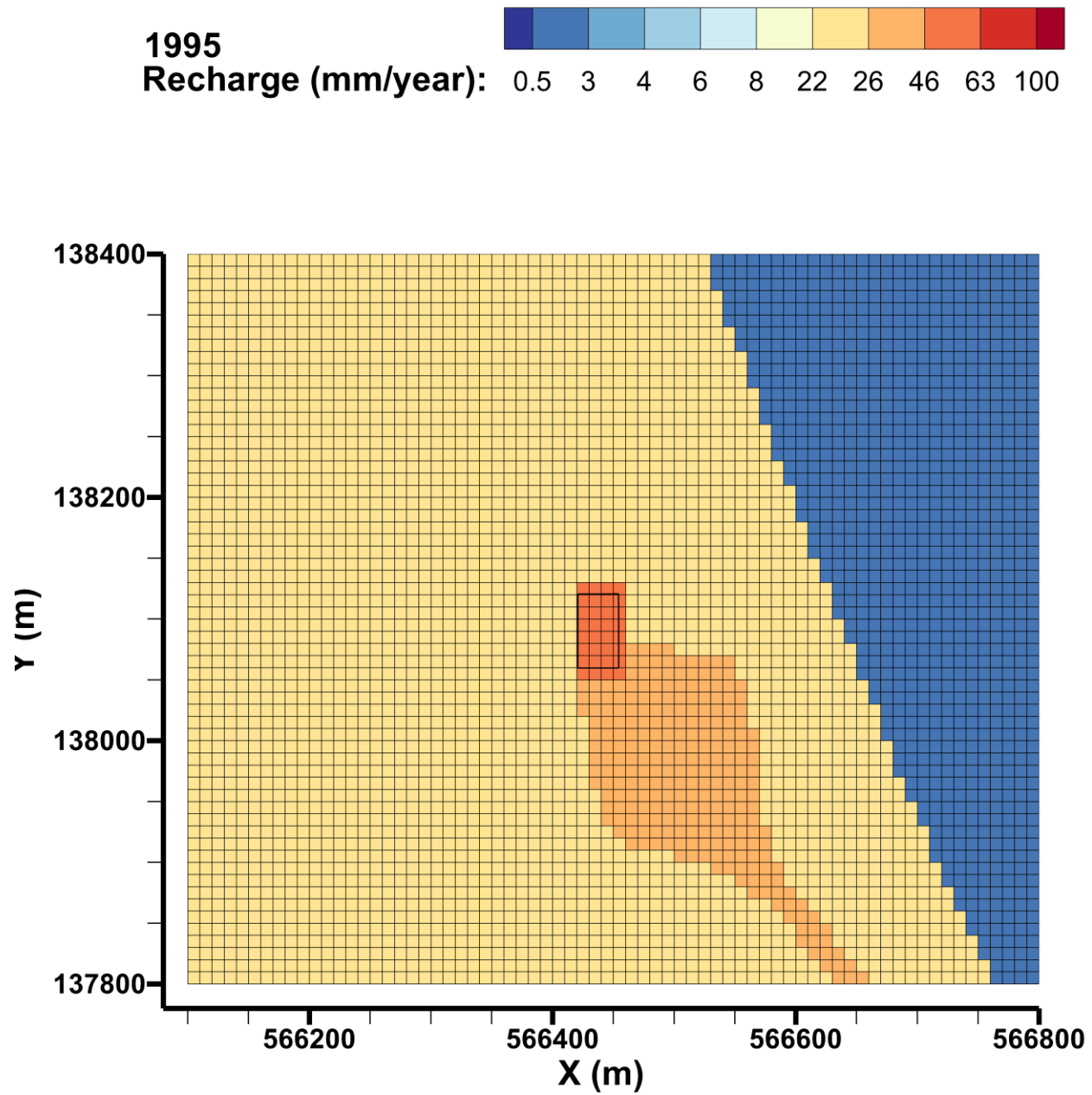
Model recharge was estimated using the RET (ECF-HANFORD-15-0019). The RET assigns soil infiltration rates for the CA vadose zone models based on land use, surface cover information from multiple sources (including existing buildings and structures, waste site footprints, and natural vegetative cover), and soil survey information. Planned future actions for waste site closure are used to develop future recharge estimates through the end of the modeling period. The RET generates spatial representations of recharge estimates for each year from 1943 until recharge reaches a final post-closure condition. These yearly recharge estimates for the model domain are then post-processed to generate the STOMP boundary condition input. The steady-state simulation uses the 1943 RET recharge values for the entire simulation under the assumption that the 1943 recharge is representative of pre-Hanford Site conditions. Recharge rates from every output year from the RET are used as the transient boundary conditions.

Natural recharge within the model domain is spatially variable. Figures of the spatial distribution of RET recharge estimates for the SALDS model are shown for every year there is a change in any recharge estimate in Appendix C. Figure 4-12 to Figure 4-15 show the RET recharge estimates for the SALDS model for 1943, 1995, 2050, and 2550. The pre-Hanford Site recharge rate distribution is determined by the soil type Burbank Loamy Sand covered with mature shrub-steppe plant communities (Figure 4-12). The recharge rate for this soil with mature vegetation is 3.0 mm/yr. As shown in Figure 4-7, development of the SALDS site (600-211) resulted in variable recharge rates over time. Development, including excavation, caused surface disturbances resulting in increased recharge rates. SALDS began operating in 1995 with an estimated recharge rate of 63 mm/yr (Figure 4-13). In 2050, a barrier with an assumed recharge rate of 0.5 mm/yr is planned to cover the site (Figure 4-14). The surface barrier is assumed to have a design life of 500 years, after which the affected area will return to natural conditions with an assigned recharge rate of 4.0 mm/yr (Figure 4-15).



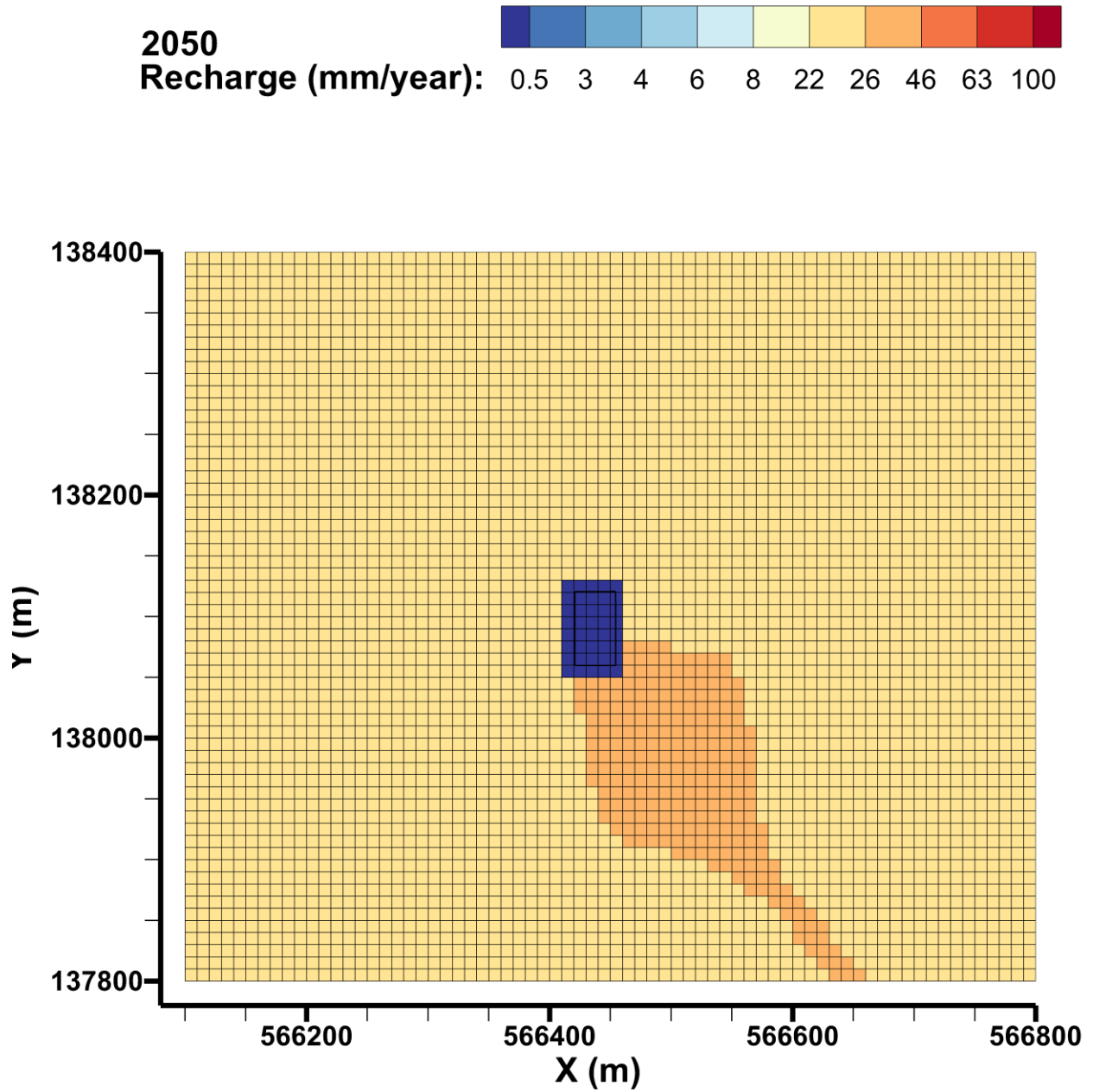
CA_v4-2_salds_SS_RET_rch_1943_CF_2020-07-02

Figure 4-12. Transient Recharge Estimates for the SALDS Model, 1943



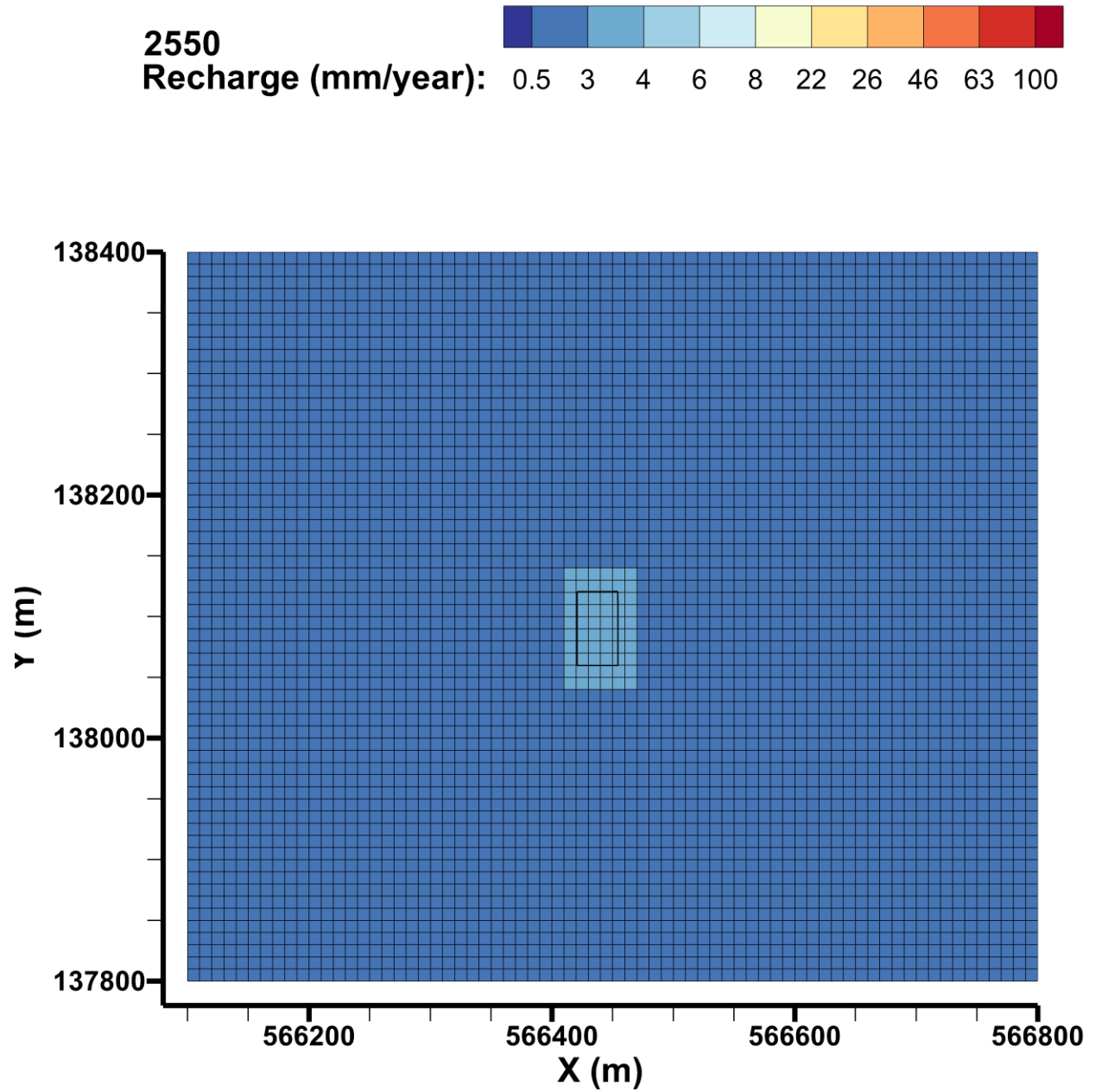
CA_v4-2_salds_SS_RET_rch_1995_CF_2020-07-02

Figure 4-13. Transient Recharge Estimates for the SALDS Model, 1995



CA_v4-2_salds_SS_RET_rch_2050_CF_2020-07-02

Figure 4-14. Transient Recharge Estimates for the SALDS Model, 2050



CA_v4-2_salds_SS_RET_rch_2550_CF_2020-07-02

Figure 4-15. Transient Recharge Estimates for the SALDS Model, 2550

Example time series charts of natural recharge rates for selected locations within the model domain (locations shown in Figure 4-16) are shown in Figure 4-17 through Figure 4-19. Location B (Figure 4-18) on Figure 4-16 represents the SALDS site (600-211, Figure 4-7). The pre-Hanford Site recharge rate is 3.0 mm/yr as determined by the soil type Burbank Loamy Sand covered with mature shrub-steppe plant communities. Development of this area is marked by an initial increase in recharge depending on the activities taking place within the model boundary. At location B, a disposition of “disturbed sand” due to excavation activities and other disturbances is reached at 1995 (when SALDS was constructed), with an assigned recharge rate of 63 mm/yr. This value is consistent with rates measured in unvegetated sands (Table 4.15 in PNNL-14702, *Vadose Zone Hydrogeology Data Package for Hanford Assessments*).

Before reaching the high recharge rates, a cheatgrass cover occurs with a recharge rate of 26 mm/yr for Burbank Loamy Sand. Location B is planned to receive a barrier in 2050 with an assumed rate of 0.5 mm/yr for an expected design life of 500 years. After the expected design life, a final estimated recharge rate of 4 mm/yr is assumed at this location. Location A (Figure 4-17) and location C (Figure 4-19) are not on a waste site, but the area of location C was disturbed by installation of the pipeline conveying effluent to SALDS. For location A, the recharge rate, after development began, increases to 26 mm/yr due to the appearance of a cheatgrass cover on Burbank Loamy Sand. For location C, soil disturbance is assumed to increase the recharge rate to 46 mm/yr. A revegetation cycle with a linear rate decrease over 30 years down to 4.0 mm/yr is imposed on both locations in 2070. There is no barrier emplaced at these locations and the 4.0 mm/yr rate was therefore used from 2100-2270.

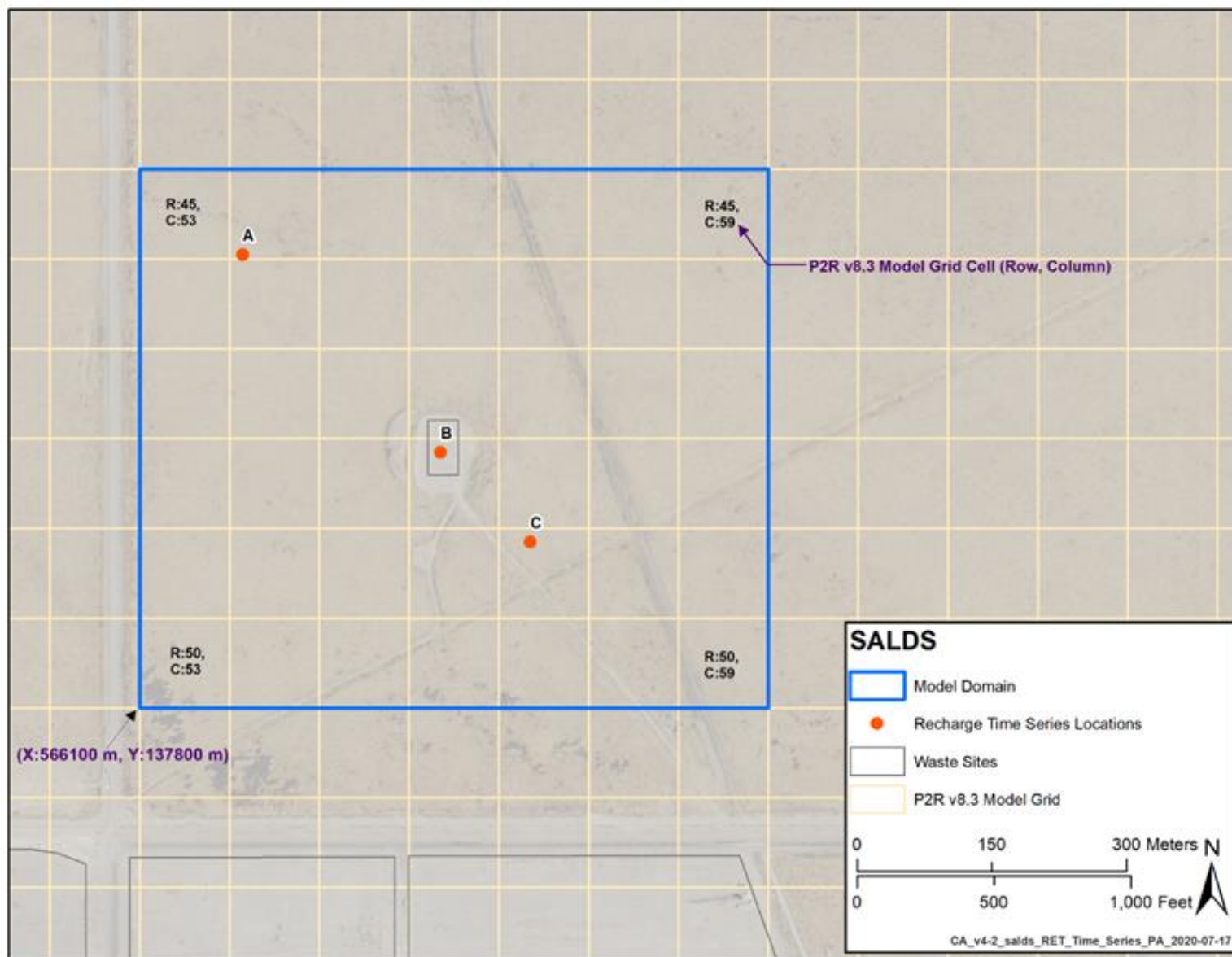
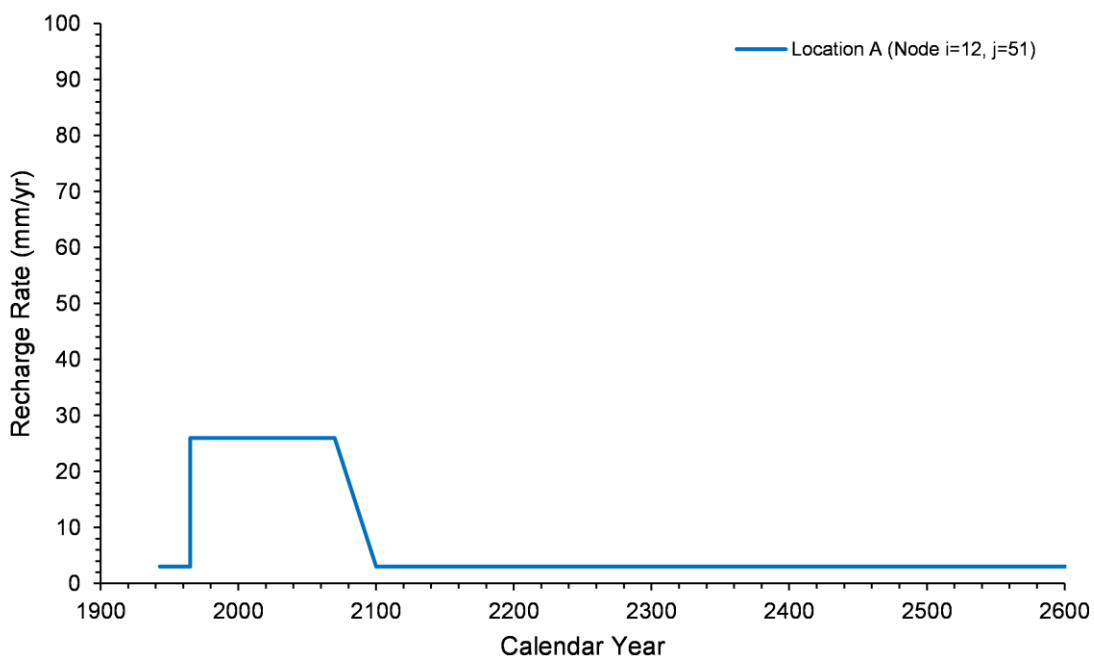
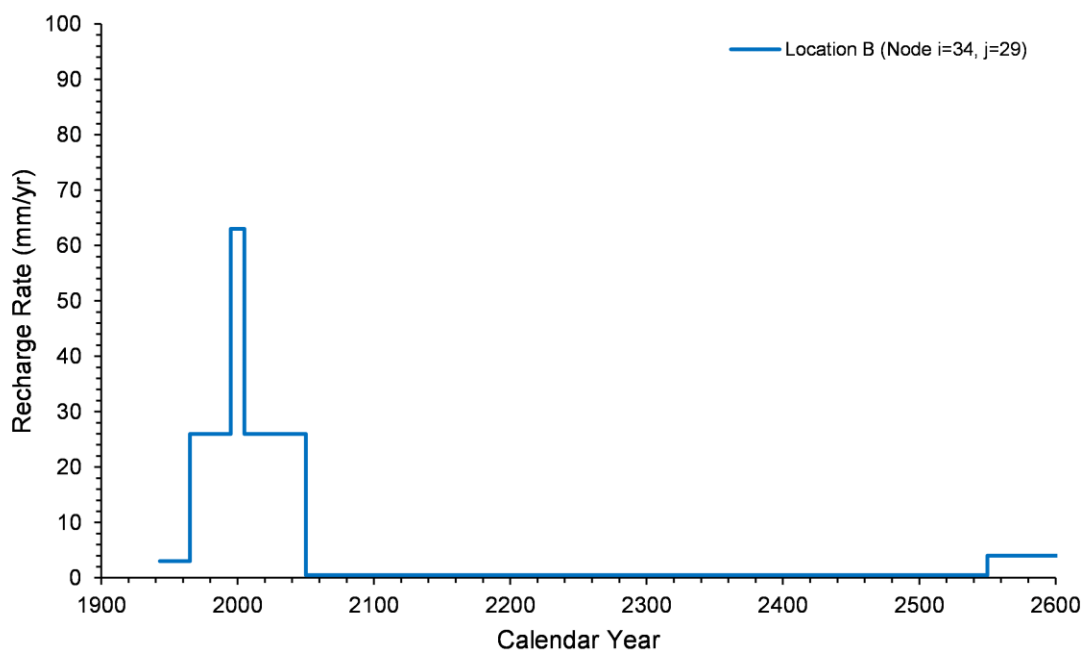


Figure 4-16. Locations of Recharge Rate Time Series Examples



CA_v4-2_salds_recharge_rate_Location_A_pa_2020-06-30

Figure 4-17. Time Series of Natural Recharge Rates, Location A



CA_v4-2_salds_recharge_rate_Location_B_pa_2020-06-30

Figure 4-18. Time Series of Natural Recharge Rates, Location B

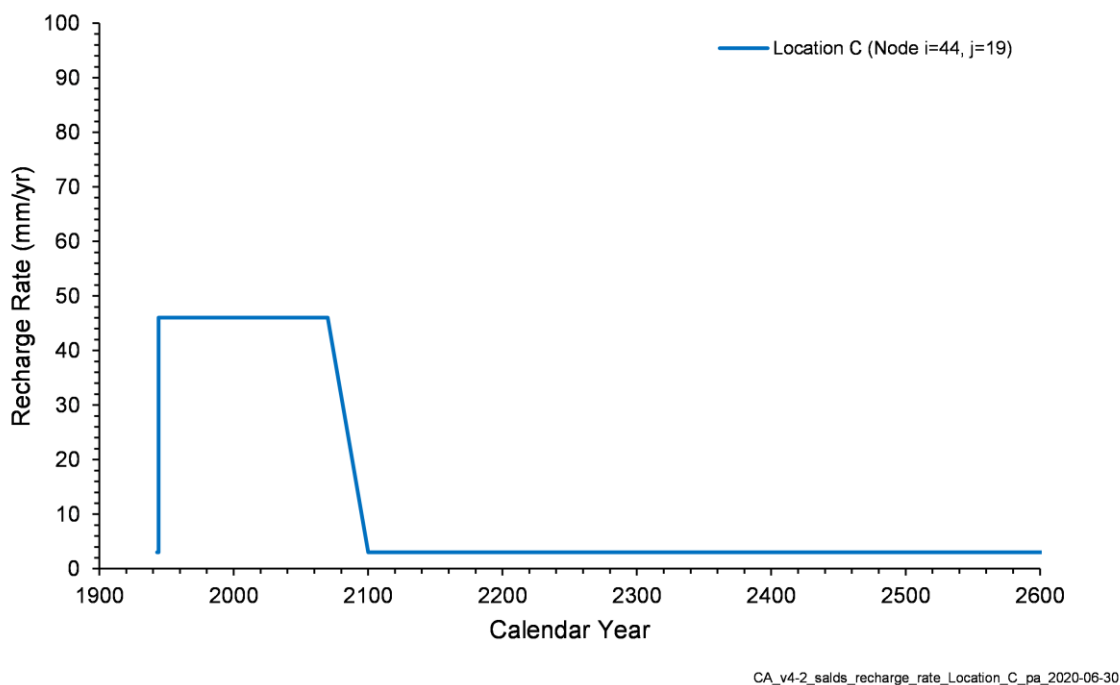


Figure 4-19. Time Series of Natural Recharge Rates, Location C

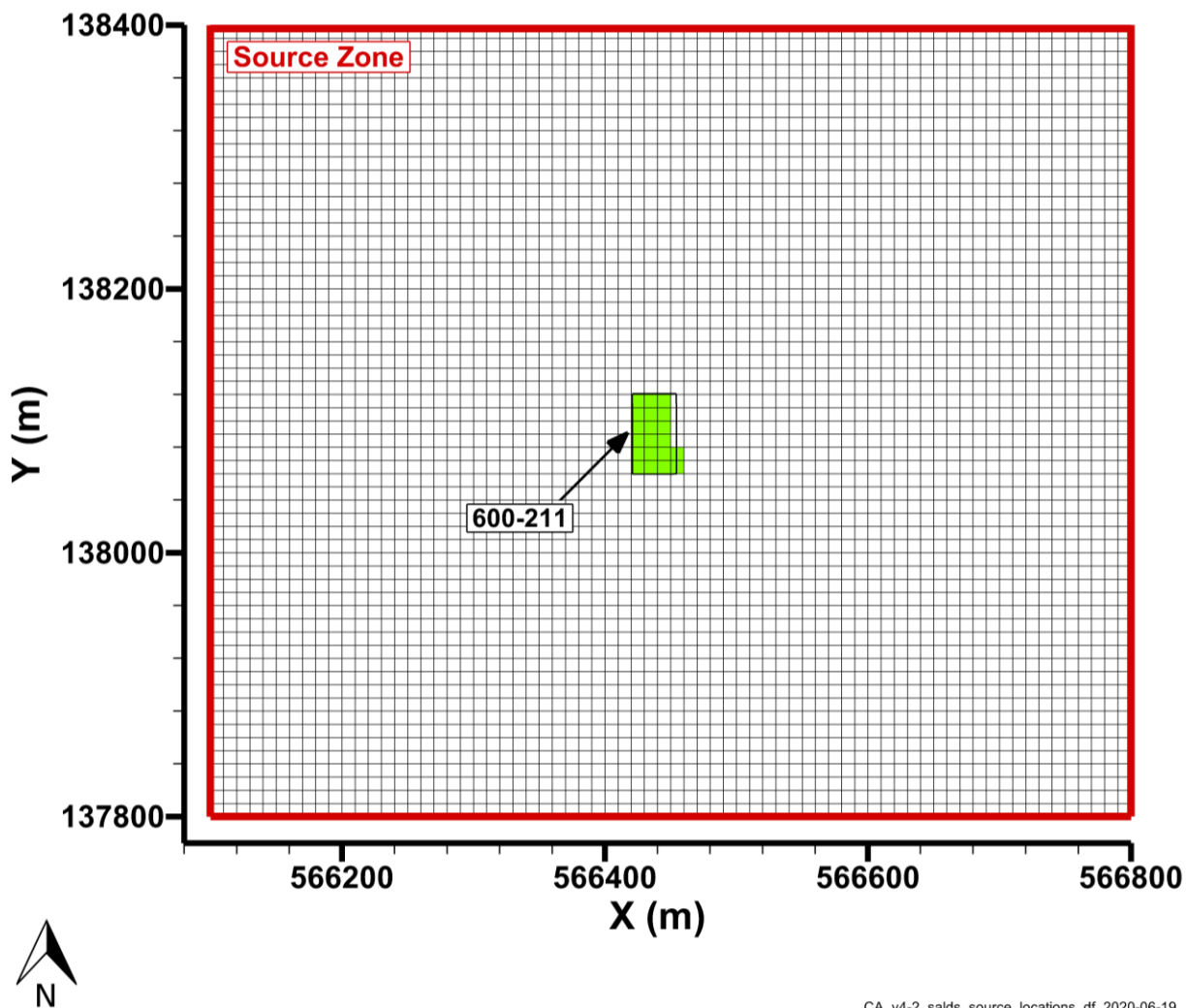
4.8.2 Lateral and Bottom Boundaries

Lateral boundaries for the model are assumed to be zero-flux boundaries for both contaminant transport and water flow. The locations of the lateral boundaries were selected in an iterative procedure to ensure that the contaminant plumes would not reach the model boundary. Source zone waste sites with radionuclide and liquid releases were at least 100 m away from the model boundary so that the releases would not affect soil moisture or contents at or near the boundary. For elongated waste sites extending into adjacent models, the assumption is that bifurcation of a waste site by a model boundary does not lead to soil moisture gradients across the boundary and that zero-flux boundaries are therefore appropriate for such waste sites.

The bottom of the model was assumed to be coincident with the water table at the model location, as estimated from the 2017 water table elevation (ECF-HANFORD-17-0120, *Preparation of the March 2017 Hanford Site Water Table and Potentiometric Surface Maps*). This boundary was represented by a Dirichlet boundary condition with a pressure of 101,325 Pa.

4.9 Source Nodes

Radionuclides and water discharged from the waste site are introduced to this model at source nodes. The distribution of these source nodes is shown in Figure 4-20. The STOMP Source Cards (i.e., specific information on source location and releases in the STOMP input file) were built using waste site footprints, source inventory, and the model grid. A discussion of the source node allocation process is found in CP-63515.



Note: Black cells indicate source nodes with input from multiple sites.

Figure 4-20. Distribution of Source Nodes in the SALDS Model

4.9.1 Data Reduction

The solid waste inventory from ECF-HANFORD-19-0112 described in Section 4.5.1.2 is released over approximately 10,000 years, with the total release timespan varying by waste site. These long release periods had many timesteps, resulting in large STOMP Source Cards. To accommodate the size limitations of STOMP Source Cards, the original inventory datasets were modified to release the solid waste inventory in a number of timesteps that is compatible with the Source Card size limitations. The reduced datasets were checked to ensure they adequately represent the original inventory amounts and release rates. Additional information regarding the data reduction methodology is documented in ECF-HANFORD-20-0006.

4.10 Modeling Assumptions

The development of the SALDS model required several conceptual and simulation assumptions. The major assumptions are as follows:

- The vadose zone model consists of a system of HSUs derived from the Central Plateau Vadose Zone Geoframework Model (CP-60925). The geoframework is a three-dimensional representation of the subsurface beneath the Central Plateau, vertically extending from the ground surface to the top of the Columbia River Basalt Group. The geoframework model is constructed using a combination of lithologic and sequence stratigraphic interpretations, leading to the definition of a series of HSUs. With this approach, correlated, hydraulically significant units are mapped while still representing the interpretations of lithologically heterogeneous features. The HSU surfaces used in generating the SALDS model are from an update to CP-60925, ECF-HANFORD-18-0035.
- The anisotropic equivalent homogeneous media (EHM) approach is used to simulate flow and transport in the heterogeneous Central Plateau HSUs. The EHM approach is recommended by Yeh et al., 2015, “Flow Through Heterogeneous Geologic Media,” for systems with large-scale HSUs. With this approach, an HSU has two main characteristics: (1) representative hydraulic property and parameter values are applied that are equivalently homogeneous (i.e., constant) in space, and (2) the effects of heterogeneity on flow are described using an anisotropic unsaturated hydraulic conductivity. An important feature of an anisotropic EHM model representation is that it captures the mean or the bulk flow characteristics of the vadose zone moisture plumes, as demonstrated by Zhang and Khaleel, 2010, “Simulating Field-Scale Moisture Flow Using a Combined Power-Averaging and Tensorial Connectivity-Tortuosity Approach.” Therefore, the contaminant peak arrival time under recharge-dominated flow conditions is adequately captured by an anisotropic EHM model representation. The anisotropic EHM approach is commonly used to model flow and transport at the Hanford Site. For instance, recent PA vadose modeling for WMA C (RPP-ENV-58782, *Performance Assessment of Waste Management Area C, Hanford Site, Washington*) used this approach to simulate subsurface flow and transport.
- For simulation of flow in unsaturated Hanford Site sediments, the soil water retention relation (i.e., the relation between soil moisture content and capillary pressure) and the unsaturated hydraulic conductivity relation (i.e., the relation between moisture content and unsaturated hydraulic conductivity) need to be provided. The unsaturated hydraulic conductivity is the product of the saturated hydraulic conductivity and the aqueous phase relative permeability. The nonhysteretic van Genuchten equation (van Genuchten, 1980) is used for the soil water retention relation. The Mualem relation (Mualem, 1976) is used for the unsaturated hydraulic conductivity relation.
- For the heterogeneous stratified sediments at the Central Plateau, upscaled hydraulic properties based on small-scale laboratory measurements are used to simulate the large, field-scale behavior. This assumption requires that each heterogeneous HSU be replaced by an anisotropic EHM with upscaled hydraulic properties. The hydraulic properties used in the CA model are on a grid-block scale which are much larger than the cores that are typically analyzed in the laboratory.
- The upscaled grid-block-scale parameter values for the water retention and relative permeability relations are obtained by applying averaging procedures to core-scale data. For the soil water retention relation, the linear upscaling scheme (Green et al., 1996, “Upscaled Soil-Water Retention Using Van Genuchten’s Function”) is applied. For the unsaturated hydraulic conductivity, the power-averaging tensorial connectivity-tortuosity (PA-TCT) method (Zhang et al., 2003, “A Tensorial Connectivity–Tortuosity Concept to Describe the Unsaturated Hydraulic Properties of Anisotropic Soils”; Zhang and Khaleel, 2010) is used to determine directionally-dependent saturated hydraulic conductivity and relative permeability tortuosity parameters that are functions of the soil moisture content. The PA-TCT upscaling method leads to a soil-moisture-dependent anisotropic unsaturated hydraulic. Applying the PA-TCT method allows for an assessment of the effects of heterogeneity on lateral flow and contaminant spreading, including plume commingling at the HSU

scale. The method has been successfully applied to evaluate various water infiltration tests performed at the Sisson and Lu field experiment site in the 200 East Area (Ye et al., 2005, “Stochastic Analysis of Moisture Plume Dynamics of a Field Injection Experiment”; Zhang and Khaleel, 2010). The field applications of the upscaled vadose zone property values based on the PA-TCT method suggests that it provides a reasonable framework for upscaling core-scale measurements, as well as an accurate simulation of moisture flow in the heterogeneous vadose zone under the Central Plateau.

- The CA vadose zone models use a “forward” modeling approach for contaminant transport in the subsurface: model transport simulations initiate at a time when contamination is not present in the subsurface, and the contaminant activity is introduced in the models as sources over time. This approach has been used to simulate Hanford Site contaminant transport resulting from liquid waste disposal (e.g., Oostrom et al., 2017, “Deep Vadose Zone Contaminant Flux Evaluation at the Hanford BY-Cribs Site Using Forward and Imposed Concentration Modeling Approaches”) and past leaks (RPP-RPT-59197, *Analysis of Past Waste Tank Leaks and Losses in the Vicinity of Waste Management Area C, Hanford Site, Washington*).
- Contaminant activity is assumed to be transported in the vadose zone by advection and hydrodynamic dispersion, which is the sum of molecular diffusion and mechanical dispersion. The two components of hydrodynamic dispersion are described by a single hydrodynamic dispersion coefficient and treated as a diffusive flux proportional to the concentration gradient. Advective transport and mechanical dispersion are computed using the flow field obtained when solving the water conservation equation. The contaminants are considered to be solutes, without affecting fluid properties like density and viscosity.
- Mechanical dispersion is assumed to be directionally dependent with a constant macroscopic macrodispersivity value for each HSU. The use of a constant (asymptotic) macrodispersivity for large-scale vadose zone CA modeling is considered appropriate (NUREG/CR-5965, *Modeling Field Scale Unsaturated Flow and Transport Processes*). Macrodispersivity values for the HSUs in the longitudinal direction, are obtained from Hanford Site field-scale numerical simulations and field experiments. Hanford Site-specific datasets include Khaleel et al., 2002, “Upscaled Flow and Transport Properties for Heterogeneous Unsaturated Media”; and PNNL-25146, *Scale-Dependent Solute Dispersion in Variably Saturated Porous Media, RPT-IGTP-009*. In the absence of unsaturated media experimental data, the CA transport models used a transverse macrodispersivity value that is 1/10th of the obtained longitudinal value.
- Contaminant sorption is simulated using a reversible linear sorption isotherm with a linear K_d . The linear sorption model approach is assumed to be adequate for modeling transport at the Hanford Site (PNNL-13895, *Hanford Contaminant Distribution Coefficient Database and Users Guide*). An important benefit of the linear adsorption assumption is that an extensive database of K_d values applicable to Hanford Site sediments is available for the contaminants of most concern over a broad range of conditions (e.g., PNNL-17154, *Geochemical Characterization Data Package for the Vadose Zone in the Single-Shell Tank Waste Management Areas at the Hanford Site*). Use of reversible linear K_d isotherms is computationally efficient and appropriate for the scale of the CA problem. Recognizing that experimental K_d values are mostly determined using sediment grain sizes < 2 m, corrections for gravel content using equations provided in PNNL-17154 are used to adjust measured values for the finer fraction applicable to HSUs with considerable gravel content.
- The spatial and temporal variable natural recharge rate is used to define the upper boundary conditions for the water conservation equation. The natural recharge rate is a term applied to define the net infiltration that migrates through the vadose zone to reach the water table. At the Hanford Site,

this rate is primarily a function of the surface soil type and type/density of vegetative cover. Effects of climate change on natural recharge over the next 10,000 years are not accounted for in the simulations.

- No moisture or contaminants are allowed to migrate across the lateral boundaries of the model domain. During development of the model domain, the proper locations of the zero flux lateral boundaries were determined in an iterative procedure.
- The simulations use a fixed water table representing 2018 conditions to increase efficiency and reduce complexity during implementation of the vadose zone models. The effects of the transient water table on contaminant transfer after 2018 to the aquifer were evaluated to validate this approach in Farrow et al., 2019, “Prediction of Long-Term Contaminant Flux from the Vadose Zone to Groundwater for Fluctuating Water Table Conditions at the Hanford Site.” Simulations for selected vadose zone models with continuing sources demonstrated that a simplification of the water table boundary condition (i.e., a static water table), could be adequately used to compute long-term predictions of contaminant flux to groundwater.
- The liquid volumes and waste site inventories for most of the CA models are obtained from the Hanford Soil Inventory Model (SIM-v2) (ECF-HANFORD-17-0079). The liquid volumes and waste site inventory for SALDS were obtained from EMDT-IN-0047. Non-radiological site liquid volumes were obtained from site-specific literature. Using geometry information, waste and non-radiological site shapes were assigned to vadose zone model grid surfaces, according to EMDT-GR-0035⁴, *Waste Site and Structure Footprint Shapefiles for Inclusion in Updated Composite Analysis*. Water volumes and SIM-v2 contaminant inventories were assigned to the model grid cells at the lowest topographic location within the site footprints.

⁴ EMDT-GR-0035, *Waste Site and Structure Footprint Shapefiles for Inclusion in Updated Composite Analysis*, Rev. 0, CH2M HILL Plateau Remediation Company, Richland, Washington. Electronic model data transmittals are stored in the Environmental Model Management Archive. A copy of the cover sheet for this EMDT is provided in Appendix F.

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5 Software Applications

Three types of calculation software are used in this modeling effort: the numerical modeling simulator eSTOMP, support software (spreadsheet and geographic information system [GIS] applications), and custom utility calculation software. Custom utility calculations software is documented under CHPRC-04032, *Composite Analysis / Cumulative Impact Evaluation (CACIE) Utility Codes Integrated Software Management Plan* and described in further detail in Section 5.3 of this ECF.

5.1 Approved Software

The eSTOMP numerical simulator has been used for the flow and transport calculations reported in this ECF. The application of the simulator is managed under the requirements of CHPRC-00176, *STOMP Software Management Plan*. Use of this software is consistent with the intended uses of STOMP at the Hanford Site as defined in CHPRC-00222, *STOMP Functional Requirements Document*. The STOMP software is actively managed by the CH2M HILL Plateau Remediation Company and approved for use at the Hanford Site as Level C software under a procedure that implements the requirements of DOE O 414.1D, *Quality Assurance*.

Build 6 of the STOMP software was used in the implementation of the model described in this document. This version was approved for use at the Hanford Site based on acceptance testing results reported in CHPRC-00515, *STOMP Acceptance Test Report*. The status of requirements for this software are maintained in CHPRC-00269, *STOMP Software Requirements Traceability Matrix*. All acceptance testing was performed to the requirements of CHPRC-00211, *STOMP Software Test Plan*. Installation testing is also required for any computer system on which STOMP is run. The installation test is specified in CHPRC-00211.

The STOMP simulator was developed by Pacific Northwest National Laboratory to simulate flow and transport over multiple phases in a subsurface environment. The water mode of the simulator uses numerical approximation techniques to solve partial differential equations that describe the conservation of aqueous mass and radionuclide activity in variably saturated porous media. These governing conservation equations, along with a corresponding set of constitutive relations that relate variables within the conservation equations, are solved numerically by using integrated-volume, finite-difference discretization to the physical domain and first- or second-order Euler discretization to the time domain. The resulting equations are nonlinear, coupled algebraic equations that are solved using the Newton-Raphson iteration.

The theoretical and numerical approaches applied in the STOMP simulator are documented in a published theory guide (PNNL-12030). The simulator has undergone a rigorous verification procedure against analytical solutions, laboratory-scale experiments, and field-scale demonstrations. The application guide (PNNL-11216) provides instructive examples in the application of the code to classical groundwater and vadose zone flow and transport problems. The user's guide (PNNL-15782) describes the general use, input file formatting, compilation, and execution of the code.

- Software Title: STOMP, parallel implementation (eSTOMP), executable eSTOMP1-chprc06-20200204-g.x
- Software Version: CHPRC Build 6
- Hanford Information Systems Inventory Identification Number: 2471

- Workstation type and property number (from which software is run): GAIA Subsurface Flow and Transport Modeling Platform, Nodes compute-0-0 through compute-0-8 inclusive, property tags: WF32991, WF32992, WF32993, WF32994, WF32995, WF32996, WF32997, WF32998, WF32999

5.1.1 Software Installation and Checkout

The software installation and checkout form for STOMP simulation software is provided as Appendix D to this ECF.

5.1.2 Statement of Valid Software Application

The application of the eSTOMP software to the vadose zone flow and transport systems is correct. The software has been used within the limits discussed in the simulator's theory guide (PNNL-12030) and user's guide (PNNL-15782). The water mode of the STOMP simulator is designed to simulate flow and transport over multiple phases in a subsurface environment, including unsaturated systems like the Hanford Site vadose zone. The simulator solves partial differential equations describing conservation of aqueous mass and radionuclide activity in variably saturated porous media, consistent with aqueous flow and contaminant transport in Hanford Site sediments. The STOMP code has been executed at research institutions and universities to address vadose zone flow and contaminant transport problems comparable to the CA unsaturated systems.

The STOMP code, including the eSTOMP parallel implementation, is developed and tested to NQA-1, *Quality Assurance Requirements for Nuclear Facility Applications*, standards by Pacific Northwest National Laboratory "by option" wherein testing conducted option by option. Therefore, an "NQA-1 Options Analysis" is provided for the model application documented in this ECF (as well as other related model applications) in CP-63515 to demonstrate that all eSTOMP code options used in this model are NQA-1 qualified.

5.2 Support Software

The following programs are classified as Support Software

- **Microsoft® Excel®** (version 2010): The tool was used to generate inventory plots and contaminant release and transfer timeseries.
- **ArcGIS®** (version 10.3.1): The tool was used to create of spatial model discretization and waste site location maps.
- **Tecplot® 360 EX** (version 2018R1): The tool was used to generate source location, recharge distribution, and mass transfer to groundwater plots.

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® ArcGIS® is a registered trademark of the Environmental Systems Research Institute, Inc., Redlands, California.

® Tecplot is a registered trademark or trademarks of Tecplot, Inc. in the United States and other countries.

5.3 Support Scripts

Generation of model input files and post-processing of model results was mostly performed with utility codes (scripts) that are managed, tested, and controlled in accordance with CHPRC-04032.

CHPRC-04032 provides a common foundation for the management of several custom-developed scripts to manage pre- and post-processing operations and inter-facet information passing between major software packages efficiently for the CA. It also provides direction for electronic management of documentation requirements at the script level with respect to individual tool functional requirements, software requirements specification, software design description, requirements tracing, test plans and reporting, and user documentation. The utility scripts developed for this project, in alphabetical order, are as follows:

- **aq_mod_avg.exe:** The Aqueous Source Averaging Tool averages aqueous source rates for user-specified waste sites and times.
- **ca_build_surface_flux.py:** The Build Surface Flux Tool maps the STOMP grid into the MODFLOW grid.
- **ca-dups.pl:** The Duplicate Source Nodes Tool identifies any source nodes that overlap spatially and writes information regarding the duplicate source node(s) to an output file.
- **ca-getmod_srf.pl:** The Surface File to P2R Tool aggregates solute flux and cumulative discharge data exiting the vadose zone model by P2R grid cell.
- **ca-ipp.pl:** The Inventory Pre-Processor Tool creates a comprehensive dataset consisting of radionuclide and aqueous volume releases as a function of time for Central Plateau sites. The dataset is input for the SRC2STOMP Tool.
- **ca-merge_srf.pl:** The STOMP Surface Merge Tool merges STOMP surface file data from two consecutive STOMP simulations (e.g., surface files for the 2018 to 12070 simulation).
- **ca-patchbowl.pl:** The Patchbowl Tool modifies STOMP soil zonation files to patch holes in the silt layers of the perching silt layer in the 200 East Area.
- **ca_RET2STOMP.py:** The RET2STOMP Tool generates the natural recharge Boundary Condition Cards for the STOMP model input file using output generated by the RET (Recharge Evolution Tool).
- **ca-rtdic.pl:** The RTD Initial Conditions Card Tool generates Initial Conditions Cards at RTD years for models with RTD sites using an input source card file and a steady-state STOMP input file.
- **ca-src2stomp.pl:** The SRC2STOMP Tool combines the site spatial information with the corresponding radionuclide inventory and creates a STOMP-readable Source Card file containing grid cell definitions of solute and/or liquid sources.
- **K2S_ROCSAN.exe:** The Kingdom2Stomp Tool reads an input file representing each node in the model and generates an output file like the input file with the addition of which geologic formation each model node represents.
- **ModelSetupFY18.jar:** The Composite Analysis STOMP Tool is a graphical user interface tool that produces STOMP input files based on user input model dimensions and material properties.

- **OC_SS_gen.exe:** The Steady-State Output Card Generator Tool reads files generated by the Composite Analysis STOMP Tool and generates a STOMP Output Control Card for the steady-state simulation.
- **OC_rad_gen.exe:** The Transport Output Card Generator Tool Creates a STOMP Output Control Card used for mass balance and transport production simulations.
- **reroute_sources.exe:** The Source Rerouting Tool redistributes wastewater volumes and contaminant inventories for the 216-U-10 Pond System and the 216-B-3 Pond System.
- **splitKingdomLayer.pl:** The SplitKingdomLayer Tool is used to split one geology surface layer file into two sub-unit surface layer files based on the information specified in the polygon file.
- **srcloc_modify.exe:** The Source Node Moving Tool moves source nodes from the locations selected by the SRC2STOMP Tool.
- **SS_input_gen.exe:** The Steady-State STOMP Input File Generator Tool generates the STOMP input file for the steady-state simulation.
- **xprt_2018_input_gen.exe:** The 2018 STOMP Input File Generator Tool generates the 1943–2018 STOMP transport input file.
- **xprt_12070_input_gen.exe:** The 12070 STOMP Input File Generator Tool generates the 2018 (or RTD year if the model has RTD remediation sites)–12070 STOMP transport input file. This code reads and modifies the 1943–2018 STOMP input file created by the 2018 STOMP Input File Generator Tool.
- **xprt_mb_input_gen.exe:** The Mass Balance STOMP Input File Generator Tool generates the mass balance STOMP transport input file. This code reads and modifies the STOMP input file created by the 2018 STOMP Input File Generator Tool.
- **xprt_RTD_input_gen.exe:** The RTD STOMP Input File Generator tool generates the 2018 – RTD year STOMP transport input file. This code reads and modifies the 1943–2018 STOMP input file created by the 2018 STOMP Input File Generator Tool.

6 Calculation

The fate and transport calculations for the SALDS model were performed using a suite of STOMP simulations: a steady-state simulation, mass balance transport simulations, and historical and forecast transport simulations (as discussed in Section 4.6). This section describes the mass balance calculations for the steady-state and transport simulations.

6.1 Steady-State Simulation

The purpose of the steady-state simulation was to verify model performance and to generate the initial primary variable (i.e., aqueous pressure) conditions within the model domain for the historical transport simulations, as discussed in Section 4.6.1. Contaminants are not simulated in the steady-state simulation, only flow. Pre-Hanford Site boundary conditions (i.e., natural recharge rates for 1943) are applied for a period of 10,000 years (from year zero to 10,000) to allow the simulation to reach steady-state conditions. Figure 6-1 compares the steady-state recharge flux into the top of the model to the flux leaving the base of the model, which represents discharge to groundwater from the model. Conditions reach equilibrium (i.e., flux in equals flux out) and remain unchanged through the end of the simulated time period, indicating that steady-state conditions have been achieved.

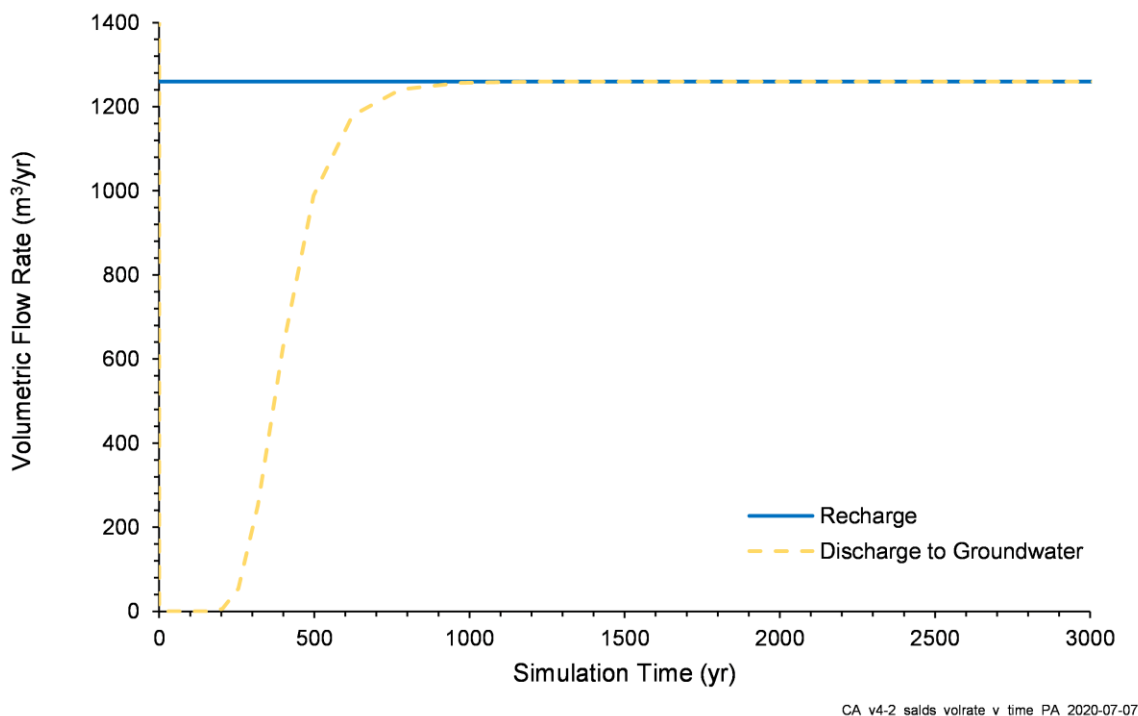


Figure 6-1. Steady-State Recharge Compared to Discharge to Groundwater Over Time

The steady-state liquid volume balance (also called mass balance) error (E) is calculated as shown in Equation 6-1 (all variables have units of volume):

$$E = (S + O) - R_P \quad (\text{Eq. 6-1})$$

where:

- E = liquid volume balance error
- S = change in liquid storage within the model domain
- O = total liquid outflow from the model domain
- R_P = total pre-Hanford Site natural recharge.

The percent relative error ($\%RE$) of the aqueous volume balance is calculated as shown in Equation 6-2:

$$\%RE = 100|E/R_P| \quad (\text{Eq. 6-2})$$

where $\%RE$ is the liquid volume percent relative error.

Change in liquid storage (S) is the difference between liquid in the model at year 10,000 and year 0. Total liquid water outflow from the model (O) is the cumulative liquid volume that passed through the bottom of the model boundary at the end of 10,000 years. The pre-Hanford Site natural recharge (R_P) is the cumulative volume of recharge applied to the top layer of the model during the simulation. The flow-only steady-state liquid volume balance is shown in Table 6-1.

Table 6-1. Liquid Volume Balance for the SALDS Model Steady-State Simulation

Natural Recharge (R_P) ^a	Change in Liquid Storage (S) ^{a,b}	Total Liquid Outflow (O) ^{a,b}	Error (E) ^a	Percent Relative Error ($\%RE$)
12,600,000	473,779	12,126,340	119	9.445E-04

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a. Volume units in m³.

b. Calculated by STOMP.

- $\%RE$ = liquid volume percent relative error
- E = liquid volume balance error
- O = total liquid outflow from the model domain
- R_P = total pre-Hanford Site natural recharge
- S = change in liquid storage within the model domain
- SALDS = State-Approved Land Disposal Site
- STOMP = Subsurface Transport Over Multiple Phases

6.2 Contaminant Transport Volume and Activity Simulations

Transient simulations were used to calculate liquid volume and activity balances, also referred to as mass balances. These simulations use the steady-state model final aqueous pressure distribution as initial aqueous pressure conditions, the transient natural recharge described in Section 4.8.1, and the waste site sources described in Section 4.5. Although run as single simulations for each radionuclide group, two sets of radionuclide activity balance evaluations were performed: the first for the historical time period from 1943 to 2018, and the second for the entire transient model duration from 1943 to 12070. Radionuclide half-life values were set to 1.0E+20 years to virtually eliminate radioactive decay. Therefore, decay corrections were not necessary, and the radionuclide activity balance could be evaluated directly.

The liquid volume balance error (E) is calculated as shown in Equation 6-3 (all variables have units of volume):

$$E = (S + O) - (I + R) \quad (\text{Eq. 6-3})$$

where:

- E = liquid volume balance error
- S = change in liquid storage within the model domain
- O = total liquid outflow from the model domain
- I = liquid inventory entering the model domain from liquid waste site releases
- R = total natural recharge.

The percent relative error ($\%RE$) of liquid volume balance is calculated as shown in Equation 6-4:

$$\%RE = 100|E/(I + R)| \quad (\text{Eq. 6-4})$$

where $\%RE$ is the liquid volume percent relative error.

The change in liquid storage within the model domain (S) is the difference between the volume of water in the model at the beginning of the simulation (1943) and the end of the mass balance analysis period (either 2018 or 12070). The total liquid outflow from the model domain (O) is the cumulative liquid volume that passed through the bottom of the model boundary by the end of the mass balance analysis period. The liquid inventory entering the model domain from liquid waste site releases (I) is the cumulative volume of liquids released to the model from the liquid waste site in the source zone during the mass balance analysis period. The natural recharge (R) is the cumulative volume of liquid applied to the top of the model from natural recharge during the mass balance analysis period. The liquid volume balance for the SALDS model for the simulation for Radionuclide Group 1 is shown in Table 6-2.

Table 6-2. Transient Liquid Volume Balances for the SALDS Model Radionuclide Group 1 Simulations

Liquid Inventory (I) ^a	Natural Recharge (R) ^a	Change in Liquid Storage (S) ^{a,b}	Total Liquid Outflow (O) ^{a,b}	Error (E) ^a	Percent Relative Error ($\%RE$)
1943–2018					
1,237,360	562,373	378,688	1,421,141	96	5.320E-03
1943–12070					
3,470,760	13,958,625	599	17,425,120	-3,665	2.103E-02

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a. Volume units in m³.

b. Calculated by STOMP.

- $\%RE$ = liquid volume percent relative error
- E = liquid volume balance error
- I = liquid inventory entering the model domain from liquid waste site releases
- O = total liquid outflow from the model domain
- R = total natural recharge
- S = change in liquid storage within the model domain
- SALDS = State-Approved Land Disposal Site
- STOMP = Subsurface Transport Over Multiple Phases

The radionuclide activity balance error (E_R) is calculated as shown in Equation 6-5 (all variables have units of activity):

$$E_R = (S_R + O_R) - I_R \quad (\text{Eq. 6-5})$$

where:

- E_R = radionuclide activity balance error
- S_R = radionuclide storage within the model domain at the end of the simulation
- O_R = total radionuclide outflow from the model domain
- I_R = radionuclide inventory entering the model domain from waste site releases.

The percent relative error ($\%RE_R$) of the radionuclide activity balance is calculated as shown in Equation 6-6:

$$\%RE_R = 100|E_R/I_R| \quad (\text{Eq. 6-6})$$

where $\%RE_R$ is the radionuclide activity balance percent relative error.

The total radionuclide outflow (O_R) is the cumulative activity of a particular radionuclide that migrated through the bottom boundary of the vadose zone model from the beginning of the simulation (1943) to the end of the mass balance analysis period (either 2018 or 12070). The radionuclide storage (S_R) is the difference in total activity of a particular radionuclide in the model from the beginning of the simulation (1943) and the end of the mass balance analysis period (2018 or 12070). Because there were no radionuclides in the model from anthropogenic sources in 1943, this can be understood as the change in total activity of a radionuclide in the model domain. The radionuclide inventory that entered the model domain from waste site releases (I_R) is the cumulative activity of the radionuclide released to the model from the liquid waste release site in the source zone. Table 6-3 shows the activity balance for the SALDS model no-decay transport simulations for Radionuclide Group 1.

Table 6-3. Transient No-Decay Activity Balances for the SALDS Model Radionuclide Group 1 Simulations

Radionuclide	Released Radionuclide Inventory (I_R) ^a	Radionuclide Storage (S_R) ^{a,b}	Radionuclide Outflow (O_R) ^{a,b}	Error (E_R) ^a	Relative Error ($\%RE_R$)
1943–2018					
C-14	0.000E+00	0.000E+00	0.000E+00	See note c	See note c
Cl-36	0.000E+00	0.000E+00	0.000E+00	See note c	See note c
H-3	4.371E+02	3.619E+01	4.003E+02	-5.998E-01	1.372E-01
I-129	0.000E+00	0.000E+00	0.000E+00	See note c	See note c
Np-237	0.000E+00	0.000E+00	0.000E+00	See note c	See note c
Re-187	0.000E+00	0.000E+00	0.000E+00	See note c	See note c
Sr-90	0.000E+00	0.000E+00	0.000E+00	See note c	See note c
Tc-99	0.000E+00	0.000E+00	0.000E+00	See note c	See note c

Table 6-3. Transient No-Decay Activity Balances for the SALDS Model Radionuclide Group 1 Simulations

Radionuclide	Released Radionuclide Inventory (I_R) ^a	Radionuclide Storage (S_R) ^{a,b}	Radionuclide Outflow (O_R) ^{a,b}	Error (E_R) ^a	Relative Error ($\%RE_R$)
1943–12070					
C-14	0.000E+00	0.000E+00	0.000E+00	See note c	See note c
Cl-36	0.000E+00	0.000E+00	0.000E+00	See note c	See note c
H-3	1.097E+03	0.000E+00	1.095E+03	-1.916E+00	1.746E-01
I-129	0.000E+00	0.000E+00	0.000E+00	See note c	See note c
Np-237	0.000E+00	0.000E+00	0.000E+00	See note c	See note c
Re-187	0.000E+00	0.000E+00	0.000E+00	See note c	See note c
Sr-90	0.000E+00	0.000E+00	0.000E+00	See note c	See note c
Tc-99	0.000E+00	0.000E+00	0.000E+00	See note c	See note c

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a. Units are in Curies.

b. Calculated by STOMP.

c. The radionuclide has no inventory.

$\%RE_R$ = percent relative error of the radionuclide activity balance

E_R = radionuclide activity balance error

I_R = radionuclide inventory entering the model domain from waste site releases

O_R = total radionuclide outflow from the model domain

S_R = radionuclide outflow from the model domain

SALDS = State-Approved Land Disposal Site

STOMP = Subsurface Transport Over Multiple Phases

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7 Results

This chapter presents the results of the transport simulations. These results include the calculation of cumulative radionuclide activity transferred to the groundwater and the cumulative activity remaining in the vadose zone at the end of the historical simulation (1943–2018) and the CA evaluation (i.e., forecast) simulation (2018–12070).

For each of the 16 radionuclides, Table 7-1 and Table 7-2 list the total activity discharged to the groundwater and the total activity remaining in the vadose zone. Table 7-1 shows these data at the end of the historical simulation (1943–2018), and Table 7-2 shows these data at the end of the forecast simulation (2018–12070).

The data presented in Table 7-1 and Table 7-2 are presented graphically in Section 7.1 through 7.3. These sections each present the data for one radionuclide. The cumulative activity of radionuclides discharged to the groundwater presented in Table 7-1 are shown spatially, aggregated by P2R grid cell, in Figure 7-1 and similar figures. The cumulative activity discharged to groundwater and the cumulative inventory released to the model shown in Table 7-1 for 1943–2018 and Table 7-2 for 2018–12070, is shown through time, first by figures which show the data from 1943–2018 (like Figure 7-3) and then by figures which show the data from 1943–12070 (like Figure 7-4). Additional figures showing radionuclide arrival to the groundwater through time for P2R grid cells in this model are shown in Appendix E.

Table 7-1. SALDS Model Radionuclide Activity Transfer to Groundwater from 1943–2018 and Remaining Activity in the Vadose Zone at 2018

Radionuclide	1943–2018 Inventory Released to Vadose Zone (Ci)	1943–2018 Activity Transferred to Groundwater (Ci)	1943–2018 Percent Activity Transferred to Groundwater ^a	Activity Remaining in Vadose Zone at 2018 (Ci)	Percent Activity Remaining in Vadose Zone at 2018 ^a
Radionuclide Group 1					
C-14	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Cl-36	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
H-3	4.370E+02	3.593E+02	82.2	2.122E+01	4.9
I-129	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Np-237	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Re-187	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Sr-90	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Tc-99	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Radionuclide Group 2					
U-232	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
U-233	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
U-234	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
U-235	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
U-236	0.000E+00	0.000E+00	See note b	0.000E+00	See note b

Table 7-1. SALDS Model Radionuclide Activity Transfer to Groundwater from 1943–2018 and Remaining Activity in the Vadose Zone at 2018

Radionuclide	1943–2018 Inventory Released to Vadose Zone (Ci)	1943–2018 Activity Transferred to Groundwater (Ci)	1943–2018 Percent Activity Transferred to Groundwater ^a	Activity Remaining in Vadose Zone at 2018 (Ci)	Percent Activity Remaining in Vadose Zone at 2018 ^a
U-238	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Th-230	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Ra-226	0.000E+00	0.000E+00	See note b	0.000E+00	See note b

a. The percentage or sum of percentages could differ slightly from 100 due to numerical error.

b. The radionuclide has no 1943–2018 inventory.

SALDS = State-Approved Land Disposal Site

Table 7-2. SALDS Model Radionuclide Activity Transfer to Groundwater from 2018–12070 and Remaining Activity in the Vadose Zone at 12070

Radionuclide	1943–12070 Inventory Released to Vadose Zone (Ci)	2018–12070 Activity Transferred to Groundwater (Ci)	2018–12070 Percent Activity Transferred to Groundwater ^a	Activity Remaining in Vadose Zone at 12070 (Ci)	Percent Activity Remaining in Vadose Zone at 12070 ^a
Radionuclide Group 1					
C-14	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Cl-36	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
H-3	1.097E+03	5.584E+02	50.9	0.000E+00	0.0
I-129	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Np-237	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Re-187	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Sr-90	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Tc-99	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Radionuclide Group 2					
U-232	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
U-233	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
U-234	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
U-235	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
U-236	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
U-238	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Th-230	0.000E+00	0.000E+00	See note b	0.000E+00	See note b
Ra-226	0.000E+00	0.000E+00	See note b	0.000E+00	See note b

Table 7-2. SALDS Model Radionuclide Activity Transfer to Groundwater from 2018–12070 and Remaining Activity in the Vadose Zone at 12070

Radionuclide	1943–12070 Inventory Released to Vadose Zone (Ci)	2018–12070 Activity Transferred to Groundwater (Ci)	2018–12070 Percent Activity Transferred to Groundwater ^a	Activity Remaining in Vadose Zone at 12070 (Ci)	Percent Activity Remaining in Vadose Zone at 12070 ^a
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a. The percentage or sum of percentages could differ slightly from 100 due to numerical error.

b. The radionuclide has no 1943–12070 inventory.

SALDS = State-Approved Land Disposal Site

Further description of the fate and transport of each radionuclide is outlined in Sections 7.1 through 7.16. Results presented in the sections show cumulative activity of the radionuclide discharged to groundwater over the historical (1943–2018) and forecast (2018–12070) simulations, and figures showing the cumulative activity released from the sources compared to the transfer rate to groundwater for the historical (1943–2018) and entire (1943–12070) modeled periods.

7.1 C-14 Fate and Transport Results

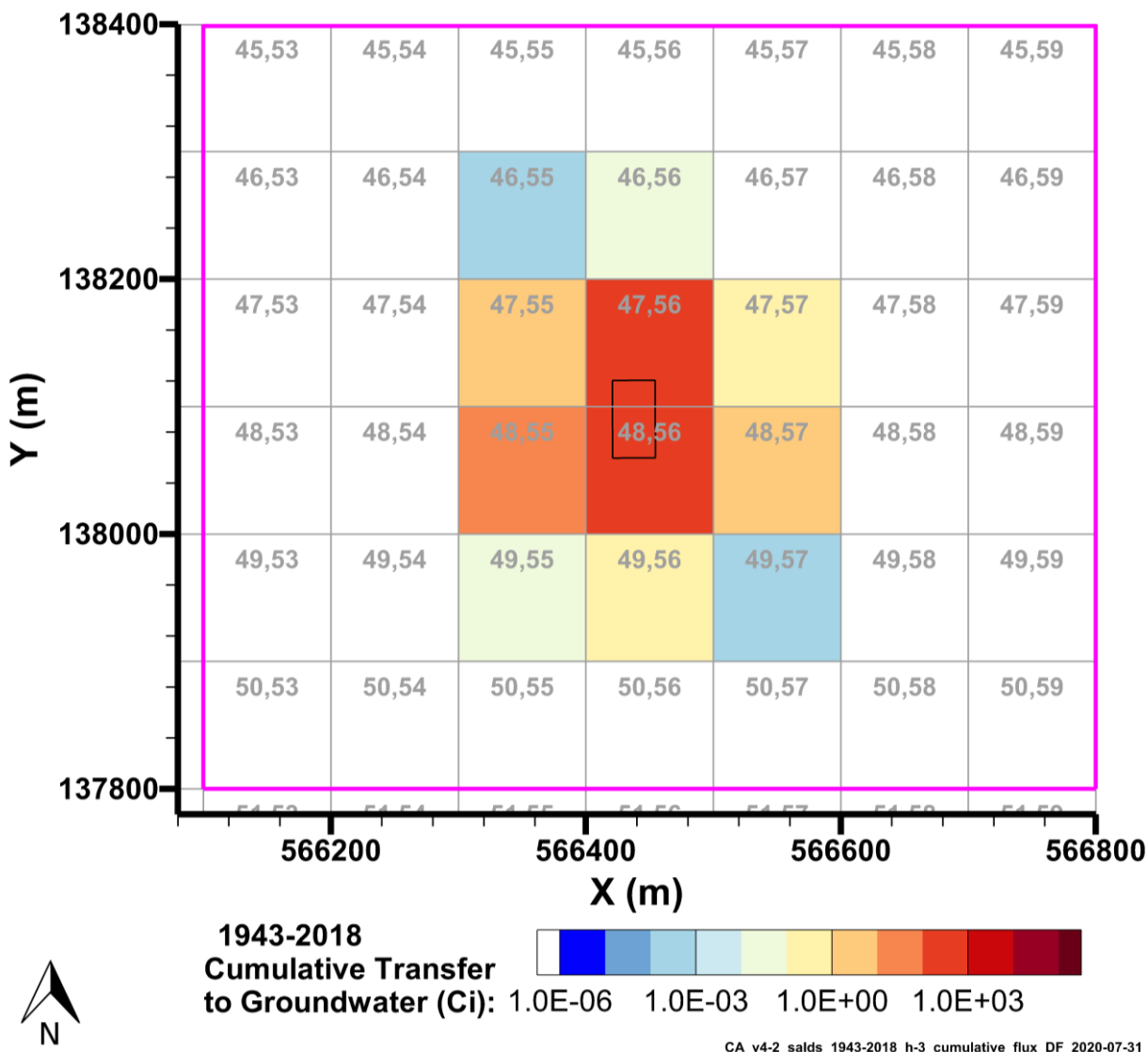
Due to a lack of inventory, transport of C-14 was not calculated in this model.

7.2 Cl-36 Fate and Transport Results

Due to a lack of inventory, transport of Cl-36 was not calculated in this model.

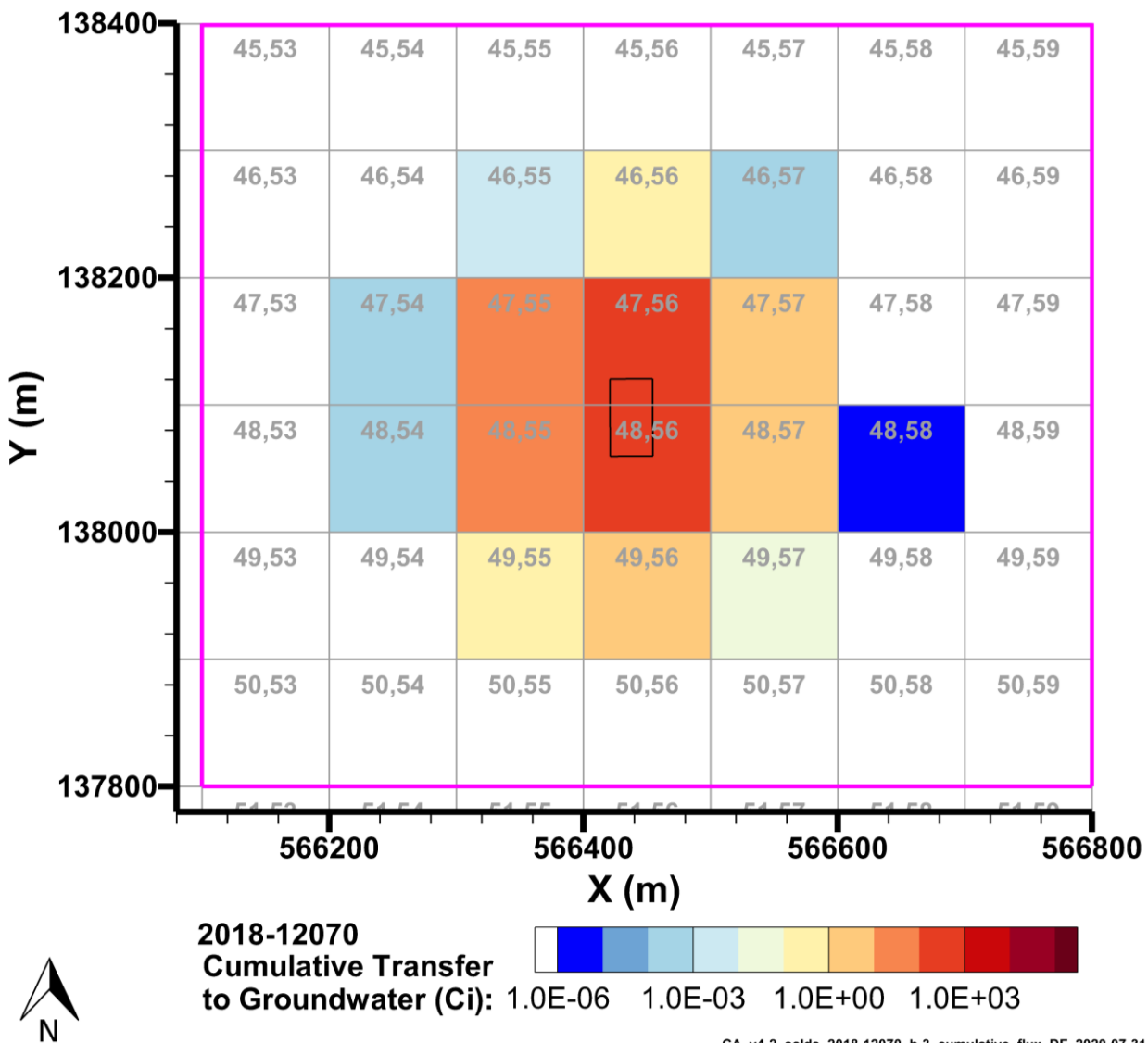
7.3 H-3 Fate and Transport Results

This model simulated release and transport of H-3. The cumulative release of H-3 into groundwater is shown aggregated by P2R grid cell in Figure 7-1 and Figure 7-2 for 1943–2018 and 2018–12070, respectively. The inventory released to the SALDS model and the transfer of H-3 to groundwater are shown from 1943–2018 in Figure 7-3 and from 1943–12070 in Figure 7-4. Figure 7-5 through Figure 7-11 show the flux of H-3 to groundwater in Ci/yr. These figures are generated at times with peak fluxes (local maxima) and during periods with gradual decline, as shown in Figure 7-3 and Figure 7-4. A figure for 2018, Figure 7-7, is also included to demonstrate the initial flux conditions for the 2018–12070 simulation.



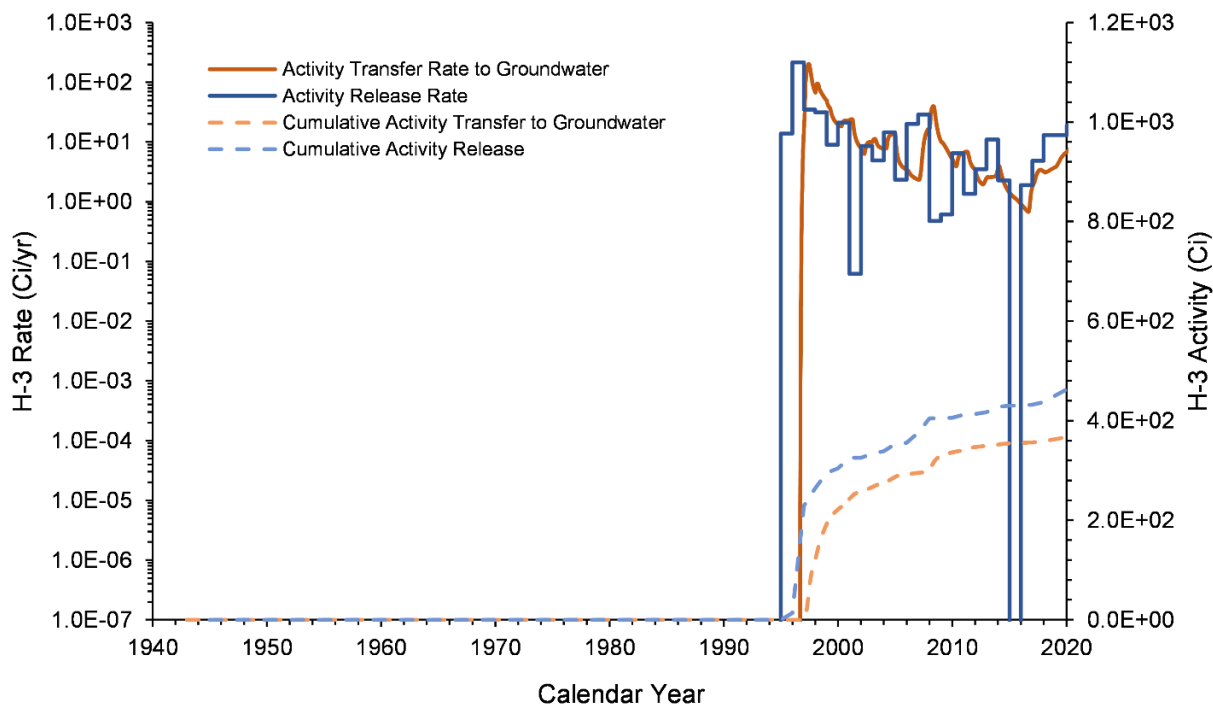
Note: source zone outlined in pink.

**Figure 7-1. Cumulative H-3 Activity Discharged to Groundwater
from the SALDS Model from 1943–2018 per P2R Grid Cell**



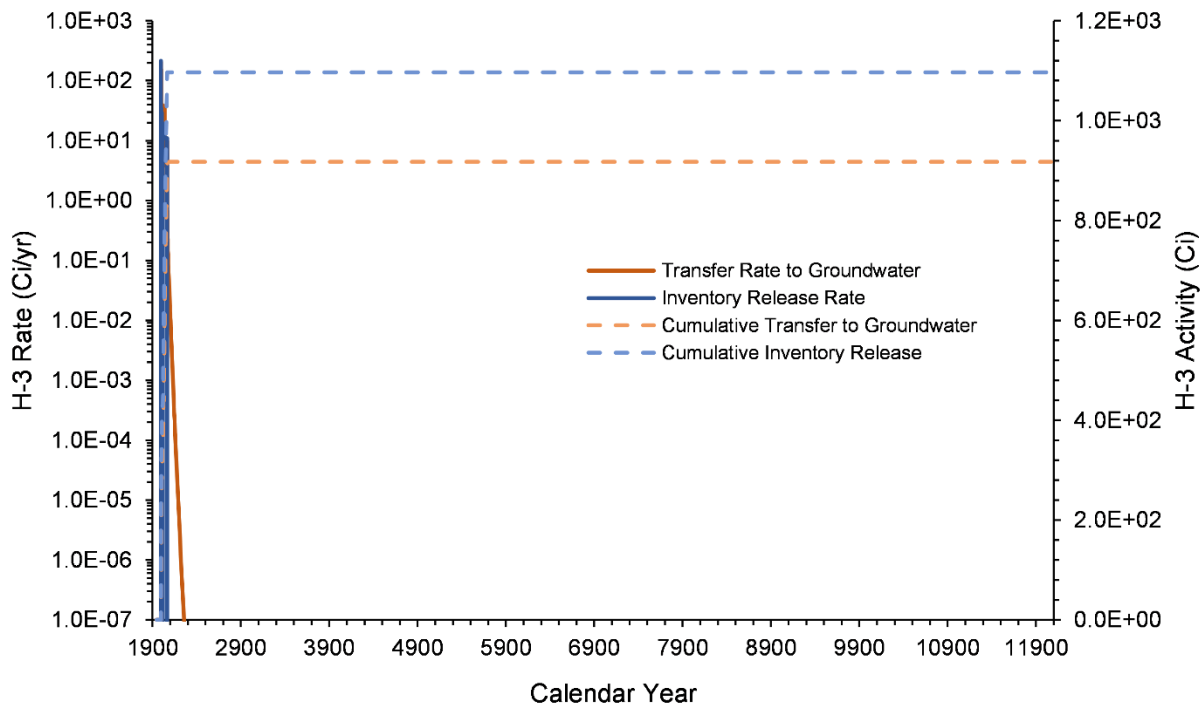
Note: source zone outlined in pink.

**Figure 7-2. Cumulative H-3 Activity Discharged to Groundwater
from the SALDS Model from 2018–2070 per P2R Grid Cell**



CA_v4-2_salds_H-3_1943-2018_rate_and_cumulative_v_time_PA_2020-07-07

Figure 7-3. H-3 Inventory Release from Waste Sites and Transfer to Groundwater for the SALDS Model from 1943–2018



CA_v4-2_salds_H-3_1943-12070_rate_and_cumulative_v_time_PA_2020-07-07

Figure 7-4. H-3 Inventory Release from Waste Sites and Transfer to Groundwater for the SALDS Model from 1943–12070

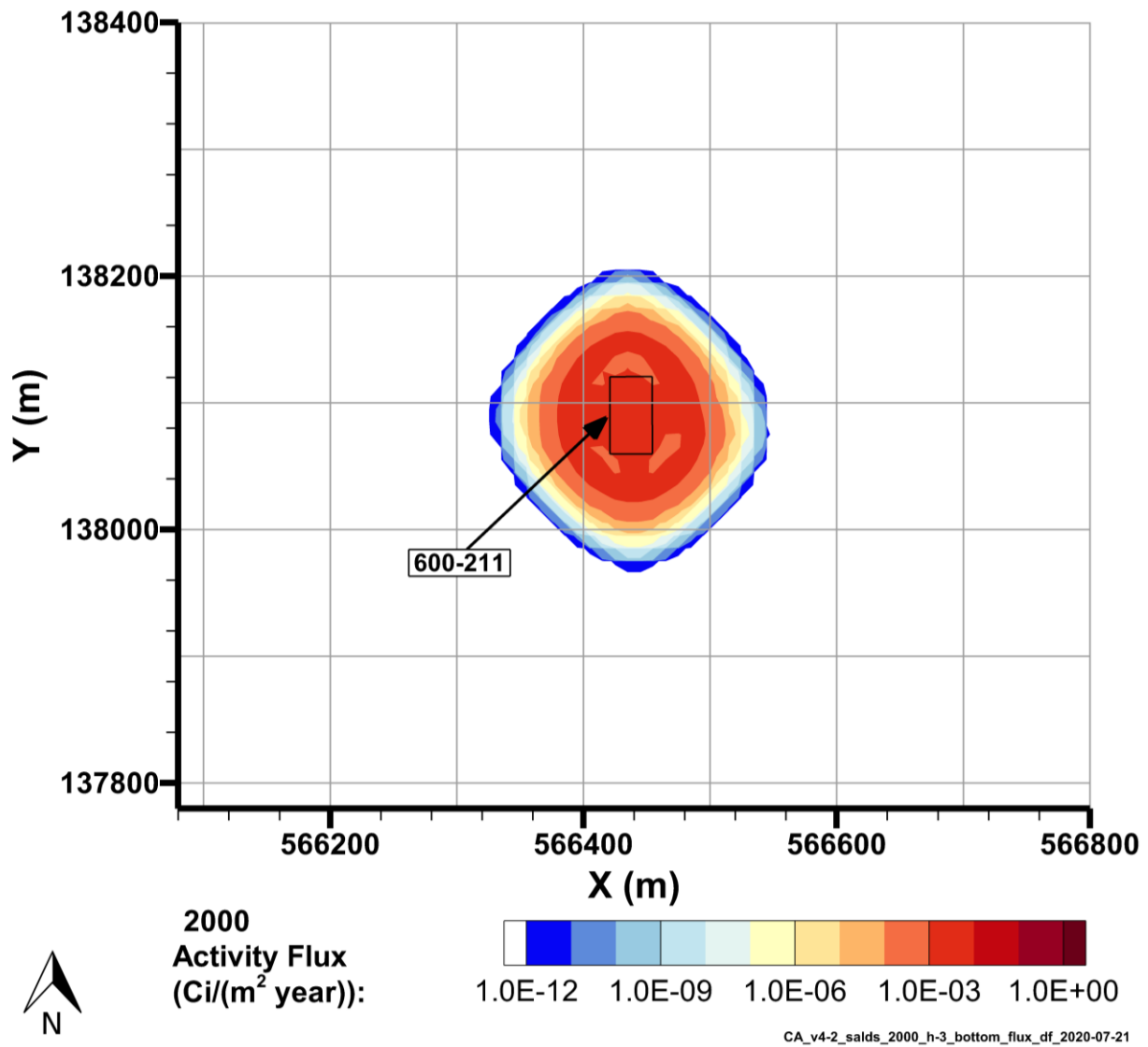


Figure 7-5. H-3 Flux to Groundwater, 2000

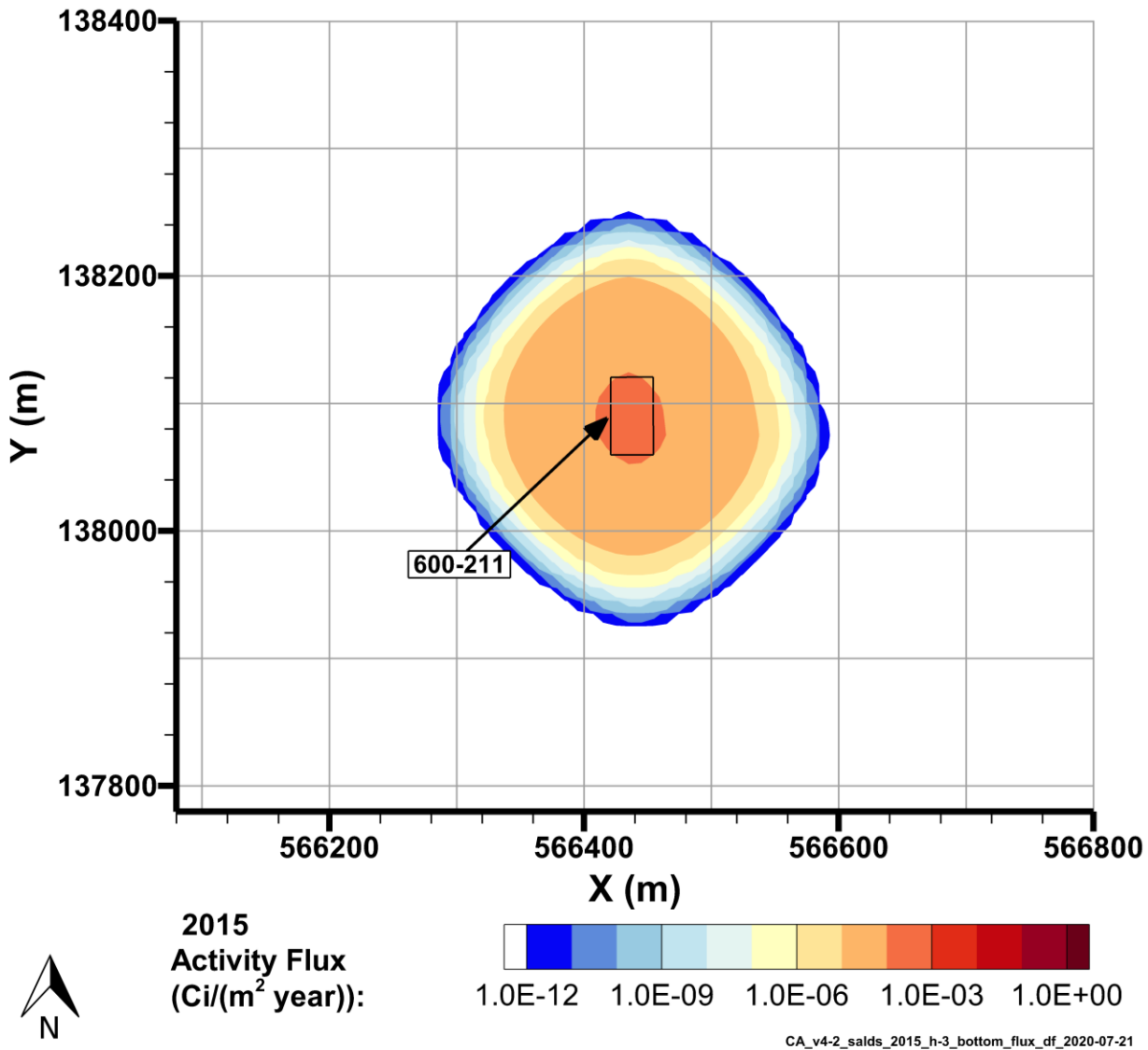


Figure 7-6. H-3 Flux to Groundwater, 2015

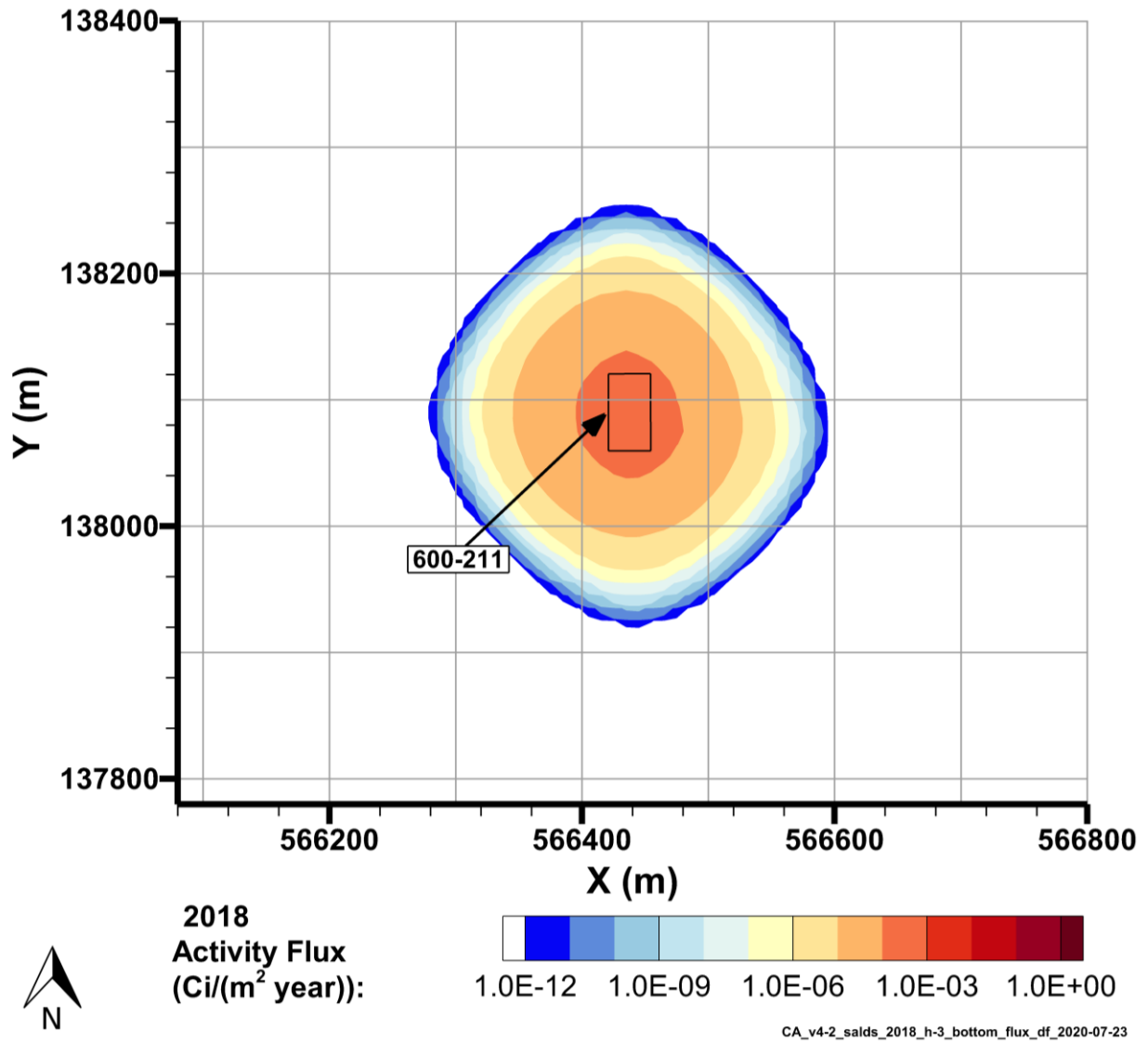


Figure 7-7. H-3 Flux to Groundwater, 2018

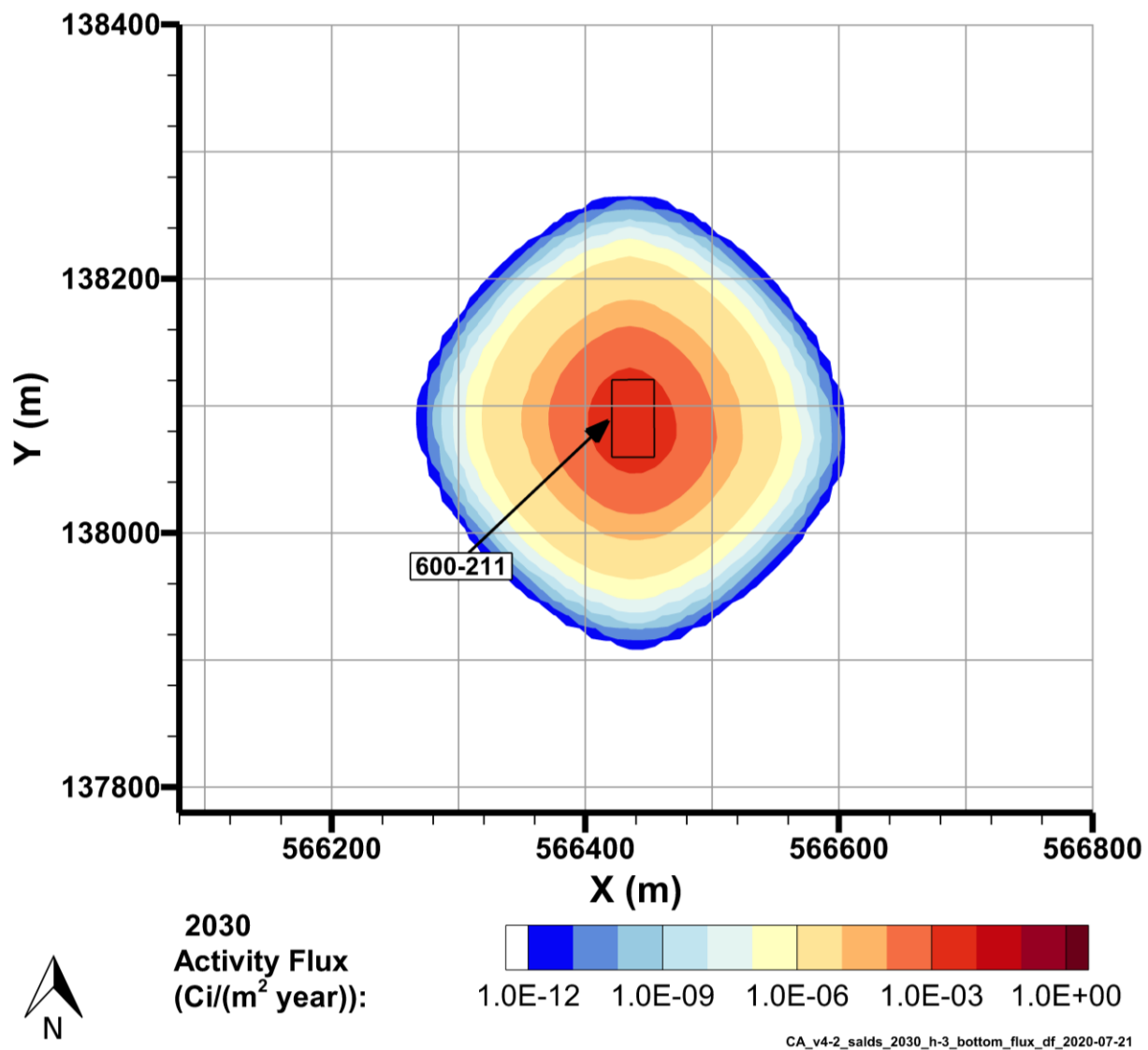


Figure 7-8. H-3 Flux to Groundwater, 2030

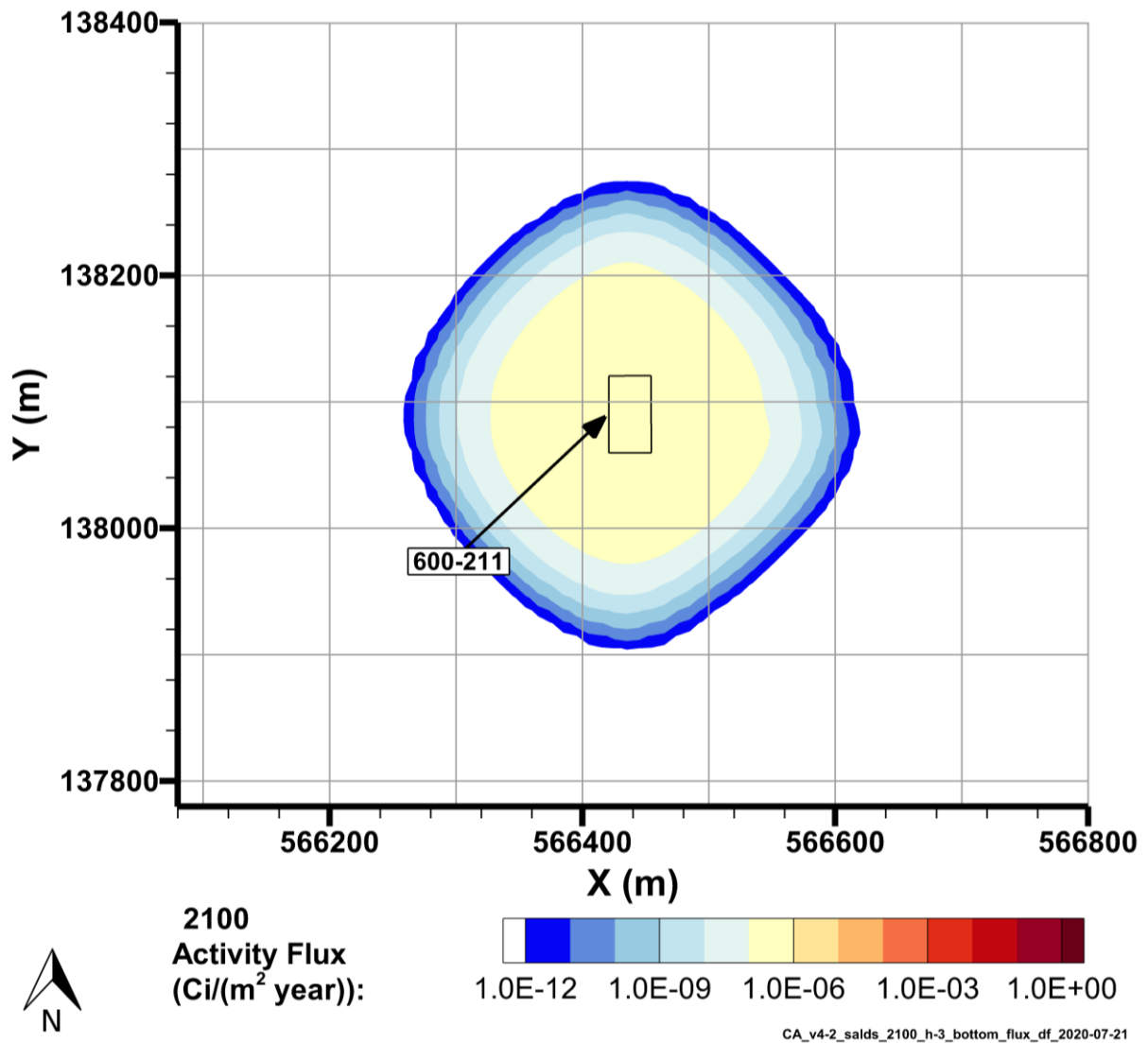


Figure 7-9. H-3 Flux to Groundwater, 2100

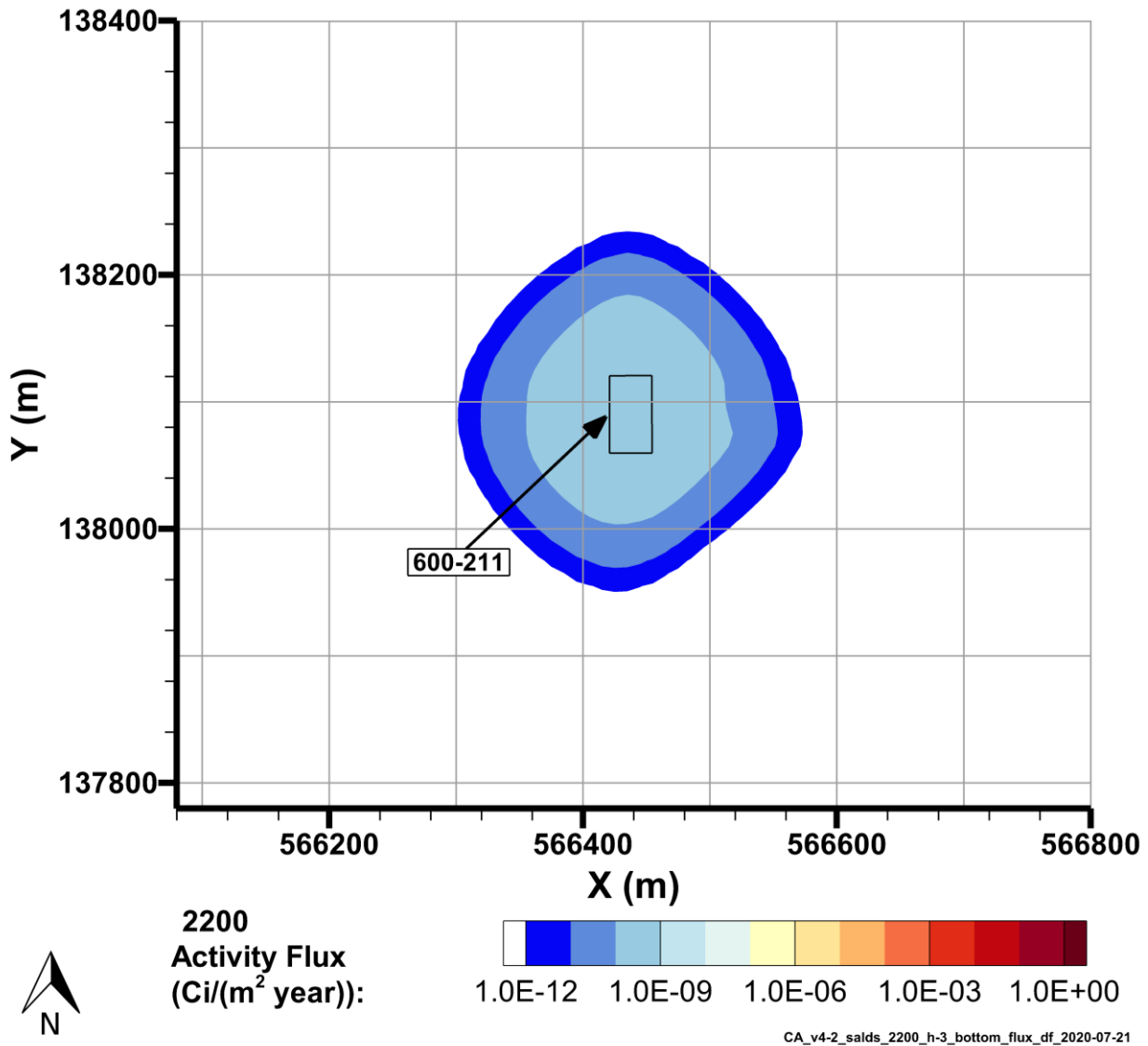


Figure 7-10. H-3 Flux to Groundwater, 2200

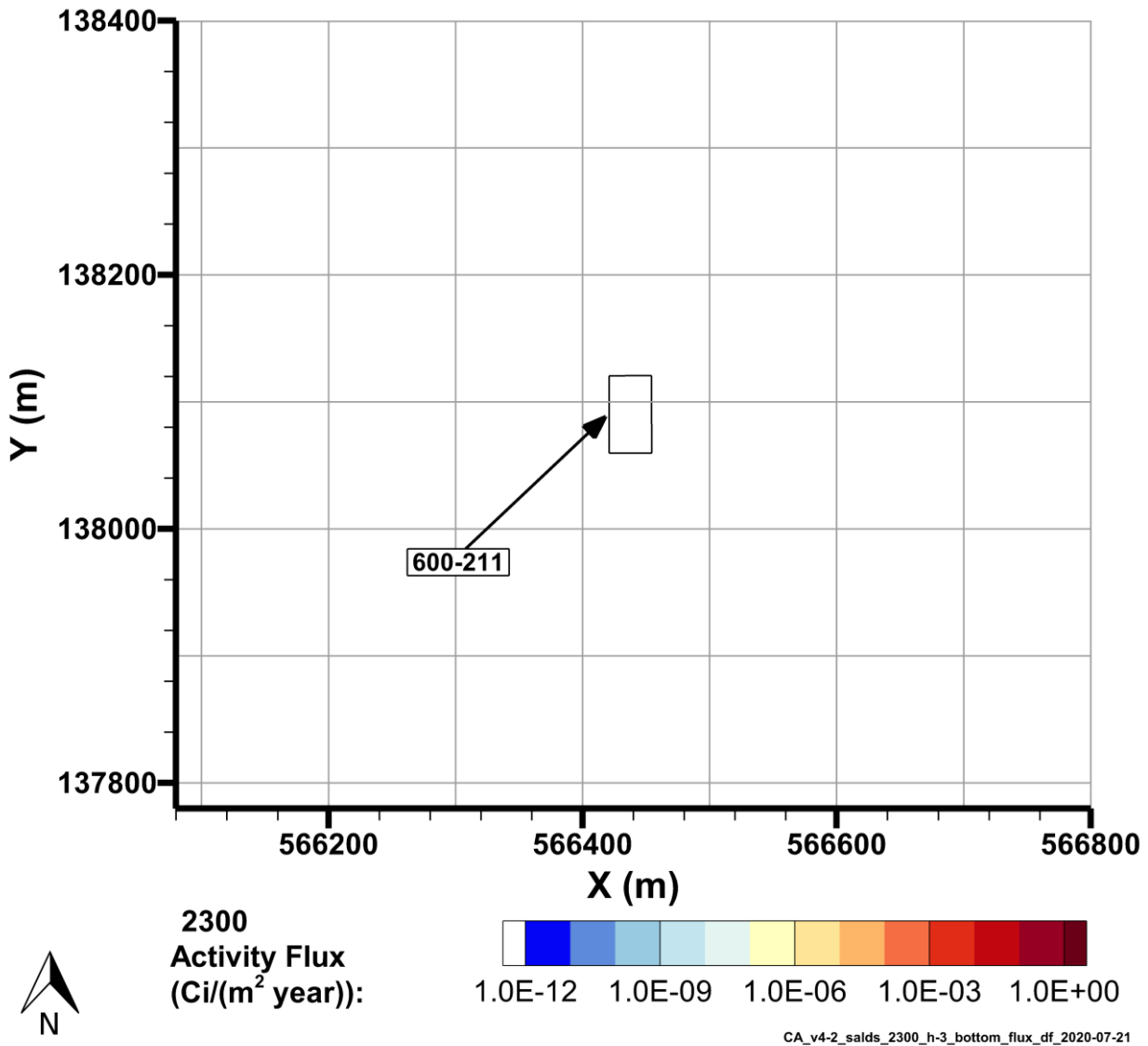


Figure 7-11. H-3 Flux to Groundwater, 2300

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Appendix A

Checking Documentation for the State-Approved Land Disposal Site Model

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Contents

A1 Introduction..... A-1


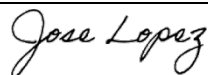
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A1 Introduction

This appendix contains documentation of checks completed by the modeling team and from qualified employees outside of the modeling team.


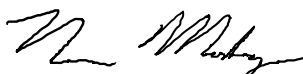
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Peer Reviewer Name:	Jose Lopez			
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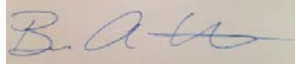
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Peer Reviewer Name:	Jose Lopez			
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Date Completed	Modeler: 03-23-2020		Peer Reviewer: 3-27-2020	
Name	Praveena Allena		Jose Lopez	
Signature				

Model Check 2 – Transport XPRT Part B				
Model (full name):	SALDS			
Modeler Name:	Praveena Allena			
Peer Reviewer Name:	Neira Mondragon			
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Peer Reviewer Name:	Neira Mondragon			
Task/Action/Operation	Modeler		Peer Reviewer	
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Peer Reviewer Name:	Neira Mondragon			
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Date Completed	05/07/20		05/08/20	
Name	Praveena Allena		Neira Mondragon	
Signature and Date	 05/08/2020			

Model Check 3 – Transport XPRT Part C				
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Modeler Name:	Praveena Allena			
Peer Reviewer Name:	Brian Archuleta			
Task/Action/Operation	Modeler		Peer Reviewer	
	Status	Comment	Status	Comment
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Completed Simulation Title Card Check (Page 21)	<input checked="" type="checkbox"/>		<input checked="" type="checkbox"/>	
Completed Solution Control Card Check (Page 22-24)	<input checked="" type="checkbox"/>		<input checked="" type="checkbox"/>	
Completed Direct <i>input_XPRT-1</i> Copy Check (Page 25)	<input checked="" type="checkbox"/>		<input checked="" type="checkbox"/>	
Completed Solute/Fluid Interaction Card Check (Page 26)	<input checked="" type="checkbox"/>		<input checked="" type="checkbox"/>	
Completed Output Control Card Check (Page 27)	<input checked="" type="checkbox"/>		<input checked="" type="checkbox"/>	
Completed Surface Card Check (Page 28)	<input checked="" type="checkbox"/>		<input checked="" type="checkbox"/>	
Input File Check – MB2 simulation				
Completed Simulation Title Card Check (Page 31)	<input type="checkbox"/>	NA	<input type="checkbox"/>	N/A Does not have a mb2 file.
Completed Solution Control Card Check (Page 32-234)	<input type="checkbox"/>	NA	<input type="checkbox"/>	N/A Does not have a mb2 file.

Model Check 3 – Transport XPRT Part C				
Model (full name):	SALDS			
Modeler Name:	Praveena Allena			
Peer Reviewer Name:	Brian Archuleta			
Task/Action/Operation	Modeler		Peer Reviewer	
	Status	Comment	Status	Comment
Completed Direct <i>input_XPRT-1</i> Copy Check (Page 35)	<input type="checkbox"/>	NA	<input type="checkbox"/>	N/A Does not have a mb2 file.
Completed Solute/Fluid Interaction Card Check (Page 36)	<input type="checkbox"/>	NA	<input type="checkbox"/>	N/A Does not have a mb2 file.
Completed Output Control Card Check (Page 37)	<input type="checkbox"/>	NA	<input type="checkbox"/>	N/A Does not have a mb2 file.
Completed Surface Card Check (Page 38)	<input type="checkbox"/>	NA	<input type="checkbox"/>	N/A Does not have a mb2 file.
After completion by both the modeler and peer-reviewer, the form should be moved to the CompletedForms folder. The form should not be signed until both have completed the check and all issues have been resolved.				
	Modeler		Peer Reviewer	
Date Completed	05/06/2020		05-12-2020	
Name	Praveena Allena		Brian Archuleta	
Signature	 05-12-2020			

Model Check 4– Transport XPRT Part D				
Model (full name):	SALDS			
Modeler Name:	Praveena Allena			
Peer Reviewer Name:	Andrew Murphy			
Task/Action/Operation	Modeler		Peer Reviewer	
	Status	Comment	Status	Comment
<p>Check list follows sections in CA-XPRT-12070-Input-File-Check-PartD-*.pptx</p> <p>Modelers: \CAVE\v4-2\supportfiles\CheckingDocs\xprt-PartD</p> <p>Peer Reviewers: \Rel.061\vadose\Peer-Checking-xprt-D\CheckingDocs</p>				
Completed tool qualification checks (pages 11-12 of CA-XPRT-12070-Input-File-Check-PartD-*.pptx)	<input checked="" type="checkbox"/>		<input checked="" type="checkbox"/>	
Completed <i>xprt_12070_input_gen.f</i> tool input check (Pages 14-15)	<input checked="" type="checkbox"/>		<input checked="" type="checkbox"/>	
Input File Check: xpirt-1-12070 simulation				
Completed Simulation Title Card Check (Page 18)	<input checked="" type="checkbox"/>		<input checked="" type="checkbox"/>	
Completed Solution Control Card Check (Page 19)	<input checked="" type="checkbox"/>		<input checked="" type="checkbox"/>	
Completed Direct <i>input_XPRT-1</i> Copy Check (Page 20)	<input checked="" type="checkbox"/>		<input checked="" type="checkbox"/>	
Completed Output Control Card Check (Page 21)	<input checked="" type="checkbox"/>		<input checked="" type="checkbox"/>	
Input File Check: xpirt-2-12070 simulation				
Completed Simulation Title Card Check (Page 24)	<input type="checkbox"/>	NA	<input type="checkbox"/>	NA
Completed Solution Control Card Check (Page 25)	<input type="checkbox"/>	NA	<input type="checkbox"/>	NA
Completed Direct <i>input_XPRT-1</i> Copy Check (Page 26)	<input type="checkbox"/>	NA	<input type="checkbox"/>	NA
Completed Output Control Card Check (Page 27)	<input type="checkbox"/>	NA	<input type="checkbox"/>	NA

Model Check 4– Transport XPRT Part D				
Model (full name):	SALDS			
Modeler Name:	Praveena Allena			
Peer Reviewer Name:	Andrew Murphy			
Task/Action/Operation	Modeler		Peer Reviewer	
	Status	Comment	Status	Comment
After completion by both the modeler and peer-reviewer, the form should be moved to the CompletedForms folder. The form should not be signed until both have completed the check and all issues have been resolved.				
	Modeler		Peer Reviewer	
Date Completed	05-09-2020		5-15-2020	
Name	Praveena Allena		Andrew Murphy	
Signature and Date	 05-15-2020		 5-15-2020	

Appendix B

Cross-Sections of the Hydrostratigraphy in the State- Approved Land Disposal Site Model

(Electronic Appendix)

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Contents

B1 Introduction..... B-1

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B1 Introduction

This appendix is a folder containing two subfolders, SouthToNorth and WestToEast. Both contain images of cross-sections through the model showcasing the hydrostratigraphy; the first from south to north and the second from west to east.

The contents of this electronic appendix are stored in the Electronic Modeling Management Archive (EMMA) indexed to this ECF by document number.

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Appendix C

Charts of Recharge to the State-Approved Land Disposal Site Model as Defined by the Recharge Evolution Tool

(Electronic Appendix)

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Contents

C1 Introduction.....C-1

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C1 Introduction

This appendix is a folder of images. Each image is a map of the annual recharge rate at the surface of the model, as assigned by the Recharge Evolution Tool, per grid cell in the model for each year where any recharge rate is different than the preceding year.

The contents of this electronic appendix are stored in the Electronic Modeling Management Archive (EMMA) indexed to this ECF by document number.

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Appendix D

Software Installation and Checkout Forms

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Contents

D1 Introduction.....D-1

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D1 Introduction

This appendix is a portable document file showing the completed Software Installation and Checkout form.

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CHPRC SOFTWARE INSTALLATION AND CHECKOUT FORM**Software Owner Instructions:**

Complete Fields 1-13, then run test cases in Field 14. Compare test case results listed in Field 15 to corresponding Test Report outputs. If results are the same, sign and date Field 19. If not, resolve differences and repeat above steps.

Software Subject Matter Expert Instructions:

Assign test personnel. Approve the installation of the code by signing and dating Field 21, then maintain form as part of the software support documentation.

GENERAL INFORMATION:

1. Software Name: STOMP (Subsurface Transport Over Multiple Phases) Software Version No.: Bld 6

EXECUTABLE INFORMATION:

2. Executable Name (include path):

Following STOMP serial and parallel mode executable files in directory [REDACTED]/bin on head node and each compute node (compute-0-0 through compute-0-8, inclusive):

MD5 File Signature	Executable File Name
4a0f738b74620bc8df4d05290b513a44	eSTOMP1-chprc06-20200204-gaia.x
6536b8e12d8c5b83dca76f2c947b6153	stomp-wae-bcg-chprc06i.x
e0cdf04bc1a2f6c55c5a1b499939f663	stomp-wae-bcg-chprc06l.x
86c58db6fac5d1b4e6cbe13041b2568b	stomp-wae-bcg-chprc07i.x
6e72340bb39f6056e232fe5ff241c4d4	stomp-wae-bd-chprc06i.x
3f837a0fb8d9f47dbcada686f542d7fc	stomp-wae-bd-chprc06l.x
7e5b4cc36a8991b3d5a8ea2ed155ce47	stomp-wae-cgsq-chprc06i.x
00a898c0c3ec06817485781ad1c9ec46	stomp-wae-cgsq-chprc06l.x
f18ff5ab5667065d8ab12657344fb6a0	stomp-wae-cgst-chprc06i.x
061af86cf21ad8435b046d0efabe971b	stomp-wae-cgst-chprc06l.x
3c8111a9855dc0e430bf3c8a7abcf37e	stomp-w-bcg-chprc06i.x
20436d615a94955a2ce8eecd8b8cba546	stomp-w-bcg-chprc06l.x
8b3df29df21d040189c3e2a50ef823bb	stomp-w-bd-chprc06i.x
066a289a75aedb933eb2536da5d7d1ff	stomp-w-bd-chprc06l.x
c8e62ad7a0d9b6fca39d8a8952ef5d8e	stomp-w-cgsq-chprc06i.x
28ad16806e1307aca51fd7bf89793e75	stomp-w-cgsq-chprc06l.x
6c25051016db2fe1f883a7caaaab1e97	stomp-w-cgst-chprc06i.x
ff9ff6f29b3469419ffaece87d7e772b	stomp-w-cgst-chprc06l.x
0c3e3fba40f5b93e71bcf9586432fd27	stomp-w-r-bcg-chprc06i.x
78492aee80a8c2d0a4e82aabf4a9c213	stomp-w-r-bcg-chprc06l.x
84b129786aba9c4be884e15e45a67389	stomp-w-r-bd-chprc06i.x
e990f1566c8099a8d54508de3da9cd88	stomp-w-r-bd-chprc06l.x
18a589a2b55aab2db290efea19b39351	stomp-w-r-cgsq-chprc06i.x
6569959476772a137df35ce874821889	stomp-w-r-cgsq-chprc06l.x

3. Executable Size (bytes): MD5 signatures above uniquely identify each executable file

COMPILATION INFORMATION:

4. Hardware System (i.e., property number or ID):

Tellus Subsurface Modeling Platform (serial STOMP executables) and compiled directly on Gaia for eSTOMP.

5. Operating System (include version number):

INSTALLATION AND CHECKOUT INFORMATION:

6. Hardware System (i.e., property number or ID):

GAIA Subsurface Flow and Transport Modeling Platform (Linux Cluster)

A-6005-149 (REV 0)

Date

Appendix E

Radionuclide Arrival to the Groundwater Through Time for Plateau to River Grid Cells in the State-Approved Land Disposal Site Model

(Electronic Appendix)

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Contents

E1 Introduction..... E-1

E1 Introduction

This appendix is a folder of portable document files. These files contain charts showing the radionuclide transfer to groundwater from the model in different configurations, as indicated by the figure titles on the charts.

The contents of this electronic appendix are stored in the Electronic Modeling Management Archive (EMMA) indexed to this ECF by document number.

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Appendix F

Title Pages for Cited Electronic Data Modeling Transmittals

(Electronic Appendix)

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Contents

F1. Introduction..... 1


F1. Introduction


This appendix contains the cover sheets for the electronic model data transmittals cited in this environmental calculation file. The electronic model data transmittals cover sheets presented in this appendix are EMDT-IN-0047¹, *SALDS Liquid Disposal Volumes and Tritium Inventory*, Rev. 0, and EMDT-GR-0035², *Waste Site and Structure Footprint Shapefiles for Inclusion in Updated Composite Analysis*, Rev. 0.


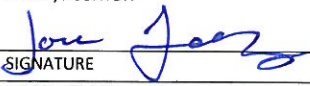
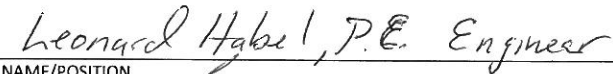

¹ EMDT-IN-0047, *SALDS Liquid Disposal Volumes and Tritium Inventory*, Rev. 0, CH2M HILL Plateau Remediation Company, Richland, Washington.

² EMDT-GR-0035, *Waste Site and Structure Footprint Shapefiles for Inclusion in Updated Composite Analysis*, Rev. 0, CH2M HILL Plateau Remediation Company, Richland, Washington.

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 Environmental Modeling Data Transmittal Cover Page	
No.: EMDT-GR-0035 <i>[Request EMDT number from Modeling Team Leader]</i>	Revision No.: 0
Title: Waste Site and Structure Footprint Shapefiles for Inclusion in Updated Composite Analysis	Date: 06/24/2019
1. Data Description <i>Provide the description of data set or data type.</i> <p>Ehsit is a shapefile of known or suspected waste sites across the Hanford site (3,390 features in this version). Bggenexs is a shapefile of existing buildings/structures across the Hanford site (2,443 features in this version).</p>	
2. Data Intended Use <i>Identify the data's intended use. Describe the rationale for its selection and how the data will be incorporated into a model, report, or database. Include discussion of the extent to which the data demonstrate the properties of interest.</i> <p>These shapefiles provide the footprints to identify features commonly modeled/reported. They identify the location of where these features are on the Hanford site and the extent of their domains.</p>	
3. Data Sources <i>List databases, documents, etc. – provide sufficient detail to enable data to be located by independent reviewer</i> <p>These were obtained as part of the data transfer to create the 2017 HIGRV. These files were originally sent as a feature dataset within an ArcGIS geodatabase by Margo Aye at Jacobs, to Jose Lopez at INTERA via email on 7/26/2018.</p> <p>The original geodatabase and shapefiles can be found at:</p> <p>S:\PSC\CHPRC.C003.HANOFF\Rel.044\HIGRV2017\Data\MargoAye@Jacobs</p>	
4. Impact of Use or Nonuse of Data <i>Describe the importance of the data to the model, report, and/or conclusions which they support. Identify the value added and discuss the impacts of not using the data.</i> <p>This dataset has supported, and still supports, a variety of Hanford projects. These can be used as visual aids by generating figures for reports, presentations, or for discussions. Attributes, such as inventory, are also mapped to these features to evaluate their impact. Excluding this dataset would impact a project's ability to identify a site spatially with a reliable source.</p>	
5. Prior Uses <i>Identify the data's prior uses. Describe whether the data have been used in similar applications by the scientific or regulatory community. Include the associated verification processes and prior reviews and review results.</i> <p>Ehsit and bggenexs have been used to support the Hanford Groundwater Annual Reports. Figures in the report incorporate these datasets. The Hanford Interactive Groundwater Viewer (HIGRV) of the annual report also use these datasets.</p>	

	<h2 style="margin: 0;">Environmental Modeling Data Transmittal Cover Page</h2>
No.: EMDT-GR-0035 <i>[Request EMDT number from Modeling Team Leader]</i>	Revision No.: 0
Title: Waste Site and Structure Footprint Shapefiles for Inclusion in Updated Composite Analysis	Date: 06/24/2019
<p>6. Data Acquisition Method(s)</p> <p><i>Describe the data acquisition method and associated QA/QC, considering the following:</i></p> <ol style="list-style-type: none"> a. <i>Qualifications of personnel or organizations generating the data;</i> b. <i>Technical adequacy of equipment and procedures used;</i> c. <i>Environmental and programmatic conditions if germane to the data quality;</i> d. <i>The extent to which acquisition processes reflect modeling requirements;</i> e. <i>The quality and reliability of the measurement control program;</i> f. <i>The degree to which independent audits of the process were conducted;</i> g. <i>Extent and reliability of the associated documentation.</i> <p><i>For databases, identify query language used to obtain data from database (SQL, etc.), briefly describe the query description and attach copy</i></p> <p>As mentioned in section 3, these files were given to INTERA by Margo Aye. Margo Aye is the GIS Lead Soil and Ground Water at Jacobs. Margo retrieved this data from the Mission Support Alliance (MSA) Central Mapping Services server. Ehsit was retrieved on 12/14/2017 and bggenexs on 12/17/2017.</p>	
<p>7. Corroborating Data</p> <p><i>Identify and discuss any corroborating datasets. Provide any documentation that confirms the corroborating data substantiate existing parameter values, distributions, or data quality.</i></p> <p>Not applicable.</p>	
<p>8. Data Quality Considerations</p> <p><i>Discuss data quality considerations not identified in other sections. Include discussion of data quality indicators (i.e., accuracy, precision, representativeness, completeness, and comparability).</i></p> <p>Waste site (and structure) data are compiled using a variety of methods including translations from annotated field maps, estimates based on published reports, and digitizing from aerial photography/scanned drawings/global positioning surveys. Mapped location is based on the best available information at the time. As new data becomes available, mapped location is modified to account for newly identified information.</p>	
<p>9. Assumptions and Limitations on Data Use</p> <p><i>Document known uncertainties, assumptions, constraints or limits on data.</i></p> <p>Due to the explanation in section 8, there may be a level of uncertainty behind this dataset. None of the mapped locations are absolute. Features may have changed/removed/added throughout different iterations of this dataset.</p>	

 Environmental Modeling Data Transmittal Cover Page	
No.: EMDT-GR-0035 <i>[Request EMDT number from Modeling Team Leader]</i>	Revision No.: 0
Title: Waste Site and Structure Footprint Shapefiles for Inclusion in Updated Composite Analysis	Date: 06/24/2019
Data Configuration Item Submittal:	
Data Provider Submittal	Jose Lopez/GIS Analyst NAME/POSITION  SIGNATURE
	G-24-19 DATE
Data Configuration Item Review and Verification:	
10. Verification Process Describe steps taken to verify that these data are appropriate for intended use, noting any limitations <i>I reviewed this document and the data provided by Margo Aye on July 26, 2018. The information stated herein is accurate.</i>	
11. Summary of Data Review The review shall ensure that the report meets the listed criteria. Consideration includes ensuring that the data collection method employed was appropriate for the type of data being considered and confidence in the data acquisition and subsequent processing methodology is warranted.	
Is documentation technically adequate, complete, and correct? <input checked="" type="checkbox"/> Yes [] No	
Are uncertainties and limitations on appropriate use of data discussed? <input checked="" type="checkbox"/> Yes [] No	
Are the assumptions, constraints, bounds, or limits on the data identified? <input checked="" type="checkbox"/> Yes [] No	
Data Reviewer Approval	Approval of Data Configuration Item  NAME/POSITION  SIGNATURE
	6/24/2019 DATE

**EMDT accepted for Composite Analysis input in
Data Readiness Review on 12/2/2019.**

ENVIRONMENTAL MODELING DATA TRANSMITTAL COVER PAGE

No.: EMDT-IN-0047

Revision No: 0

(Request EMDT number for Modeling Team Leader)

Title: SALDS Liquid Disposal Volumes and Tritium Inventory

Date: 3/12/2020

1. Data Description

Provide the description of data set or data type.

Water and tritium releases to the State-Approved Land Disposal Site (SALDS) from the start of operations in December 1995 through September 2017, and the estimated future water and tritium releases for October 2017 through the end of operations in Year 2065.

2. Data Intended Use

Identify the data's intended use. Describe the rationale for its selection and how the data will be incorporated into a model, report, or database. Include discussion of the extent to which the data demonstrates the properties of interest.

The data will be used as input to simulations of tritium migration through the vadose zone. SALDS receives treated effluent from the Effluent Treatment Facility (ETF) operated by Washington River Protection Solutions (WRPS), and WRPS provided the data of past water and tritium releases to SALDS and estimated future water and tritium releases. These data will be incorporated into a flow and transport model of the vadose zone beneath SALDS using the Subsurface Transport over Multiple Phases (STOMP) model code.

3. Data Sources

List databases, documents, etc. - provide sufficient detail to enable data to be located by independent reviewer.

RPP-CALC-61876, 2017, Estimated Tritium Discharges to the State Approved Land Disposal Site for Use in Groundwater Modeling, Rev. 0, Washington River Protection Solutions, Richland, Washington. Contains the estimated future water and tritium releases to the SALDS starting in October 2017 and includes the methodology for estimating the releases.

RPP-CALC-61950, 2018, Fate and Transport Analysis of Historical and Future Tritium Releases from the State Approved Land Disposal Site, FY 2018, Rev. 0, Washington River Protection Solutions, Richland, Washington. Table A-1 in Appendix A contains the historical water and tritium releases from SALDS for December 1995 through September 2017. This appendix also lists the estimated future releases for October 2017 through Year 2065.

4. Impact of Use or Nonuse of Data

Describe the importance of the data to the model, report, and/or conclusions which they support. Identify the value added and discuss the impacts of not using the data.

The purpose of the model simulations is to evaluate future migration and fate of tritium from the SALDS as part of the Composite Analysis (CA) and Cumulative Impacts Evaluation (CIE) activities. Historical releases of water and tritium from SALDS and estimates of future water and tritium releases are necessary inputs to the modeling.

5. Prior Use

Identify the data's prior uses. Describe whether the data have been used in similar applications by the scientific or regulatory community. Include the associated verification processes and prior reviews and review results.

ENVIRONMENTAL MODELING DATA TRANSMITTAL COVER PAGE (Continued)**No.:** EMDT-IN-0047**Revision No:** 0*(Request EMDT number for Modeling Team Leader)***Title:** SALDS Liquid Disposal Volumes and Tritium Inventory**Date:** 3/12/2020**5. Prior Use**

These data were acquired and used for groundwater model simulations of tritium migration and fate from SALDS to meet requirements of the SALDS disposal permit (ST0004500). This work was performed in FY 2018 and is documented in RPP-CALC-61950. Historical releases from SALDS have been used for numerous Hanford Site model applications. For example, the data are used annually in groundwater modeling to evaluate operation of the 200 West Pump-and-Treat system (e.g., ECF-HANFORD-19-0014). These applications are reviewed by a checker and senior reviewer as part of the modeling process.

ECF-HANFORD-19-0014, 2019, Description of Groundwater Calculations and Assessments for the Calendar Year 2018 (CY 2018) 200 Areas Pump and Treat Report, Rev. 0, CH2M Hill Plateau Remediation Company, Richland, Washington.

6. Data Acquisition Method(s)

Describe the data acquisition method and associated QA/QC, considering the following:

- Qualifications of personnel or organizations generating the data;*
- Technical adequacy of equipment and procedures used;*
- Environmental and programmatic conditions if germane to the data quality;*
- The extent to which acquisition processes reflect modeling requirements;*
- The quality and reliability of the measurement control program;*
- The degree to which independent audits of the process were conducted;*
- Extent and reliability of the associated documentation.*

The data of water and tritium releases to the SALDS used for the CA/CIE modeling were acquired from Table A-1 of Appendix A in RPP-CALC-61950. Thus, historical releases were used through September 2017 and the estimated future releases were used from October 2017 through Year 2065. The water volumes in Table A-1 have units of gallons. These were summed and converted to cubic meters per year for input to the STOMP model preprocessor using the following equation and rounding the results to 3 significant figures:

$$\text{Volume (m}^3\text{/yr)} = \text{Volume (gal/yr)} * 3.78541 \text{ (L/gal)} * 0.001 \text{ (m}^3\text{/L)}$$

Tritium releases in Table A-1 are in units of curies and no unit conversions were needed. *For databases, identify query language used to obtain data from database (SQL, etc.), briefly describe the query description and attach copy.*

No database queries were performed.

7. Corroborating Data

Identify and discuss any corroborating datasets. Provide any documentation that confirms the corroborating data substantiate existing parameter values, distributions, or data quality.

Data on water volumes and tritium concentrations in the effluent released to the SALDS are reported quarterly to the Washington State Department of Ecology by WRPS in Discharge Monitoring Reports (DMRs). DMRs filed since 2015 are available online at the Washington State Department of Ecology website (permit number ST0004500).

8. Data Quality Considerations

Discuss data quality considerations not identified in other sections. Include discussion of data quality indicators (i.e., accuracy, precision, representativeness, completeness, and comparability).

ENVIRONMENTAL MODELING DATA TRANSMITTAL COVER PAGE (Continued)

No.: EMDT-IN-0047

Revision No: 0

(Request EMDT number for Modeling Team Leader)

Title: SALDS Liquid Disposal Volumes and Tritium Inventory

Date: 3/12/2020

8. Data Quality Considerations

The historical data on water and tritium releases contained in Table A-1 of RPP-CALC-61950 were compared to values in the DMRs and were deemed acceptable for vadose zone transport simulations (see part 10 of the EMDT). The estimates of future water discharges and tritium releases in RPP-CALC-61876 are the best available estimates of future releases to the SALDS.

9. Assumptions and Limitations on Data Use

Document known uncertainties, assumptions, constraints or limits on data.

The assumptions used to prepare the estimates of future water and tritium releases to the SALDS are documented in RPP-CALC-61876. Future estimates of facility operation are always uncertain, but the information contained in RPP-CALC-61876 and in Table A-1 of RPP-CALC-61950 are the best available to support simulations of tritium migration and fate from the SALDS.

DATA CONFIGURATION ITEM SUBMITTAL:**Data Provider Submittal:**

Position: JP McDonald / Sr. Hydrogeologist

John McDonald

Print First and Last Name

Signature

3/12/2020

Date

DATA CONFIGURATION ITEM REVIEW AND VERIFICATION:**10. Verification Process**

Describe steps taken to verify that these data are appropriate for intended use, noting any limitations.

The water volume and tritium release values contained in Table A-1 of RPP-CALC-61950 were spot checked against data from the DMRs. The tritium release values matched well, but some differences were noted in the water discharge volumes. Thus, a check of all water volumes against the DMRs was performed. Differences of greater than 1 percent occurred for the following months:

Oct 1997: 2,619,889 gal in Table A-1; 2,570,000 gal in DMR (1.9% difference)
 Oct 1998: 3,262,365 gal in Table A-1; 2,468,000 gal in DMR (32.2% difference)
 Mar 1999: 1,030,350 gal in Table A-1; 1,009,000 gal in DMR (2.1% difference)
 Apr 1999: 2,622,182 gal in Table A-1; 2,895,000 gal in DMR (9.4% difference)
 Nov 2001: 3,705,367 gal in Table A-1; 3,769,000 gal in DMR (1.7% difference)
 Feb 2012: 1,820,569 gal in Table A-1; 1,988,000 gal in DMR (8.4% difference)

Input to the STOMP model consists of annual values distributed evenly throughout the year. In terms of annual volumes, the differences are low:

1997: 15,262,603 gal sum from Table A-1; 15,213,054 gal sum from DMR (0.3% difference)
 1998: 28,322,095 gal sum from Table A-1; 27,527,000 gal sum from DMR (2.9% difference)
 1999: 23,068,191 gal sum from Table A-1; 23,320,000 gal sum from DMR (1.1% difference)
 2001: 25,922,535 gal sum from Table A-1; 25,985,000 gal sum from DMR (0.2% difference)
 2012: 9,454,636 gal sum from Table A-1; 9,623,000 gal sum from DMR (1.7% difference)

These differences were deemed acceptable for vadose zone transport simulations and the data from Table A-1 can be used for input to the STOMP model.

11. Summary of Data Review

ENVIRONMENTAL MODELING DATA TRANSMITTAL COVER PAGE (Continued)

No.: EMDT-IN-0047

Revision No: 0

(Request EMDT number for Modeling Team Leader)

Title: SALDS Liquid Disposal Volumes and Tritium Inventory

Date: 3/12/2020

11. Summary of Data Review

The review shall ensure that the report meets the listed criteria. Consideration includes ensuring that the data collection method employed was appropriate for the type of data being considered and confidence in the data acquisition and subsequent processing methodology is warranted.

Is documentation technically adequate, complete, and correct?

☒ Yes ☐ No

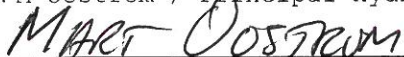
Are uncertainties and limitations on appropriate use of data discussed?

☒ Yes ☐ No

Are the assumptions, constraints, bounds, or limits on the data identified?

☒ Yes ☐ No**APPROVAL OF DATA CONFIGURATION ITEM:****Data Reviewer Approval:**

Position: M Oostrom / Principal Hydrogeologist



Print First and Last Name



Signature

03/12/20

Date

**EMDT accepted for Composite Analysis input in
Data Readiness Review on 10/8/2020.**