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# Effective Removal of Anionic Re(VII) by Surface-Modified $\text{Ti}_2\text{CT}_x$ MXene Nanocomposites: Implications for Tc(VII) Sequestration

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**ABSTRACT:**

Environmental contamination by  $^{99}\text{Tc(VII)}$  from radioactive wastewater streams is of particular concern due to the long half-life of  $^{99}\text{Tc}$  and high mobility of pertechnetate. Herein, we report a novel MXene-polyelectrolyte nanocomposite with three-dimensional networks for enhanced removal of pertechnetate, which is pertechnetate simulant. The introduction of poly(diallyldimethylammonium chloride) (PDDA) regulates the surface charge and improves the stability of  $\text{Ti}_2\text{CT}_x$  nanosheet, resulting in  $\text{Re(VII)}$  removal capacity of up to  $363 \text{ mg g}^{-1}$ , and fast sorption kinetics. The  $\text{Ti}_2\text{CT}_x/\text{PDDA}$  nanocomposite furthermore exhibits good selectivity for  $\text{ReO}_4^-$  when competing anions (such as  $\text{Cl}^-$  and  $\text{SO}_4^{2-}$ ) coexist at a concentration of 1800 times. The immobilization mechanism was confirmed as a sorption-reduction process by batch sorption experiments and X-ray photoelectron spectroscopy. The pH-dependent reducing activity of  $\text{Ti}_2\text{CT}_x/\text{PDDA}$  nanocomposite toward  $\text{Re(VII)}$  was clarified by X-ray absorption spectroscopy. As the pH increases, the local environment gradually changes from octahedral-coordinated  $\text{Re(IV)}$  to tetrahedral-coordinated  $\text{Re(VII)}$ . The overall results suggest that  $\text{Ti}_2\text{CT}_x/\text{PDDA}$  nanocomposite may be a promising candidate for efficient elimination of Tc contamination. The reported surface modification strategy might result in applications of MXene-based materials in environmental remediation of other oxidized anion pollutants.

## INTRODUCTION

Technetium-99 ( $^{99}\text{Tc}$ ), a  $\beta$  decay emitter produced from nuclear fission of  $^{235}\text{U}$  and fabrication of weapon grade plutonium, is a particularly problematic radionuclide in legacy nuclear waste due to long half-life ( $2.13 \times 10^5$  years) and high thermal neutron fission yield (6.1%).<sup>1,2</sup> In oxygenated aqueous solutions, Tc is usually present as Tc(VII) in the form of pertechnetate anion,  $\text{TcO}_4^-$ .<sup>3,4</sup> Because of high water solubility and low retention on soil and natural minerals,  $\text{TcO}_4^-$  exhibits a high environmental mobility, which results in Tc-contamination around nuclear waste reprocessing and storage sites.<sup>5-7</sup> Accordingly, there is need for efficient and rapid removal of Tc(VII) from radioactive environmental wastewater, which is a challenging task.

Several types of solid sorbents have been explored for the remediation of Tc(VII) and other toxic anions, including layered double hydroxides (LDHs),<sup>8,9</sup> activated carbon,<sup>10</sup> ion-exchange resins,<sup>11,12</sup> functionalized polymers,<sup>13-15</sup> porous aromatic frameworks,<sup>16</sup> and metal organic frameworks (MOFs).<sup>17-20</sup> For instance, Wang et al. synthesized a water-stable and radiation-resistant cationic MOF, SCU-101, that selectively recognizes  $\text{TcO}_4^-$  via strong hydrogen-bond interactions.<sup>21</sup> As an alternative to ion-exchange, a promising strategy for Tc remediation is reduction of Tc(VII) to insoluble and relatively immobile Tc(IV).<sup>22</sup> Techniques including radiolysis reduction (via  $\gamma$  radiation in alkaline solution),<sup>23,24</sup> bioreduction,<sup>5,25,26</sup> photoreduction<sup>27</sup> and electroreduction<sup>28,29</sup> have been developed for Tc(VII) immobilization. Chemical reduction is attractive due to ease of implementation, rapid kinetics, and low cost. Current candidate solid reduction agents focus on a handful of materials including zero valent iron,<sup>30,31</sup> Fe(II)-containing compounds,<sup>22,32</sup> sulfides,<sup>33,34</sup> Sn(II/IV) phosphate,<sup>35</sup> and composites of these constituents. Limitations of Tc(VII) reductants include low removal capacity and poor selectivity, and novel materials with reductive activity and sequestration capacity for Tc(VII) remain a goal.

Two-dimensional transition metal carbide (MXene) is a new class of layered nanomaterials that has emerged in the energy and environmental fields due to

intriguing physicochemical properties.<sup>36-39</sup> In particular, it has been demonstrated that multilayered MXenes are promising candidates for clean-up of several radionuclide cations (e.g.  $^{238}\text{UO}_2^{2+}$ ,  $^{232}\text{Th}^{4+}$ ,  $^{133}\text{Ba}^{2+}$  and  $^{107}\text{Pd}^{2+}$ ) by versatile interaction pathways that include coordination, ion exchange and reduction immobilization.<sup>40-45</sup> For sequestration of anions such as  $\text{UO}_2(\text{CO}_3)_3^{4-}$  and  $\text{ReO}_4^-$ , MXenes are generally ineffective due to their intrinsically negatively charged surfaces and the limited number of active sites on multilayer structures.<sup>40</sup> Although effective anion adsorption and reduction by MXenes has been demonstrated under very acidic conditions, when the surface charge becomes positive,<sup>46,47</sup> the narrow useful pH range may restrict practical environmental applications. A straightforward strategy to overcome this challenge is regulation of MXene surface properties through hybridization and modification. Herein, we report a simple, cost-effective and scalable method to fabricate  $\text{Ti}_2\text{CT}_x$ /poly(diallyldimethylammonium chloride) (PDDA) nanocomposites for enhanced removal of  $\text{ReO}_4^-$  from aqueous solutions. These components were chosen because  $\text{Ti}_2\text{CT}_x$  is a highly active MXene with reducing capability, and PDDA is a common cationic polyelectrolyte that is widely used in water treatment and mining.<sup>48,49</sup>  $\text{ReO}_4^-$  is sufficiently chemically similar to  $\text{TcO}_4^-$  to serve as a non-radioactive surrogate, though extrapolation from  $\text{ReO}_4^-$  to  $\text{TcO}_4^-$  must be judicious.<sup>50-52</sup> In the present work, different morphologies of  $\text{Ti}_2\text{CT}_x$  (multilayered or nanosheet MXene) and their nanocomposites with PDDA were synthesized and characterized. Re(VII) removal was evaluated by batch sorption and anion competition experiments. The removal mechanisms were clarified using X-ray absorption spectroscopy and photoelectron spectroscopy.

## EXPERIMENTAL DETAILS

**Synthesis of  $\text{Ti}_2\text{CT}_x$  and its composites.** 4.07 g LiF was added to 66.7 mL of 7.5 M HCl and stirred for several minutes to yield a transparent solution. 2.4 g  $\text{Ti}_2\text{AlC}$  powder was added in batches of ca. 80 mg each during 20 min, after which the

solution was stirred at 35 °C for 72 h to ensure complete etching of Al. The product was washed with deionized water and centrifuged for 5 min at 3500 rpm. The wash/centrifuge cycle was repeated until a stable dark-red supernatant was obtained, indicating a spontaneous delamination of  $\text{Ti}_2\text{CT}_x$  MXene. Finally, deionized water was added, followed by shaking and centrifugation at 2000 rpm for 30 min to collect the colloidal solution of  $\text{Ti}_2\text{CT}_x$  nanosheets (TCNS). The concentration of LiF ( $C_{\text{LiF}}$ , optimized here as 2.35 M) is a key parameter for high-yield synthesis of TCNS (~80 wt% yield based on added  $\text{Ti}_2\text{AlCl}_x$ ). No delamination was evident for  $C_{\text{LiF}}$  less than 2.3 M, presumably indicating this as the minimum concentration of  $\text{Li}^+$  needed to sufficiently weaken the interaction between MXene interlayers.<sup>53</sup> Higher  $C_{\text{LiF}}$  (e.g. 2.7 M) resulted in continued dissolution of TCNS and formation of  $\text{Li}_3\text{AlF}_6$  impurity. The synthesis of multilayered  $\text{Ti}_2\text{CT}_x$  (MTC) was as reported previously,<sup>40</sup> using  $C_{\text{LiF}} = 1\text{M}$ .

The  $\text{Ti}_2\text{CT}_x$  nanosheet/PDDA composite (TCNS-P) was synthesized by dropwise addition of 100 mL 2 wt% PDDA ( $M_w = 100000\text{-}200000$ ) aqueous solution into 100 mL TCNS colloidal solution ( $10\text{ mg mL}^{-1}$ ) under vigorous stirring. The mixture was further stirred for 2 h under argon, after which the sediment was centrifuged at 10000 rpm and washed with deionized water for three cycles to remove excess PDDA. A black aerogel cake of TCNS-P was obtained by drying the gel-like product in a freeze dryer at  $-60\text{ °C}$  for 48 h (Figure S1). The composite of multilayered  $\text{Ti}_2\text{CT}_x$ /PDDA (MTC-P) was prepared by this same procedure, but using MTC instead of TCNS.

**Batch Sorption Experiments.** Re(VII) sorption experiments were conducted in an anaerobic glove box at 303 K. In a typical procedure, 8 mg sorbent sample was added to 20 mL of  $100\text{ mg L}^{-1}$  Re(VII) ( $\text{NH}_4\text{ReO}_4$ , Sinopharm Chemical Reagent Co. Ltd.). After adding 0.1 M NaOH or 2 M HCl to adjust the pH, the suspension was shaken slightly for about 48 h to ensure equilibrium adsorption and reduction of Re(VII). The supernatant was then separated by centrifugation at 10000 rpm, and diluted with 5 wt%  $\text{HNO}_3$  to ensure that the concentration of Re (VII) was in the range of 1-5 mg

L<sup>-1</sup>. The residual solution concentration of Re(VII) was measured using inductively coupled plasma-optical emission spectrometry (ICP-OES, Horiba JY 2000-2, detection limit below 0.01 mg L<sup>-1</sup>). The removed Re(VII) ( $q_e$ ) was calculated by the following equation:  $q_e=(C_0-C_e)V/m$ , where  $C_0$  and  $C_e$  are the initial and equilibrium concentrations of Re(VII),  $m$  is the sorbent mass, and  $V$  is the solution volume. Concentrations of 10 and 100 mmol L<sup>-1</sup> Na<sub>2</sub>SO<sub>4</sub>, NaCl, NaClO<sub>4</sub> or NaNO<sub>3</sub> were employed to evaluate effects of competitive anions on Re(VII) removal.

**X-ray absorption spectroscopy.** XANES and EXAFS spectra at the Re L<sub>3</sub> absorption edge (~10,535 eV) were collected at beamline 1W1B of the Beijing Synchrotron Radiation Facility (BSRF). The ReO<sub>2</sub> reference spectrum was measured in transmission mode, and spectra of the aqueous Re(VII) reference and other samples were measured in fluorescence mode. XANES and EXAFS data were pre-edge background subtracted, normalized, and fitted by the Athena and Artemis programs in the IFEFFIT package.<sup>54</sup> A  $k$  range of ~ 3.0 – 13.0 Å<sup>-1</sup> and Rbkg 1.1 were used to obtain the Fourier transform of Re L<sub>3</sub>  $k^3$ -weighted spectra. Four scattering pathways (Re–O<sub>1</sub>, Re–Re, Re–O<sub>MS</sub> and Re–O<sub>2</sub>, nomenclature explained below) based on the crystal structures of ReO<sub>2</sub> and KReO<sub>4</sub> were used to calculate the theoretical phase and amplitudes for the extraction of metric parameters from EXAFS data. All fitting operations were performed in  $R$  space of ~ 1.1–4.0 Å. The amplitude reduction factor ( $S_0^2$ ) was set to 0.83 as determined from EXAFS analysis of reference samples.

## RESULTS AND DISCUSSION

**Characterization of Ti<sub>2</sub>CT<sub>x</sub> and its composites.** The morphologies of as-synthesized Ti<sub>2</sub>CT<sub>x</sub> and its composites are shown in Figures 1A–1D. MTC and MTC-P have a typical graphite-like stacking structure. TCNS is dispersed nanosheets from delamination of multilayered Ti<sub>2</sub>CT<sub>x</sub> under high concentration of Li<sup>+</sup>. After interaction with PDDA, the nanosheets are assembled into a porous three-dimensional network with a pore size around 10 μm (Figure 1D). The Brunauer–Emmett–Teller

(BET) surface area of TCNS-P is  $18 \text{ m}^2 \text{ g}^{-1}$  (Figure S2), which is larger than for MTC-P ( $5 \text{ m}^2 \text{ g}^{-1}$ ), but smaller than for TCNS ( $31 \text{ m}^2 \text{ g}^{-1}$ ) probably due to coating and linking of nanosheets by PDDA. Powder XRD results for the four samples are shown in Figure 1E. The (002) peak at  $\sim 7.8^\circ$  for MTC, which is consistent with the literature,<sup>40</sup> is slightly shifted to  $7.6^\circ$  for MTC-P. For TCNS, the broadening of the (002) peak and disappearance of the non-basal (110) peak at  $61^\circ$  indicate successful exfoliation of  $\text{Ti}_2\text{CT}_x$  lamellas. Formation of TCNS-P from TCNS and PDDA results in a significant decrease in diffraction intensity and a (002) peak shift from  $8.2^\circ$  to  $6.1^\circ$ , which corresponds to an increase of  $d$ -spacing from  $10.8 \text{ \AA}$  to  $14.5 \text{ \AA}$ . These results reveal a highly disordered nature of TCNS-P and the intercalation of some PDDA molecules inside few-layered  $\text{Ti}_2\text{CT}_x$  nanosheets. The FTIR spectra in Figure S3 also demonstrate co-incorporation of PDDA and  $\text{Ti}_2\text{CT}_x$  in TCNS-P, as indicated by PDDA C=C modes at  $1474 \text{ cm}^{-1}$  and C-H stretches at  $2952 \text{ cm}^{-1}$ ,<sup>48,55,56</sup> and the Ti-O MXene band at  $560 \text{ cm}^{-1}$ .<sup>57</sup> TGA measurements provide estimated PDDA contents of 4% and 45% in MTC-P and TCNS-P, respectively (Figure 1F). The weight gain of MTC, MTC-P and TCNS in temperature range of  $350 - 500^\circ\text{C}$  corresponds to the oxidation of  $\text{Ti}_2\text{CT}_x$  MXene under air atmosphere. TCNS exhibits a lower oxidation temperature than MTC and MTC-P, suggesting the higher reactivity of nanosheet MXene. Surface modification of  $\text{Ti}_2\text{CT}_x$  MXene with PDDA was further confirmed by zeta potential data, as shown in Figure 1G. Pristine MXenes MTC and TCNS have similar zero charge points (around pH 2~3), with their zeta potentials becoming more negative as pH increases, presumably due to  $\text{Ti-O}^-$  groups on MXene surface for  $\text{pH} > 3.0$ . In contrast, MTC-P and TCNS-P exhibit positively charged surfaces over the entire studied pH range of 1.0 to 10.0. The larger positive zeta potential of TCNS-P versus MTC-P (e.g.  $+59 \text{ mV}$  vs.  $+41 \text{ mV}$  at pH 4.0), suggests that the morphology of super-thin nanosheets facilitates functionalization of  $\text{Ti}_2\text{CT}_x$  by PDDA.

**Batch Sorption Experiments.** Figure 2A shows removal of  $\text{Re(VII)}$  by the four

MXene samples as a function of pH. All samples exhibit a high Re(VII) removal capacity at acidic pH 1.0, presumably because a positively charged  $Ti_2CT_x$  surface facilitates electrostatic interaction with Re(VII) anions. As discussed below, the high reactivity of  $Ti_2CT_x$  allows reduction of Re(VII), accompanied by consumption of protons from aqueous solutions,<sup>29</sup> with the result that high acidity can promote reduction and thus removal of Re(VII) as Re(IV). However, such acidic pH has the adverse effect of  $Ti_2CT_x$  MXene dissolution (ca. 55~60% at pH 1.0 based on Ti content, Figure S4). Figure S4 also indicates that the PDDA component in TCNS-P greatly suppresses dissolution at pH > 2.0, thereby improving  $Ti_2CT_x$  nanosheet stability. The drastic decrease in removal capacity for MTC and TCNS with increasing pH above 1.0 (Figure 2A) is attributed to electrostatic repulsion between negative surface of MXene and anionic Re(VII). Modification by PDDA enables MXene to exhibit a positively charged surface nearly independent of pH, which results in attraction between TCNS-P and Re(VII) over the entire pH range of 1~10 (Figure 2A). In contrast to pH-independent high Re(VII) uptake by TCNS-P, the uptake by MTC-P decreases rapidly with increasing pH despite an apparently positive surface. The very similar XRD patterns of MTC-P and MTC (Figure 1E) suggest that modification by PDDA occurs only on the exposed exterior surface of MTC, probably due to the small interlayer spacing that renders the multilayered structures inaccessible to PDDA. The unmodified surfaces inside MTC-P may thus remain negatively charged at high pH, which hinders diffusion and uptake of Re(VII) anions.

Kinetic experiments were conducted to evaluate the removal rate and controlling mechanism for Re(VII) uptake. As illustrated in Figure 2B,  $Ti_2CT_x$  and its composites exhibit rapid sorption kinetics, with equilibration in less than 1 h. Kinetic fitting parameters based on pseudo-first-order and pseudo-second-order models are in Table S2, where it is apparent that the latter model distinctively provides high correlation coefficients and also removal capacities in good accord with experiment. The kinetic results indicate that removal of Re(VII) by  $Ti_2CT_x$  and its composites is controlled

primarily by chemisorption, rather than by mass transport.<sup>58,59</sup> Particularly noteworthy is that the Re(VII) removal capacity of TCNS-P is an order of magnitude higher than that of TCNS or MTC (Figure 2B), as a result of surface modification and more efficient utilization of interaction sites.

Figure 2C shows sorption isotherms of Re(VII) for the MXene materials. The isotherms were fitted using the Langmuir and Freundlich models with fitting parameters as summarized in Table S3. The correlation coefficients reveal that Re(VII) sorption on the PDDA modified MXenes, TCNS-P and MTC-P, is well fit by Langmuir model, while sorption by pristine MXenes, TCNS and MTC, is better represented by Freundlich model. Heterogeneous adsorption of Re(VII) on TCNS and MTC is attributed to various active MXene surface sites, such as terminating groups –OH, –O, –F and –Cl.<sup>40,41,60</sup> Introduction of PDDA screens these functional groups, with the result that electrostatic attraction of Re(VII) by quaternary ammonium cations on the polyelectrolyte PDDA backbone becomes dominant, which is better represented by Langmuir model. It is evident from the results in Figure 2 that for the studied concentration range TCNS-P exhibits a great enhancement in removal capacity compared to the other MXene samples. The maximum sorption capacity of TCNS-P, estimated as 363 mg g<sup>-1</sup>, is higher than, or at least comparable to traditional nanosorbents and ion-exchange resins (Table S4).

To assess the potential efficacy of TCNS-P in environmental applications, we investigated effects of different types and concentrations of competitive anions on Re(VII) removal at pH 4.0. The results in Figure 2D demonstrate that TCNS-P can remove at least 94% of Re(VII) even in the presence of a 1860 times higher concentration of SO<sub>4</sub><sup>2-</sup> or Cl<sup>-</sup>. The resulting distribution coefficient,  $K_d$ , is  $> 2.1 \times 10^4$  mL g<sup>-1</sup>, which indicates that TCNS-P is a promising candidate to efficiently remove Re(VII) from wastewater containing excess of competing anions. It is apparent in Figure 2D that competing ClO<sub>4</sub><sup>-</sup> substantially suppresses uptake of Re(VII), probably due to the similar structures, ionic radii and hydration energies of ClO<sub>4</sub><sup>-</sup> and ReO<sub>4</sub><sup>-</sup>.

<sup>8,61,62</sup> Nevertheless, the removal percentage of Re(VII) still remains at 77% in the presence of excess  $\text{ClO}_4^-$  at a moderate ionic strength of  $10 \text{ mmol L}^{-1}$ . Zhang et al. found that the affinity of quaternary amine for anions followed the same order with our our results, namely, perchlorate  $\gg$  sulfate  $>$  chloride. <sup>63</sup> Roach et al. also reported high perchlorate retentions of PDDA in simulated groundwater containing chloride and sulfate. <sup>64</sup> We conjecture that the alkyl backbone of PDDA may provide a weak hydrophobic environment to stabilize perrhenate/perchlorate with low hydration energies. On the other hand, Re(VII) could be simultaneously reduced by  $\text{Ti}_2\text{CT}_x$  MXene under pH 4.0 (see below), thus promoting the selectivity of TCNS-P to perrhenate. Apparent in Figure 2D is that the presence of competing  $\text{NO}_3^-$  more drastically inhibits Re(VII) removal, which may be related to the oxidizing properties of nitrate (redox potential = + 0.96 V). <sup>65</sup> Maset et al. reported the recalcitrance of Re(VII) under nitrate-containing biogeochemical reduction conditions, which is consistent with our results. <sup>7</sup> Upon increasing the amount of adsorbent from  $0.75 \text{ g L}^{-1}$  to  $5 \text{ g L}^{-1}$  under nitrate conditions, Re(VII) removal increased from about 30% to nearly 60% (Figure S5). The effect of competing anions on Re(VII) uptake at pH 2.0 (Figure S6) follows the same trend as at pH 4.0:  $\text{NO}_3^-$  (most inhibiting)  $>$   $\text{ClO}_4^-$   $>$   $\text{SO}_4^{2-} \approx \text{Cl}^-$  (least inhibiting).

**Speciation of Rhenium and Removal Mechanism.** To evaluate oxidation state information for Re-loaded TCNS-P at different pH, the Re  $L_3$  edge XANES spectra shown in Figure 3 were acquired. The 1.8 eV energy shift for the strongest XANES absorption resonance (i.e. the “white line”) between Re(IV) and Re(VII) is in good agreement with literature results. <sup>66</sup> In the acidic pH range of 1.0-4.0, the white line energies for Re-loaded TCNS-P are close to that of  $\text{ReO}_2$ , indicating predominantly reduction of Re(VII) to Re(IV) by TCNS-P. At pH 4.0, the white line energy is  $\sim 0.4$  eV higher than for  $\text{ReO}_2$ , revealing partial reduction of Re(VII). At neutral pH 7.0, the white line position indicates only minor reduction of Re(VII), and a pH 10.0 there is negligible reduction. In addition to the white line, the changes in less intense

XANES features in the range 10570~10650 eV also indicate Re oxidation state. In particular, the weak absorption resonance at 10588 eV suggests an oxidation state closer to Re(IV) under acidic conditions. Under neutral and alkaline conditions, the 10588 eV resonance is replaced by a broadened peak around 10613 eV, consistent with Re(VII) reference. Oxidation state fractions were estimated from a comparison of the measured absorption edge energy ( $E_0$ ) with a linear interpolation between  $\text{ReO}_2$  and  $\text{ReO}_4^-$  (Table S1). The estimated fraction of Re(IV) decreases from 89% to 22% as the solution pH increases from 3.0 to 7.0. The overall XANES results indicate that reduction of aqueous Re(VII) to Re(IV) by TCNS-P is strongly pH dependent.

EXAFS data can elucidate the local structure around metal atoms.<sup>67,68</sup> Figure 4 shows Re  $L_3$  edge  $k^3$ -weighted spectra, along with their Fourier transforms (FT) and the corresponding fitting results. The differing oscillation periods for low- $k$  ( $k < 7 \text{ \AA}^{-1}$ ) in Figure 4A under acidic and basic conditions suggest different first coordination shells around Re. FT fitting (Figure 4B and Table 1) shows that under acidic conditions there are six O atoms at 2.03  $\text{\AA}$  ( $\text{Re-O}_1$ ) and a Re-Re distance of 2.58  $\text{\AA}$ , which are consistent with distances of Re-O (2.01  $\text{\AA}$ ) and Re-Re (2.57  $\text{\AA}$ ) in octahedral-coordinated  $\text{ReO}_2$ . The absence of multiple scattering interactions ( $\text{Re-O}_{\text{MS}}$ ) in samples for pH 1.0-4.0 may be due to asymmetrical octahedral structures. The coordination number (N) of Re-Re is close to 2.0, indicating the formation of  $\text{ReO}_2$  one-dimensional chains in an edge-sharing fashion.<sup>1,23</sup> A significantly shorter Re-O bond with a distance of 1.74  $\text{\AA}$  ( $\text{Re-O}_2$ ) is observed for alkaline conditions (pH 10.0), which is close to the Re-O distance in tetrahedral  $\text{ReO}_4^-$  (1.73  $\text{\AA}$ ). The presence of two types of Re-O bonds ( $\text{Re-O}_1$  and  $\text{Re-O}_2$ ), and one Re-Re bond, suggests both tetrahedral and octahedral-coordinated Re for pH 4.0 and pH 7.0. Between pH 1.0 and pH 10.0, the total  $N(\text{Re-O})$  decreases from  $6.0 \pm 0.5$  to  $3.7 \pm 0.3$  (Table 1), while concomitantly  $N(\text{Re-O}_1)$  and  $N(\text{Re-Re})$  decrease, and  $N(\text{Re-O}_2)$  increases. It is apparent that the local coordination environment of Re

gradually changes from octahedral-coordinated Re(IV) to tetrahedral-coordinated Re(VII) as pH increases, in accord with less reduction.

SEM images, XPS spectra and XRD patterns were acquired to assess reduction of Re(VII) as a function of contact time with TCNS-P. As shown in Figure S7, the layered structure of  $\text{Ti}_2\text{CT}_x$  nanosheets in TCNS-P is retained after Re(VII) removal, though the porous 3D network partially collapses. Fitting of the Re-loaded TCNS-P XPS spectra in the Re 4f region (Figure 5) indicates two sets of spin-orbit split doublets. The lines around 42.6 and 45.0 eV are assigned to Re(IV)  $4f_{7/2}$  and Re(IV)  $4f_{5/2}$ , and the lines around 45.7 and 48.1 eV are assigned to Re(VII)  $4f_{7/2}$  and Re(VII)  $4f_{5/2}$ , respectively.<sup>69-71</sup> The assignment of both Re(IV) and Re(VII) from XPS is consistent with the XANES and EXAFS results. As the contact time increased from 6 h to 48 h, the Re(IV) fraction increased from approximately 30% to 85% (Figure 5), revealing that Re(VII) reduction is slower than adsorption, which achieves equilibrium within 1 h. High-resolution XPS spectra in the Ti 2p and O 1s regions (Figure S8) suggest that  $\text{Ti}_2\text{CT}_x$  in TCNS-P was gradually oxidized as Re(VII) was reduced, with evidence for loss Ti–C bonds and conversion of Ti–OH to Ti–O. Despite low crystallinity, the corrosion product of TCNS-P after 48 h reaction with Re(VII) can be roughly indexed to anatase (Figure S9), in agreement with our previous results.<sup>40</sup> As no Re-containing crystalline product was identified from the XRD patterns, we conclude that the reduced Re(IV) may be amorphous  $\text{ReO}_2$ .

Based on the combined results of sorption experiments and reaction product analysis, a postulated mechanism for removal of Re(VII) by TCNS-P is summarized as follows: Anionic  $\text{ReO}_4^-$  is first adsorbed on positively surface-charged TCNS-P via a fast electrostatic interaction. The  $\text{Ti}_2\text{CT}_x$  nanosheets then more gradually reduce some or most  $\text{ReO}_4^-$  to amorphous  $\text{ReO}_2$  chains in the confined space, while  $\text{Ti}_2\text{CT}_x$  is partially oxidized to anatase.

**Environmental Significance.** Technetium contamination is a near-surface environmental remediation concern around legacy nuclear facilities due to the high

mobility of Tc(VII). The demonstration here that surface-modified  $\text{Ti}_2\text{CT}_x$  nanosheets can significantly enhance adsorption of Re(VII), with ultimate reduction to insoluble  $\text{ReO}_2$ , suggests potential applications of TCNS-P for remediation of environmental Tc(VII). The standard redox potentials for Tc(VII) to Tc(IV) (+0.74 V) and Re(VII) to Re(IV) (+0.51 V) indicate that reduced Tc(IV) should be more stable compared with Re(IV).<sup>7,72</sup> Optimization here of conditions where Re(IV) dominates (e.g. pH 1~4) should guide development of applications of  $\text{Ti}_2\text{CT}_x$ -based materials for corresponding reduction of Tc(VII). TCNS-P has been demonstrated to behave as a slow-release electron donor, which could provide redox buffering for enhanced Tc immobilization compared with liquid reducing agents.<sup>26</sup> Identification of anatase as the product of Re(VII) reduction by  $\text{Ti}_2\text{CT}_x$  nanocomposite implies a stable matrix for sequestration of Tc(IV), which exhibits similar ionic radius and octahedral coordination structures as Ti(IV).<sup>65,73</sup> The results for Re(VII) sequestration should motivate development of Ti-based MXenes as novel materials for reduction of Tc(VII) with in situ incorporation as Tc(IV).

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**Supporting Information.** Details of characterization methods; sorption kinetics

and isotherm model fitting; N<sub>2</sub> adsorption-desorption isotherms; FTIR spectra; Ti dissolution rates; effects of m/V on Re(VII) removal under nitrate conditions; SEM images; Ti 2p and O 1s XPS spectra; XRD patterns of Re-loaded TCNS-P; estimation of oxidation states in Re-loaded samples; comparison of Re(VII) removal capacities by various adsorbents. The supporting information is available free of charge via the Internet at <http://pubs.acs.org>.

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**Figure captions:**

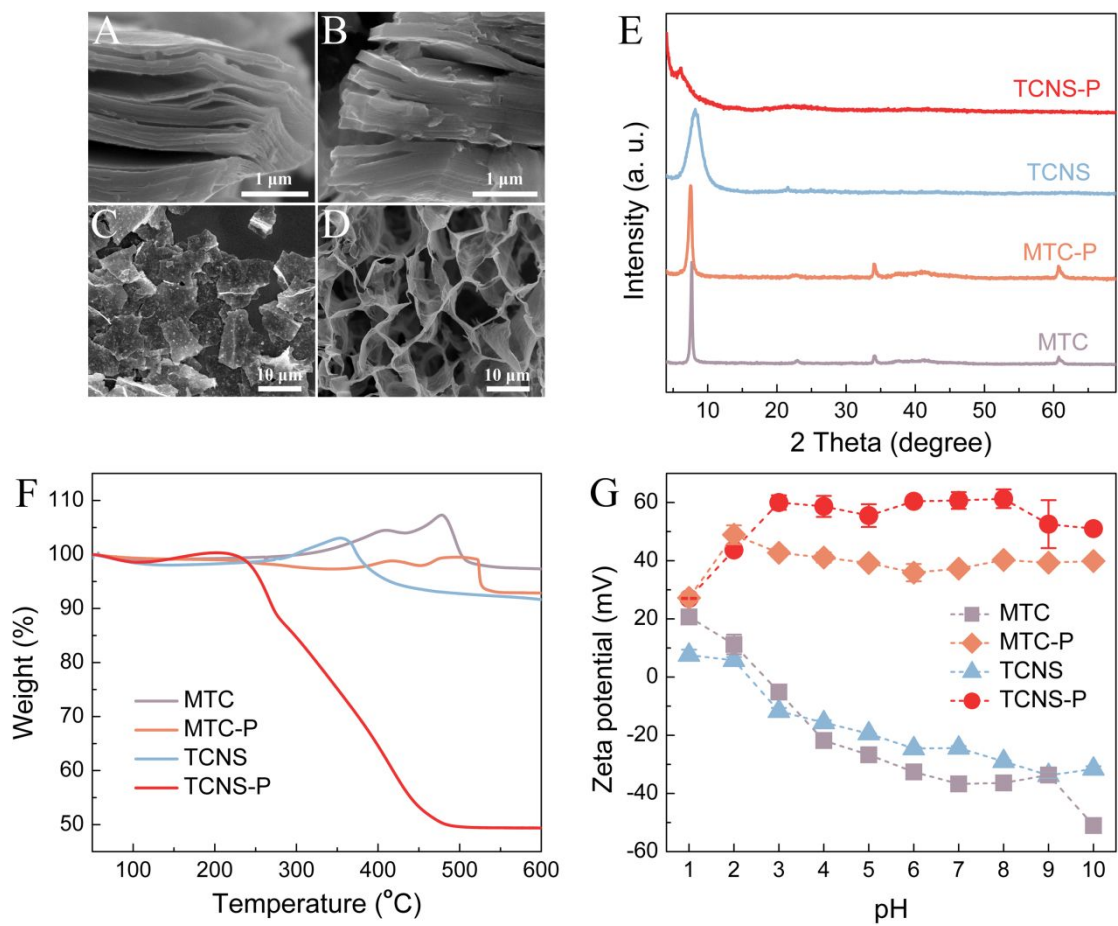
**Figure 1.** SEM images (A–D), powder XRD patterns (E), thermogravimetric weight-loss curves (F), and zeta potentials versus solution pH (G) of  $\text{Ti}_2\text{CT}_x$  and its nanocomposites. (A) MTC; (B) MTC-P; (C) TCNS; (D) TCNS-P.

**Figure 2.** (A–C) Re(VII) removal from aqueous solution by  $\text{Ti}_2\text{CT}_x$  and PDDA hybrids as a function of pH, contact time and equilibrium Re(VII) concentration. (D) Effect of competitive anion species on Re(VII) removal by TCNS-P. Experimental parameters: (A)  $m/V = 0.4 \text{ g L}^{-1}$ ,  $C_0 = 100 \text{ mg L}^{-1}$ ; (B)  $\text{pH} = 4.0 \pm 0.1$ ,  $m/V = 0.4 \text{ g L}^{-1}$ ,  $C_0 = 100 \text{ mg L}^{-1}$ ; (C)  $\text{pH} = 4.0 \pm 0.1$ ,  $m/V = 0.4 \text{ g L}^{-1}$ ,  $C_0 = 5\text{--}400 \text{ mg L}^{-1}$ ; (D)  $\text{pH} = 4.0 \pm 0.1$ ,  $m/V = 0.75 \text{ g L}^{-1}$ ,  $C_0 = 10 \text{ mg L}^{-1}$ ,  $I = 10$  or  $100 \text{ mmol L}^{-1}$ .

**Figure 3.** Re  $L_3$  edge XANES spectra of Re-loaded TCNS-P for different solution pH. Re  $L_3$  edge XANES spectra of Re(IV) and Re(VII) references are shown for comparison.

**Figure 4.** (A) Re  $L_3$  edge  $k^3$ -weighted EXAFS spectra (solid lines) and the best theoretical fits (dotted lines) for Re-loaded TCNS-P samples under different pH. (B) Corresponding non-phase shift corrected Fourier transforms. (a)  $\text{ReO}_2$  reference; (b) pH 1.0; (c) pH 2.0; (d) pH 3.0; (e) pH 4.0; (f) pH 7.0; (g) pH 10.0; (h) Aqueous Re(VII) reference.

**Figure 5.** XPS spectra of Re-loaded TCNS-P for contact times at pH 3.0. (A) 6 h; (B) 24 h; (C) 48 h.

**Figure 1**

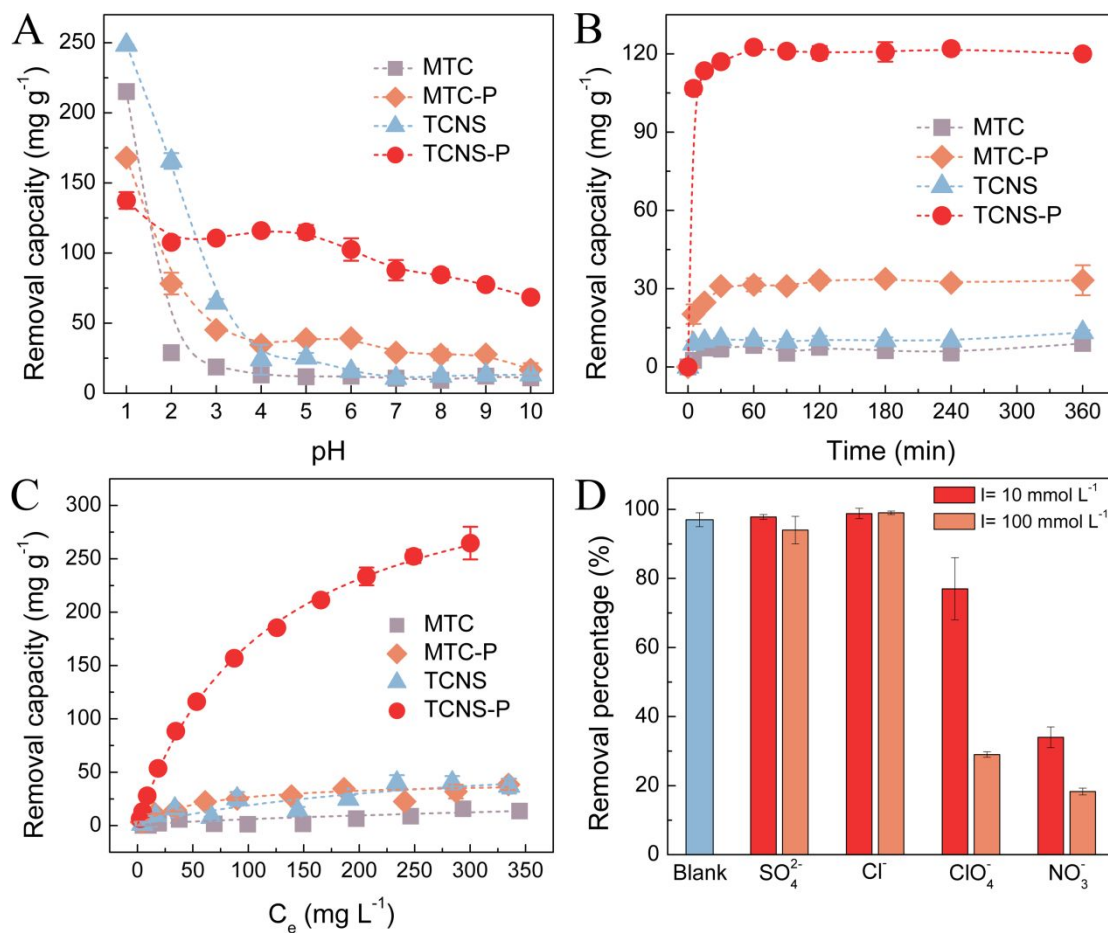
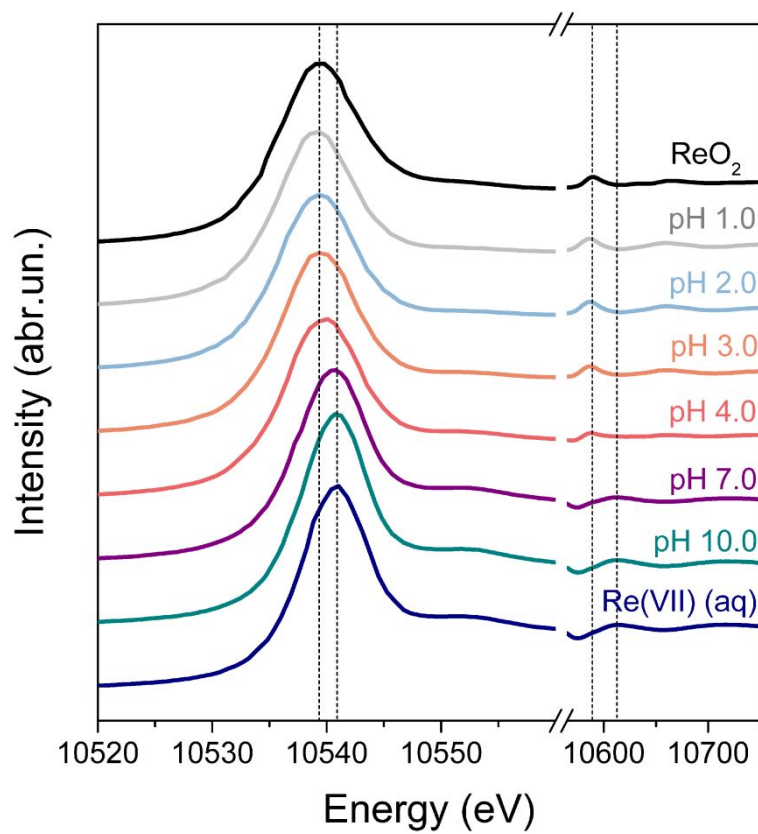
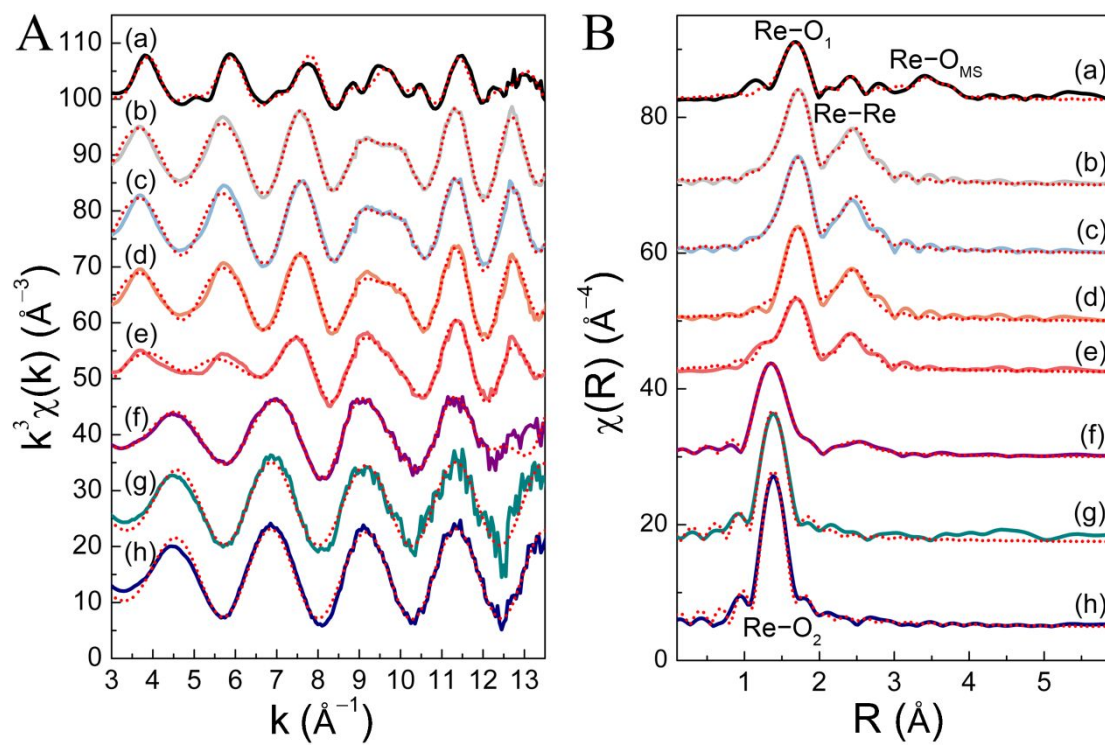


Figure 2

**Figure 3**

**Figure 4**

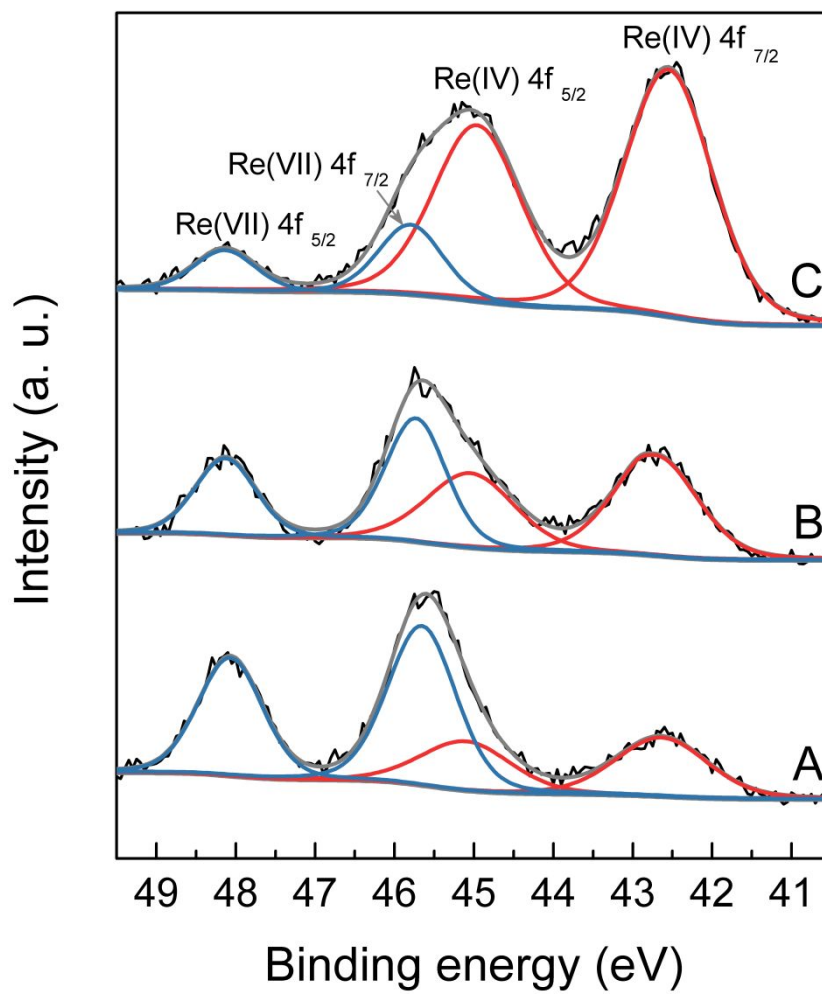
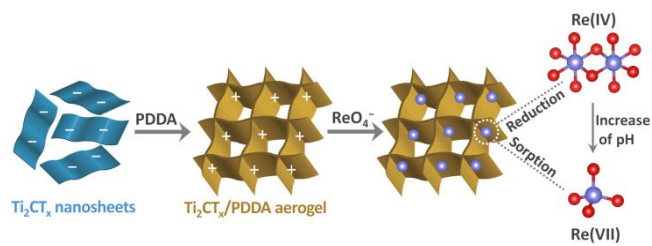


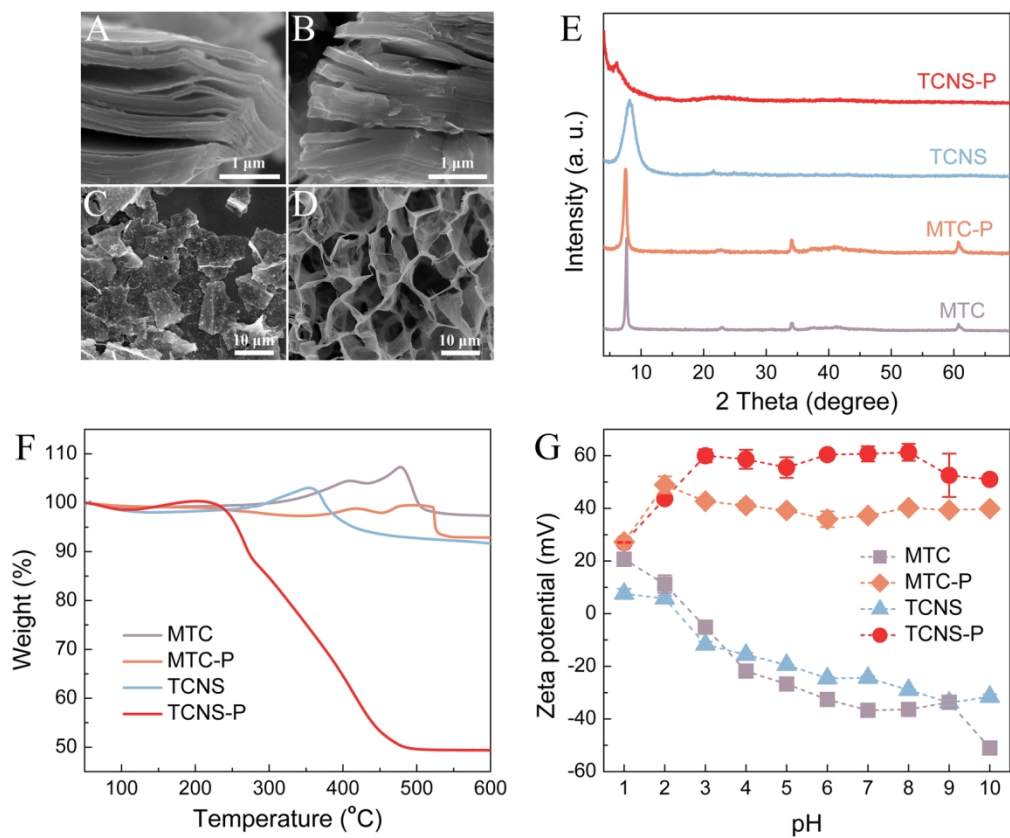
Figure 5

**Table 1.** Fitting Parameters from Least-Squares Fitting Analysis of EXAFS Spectra.

sample	path	N <sup>a</sup>	R(Å) <sup>b</sup>	σ <sup>2</sup> (Å <sup>2</sup> ) <sup>c</sup>	ΔE(eV) <sup>d</sup>	R-factor <sup>e</sup>
ReO <sub>2</sub>	Re–O <sub>1</sub>	6 <sup>f</sup>	2.01±0.01	0.002	11.2	0.031
	Re–Re	2 <sup>f</sup>	2.57±0.01	0.003		
	Re–O <sub>MS</sub>	4 <sup>f</sup>	3.89±0.01	0.003		
pH 1.0	Re–O <sub>1</sub>	6.0±0.5	2.03±0.01	0.003	8.7	0.007
	Re–Re	2.0±0.5	2.58±0.01	0.002		
pH 2.0	Re–O <sub>1</sub>	6.0±0.4	2.03±0.01	0.003	9.3	0.012
	Re–Re	2.0±0.5	2.58±0.01	0.003		
pH 3.0	Re–O <sub>1</sub>	4.7±0.7	2.03±0.01	0.001	9.1	0.017
	Re–Re	2.1±0.9	2.58±0.01	0.002		
pH 4.0	Re–O <sub>1</sub>	4.0±0.8	2.04±0.01	0.001	10.9	0.014
	Re–O <sub>2</sub>	1.5±0.5	1.73±0.01	0.002		
	Re–Re	1.4±0.5	2.58±0.01	0.002		
pH 7.0	Re–O <sub>1</sub>	0.8±0.3	2.09±0.01	0.002	4.6	0.002
	Re–O <sub>2</sub>	2.8±0.2	1.71±0.01	0.001		
	Re–Re	0.9±0.6	2.58±0.01	0.004		
pH 10.0	Re–O <sub>2</sub>	3.7±0.3	1.74±0.01	0.001	10.1	0.011
Re(VII) (aq)	Re–O <sub>2</sub>	4.0±0.2	1.73±0.01	0.001	9.9	0.014

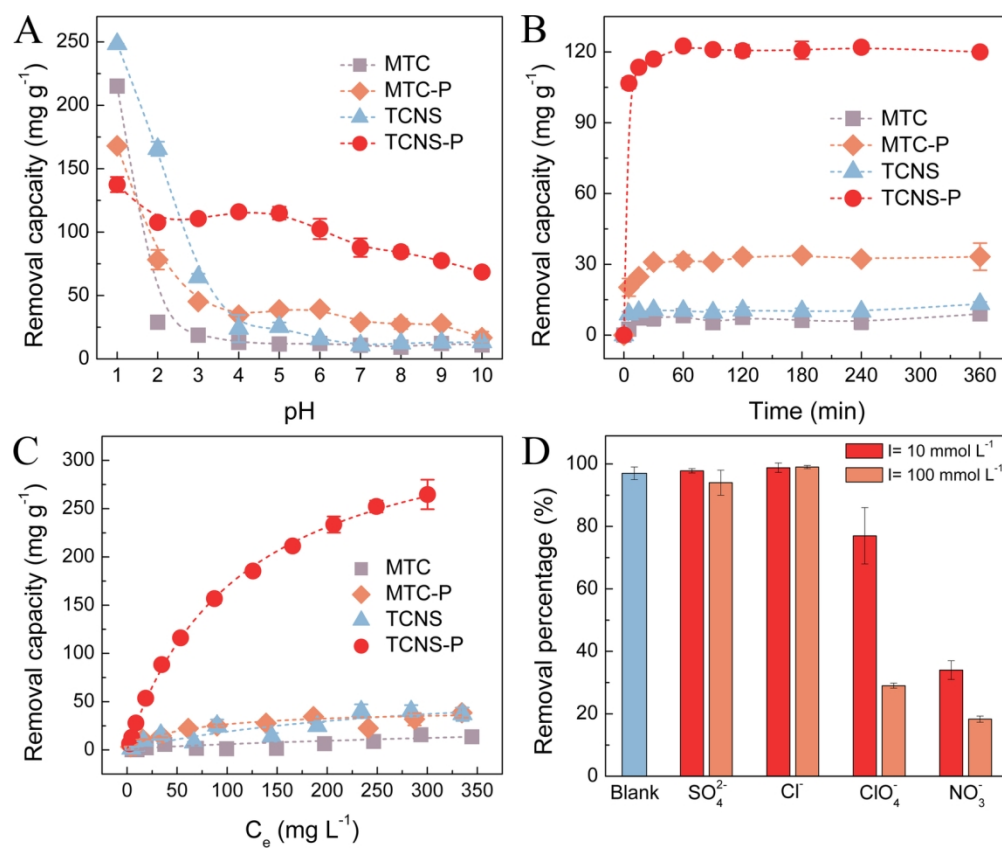
<sup>a</sup>Coordination number. <sup>b</sup>Radial distance. <sup>c</sup> Debye-Waller factor. <sup>d</sup> Energy shift relative to the calculated Fermi level. <sup>e</sup>Goodness-of-fit indicator. <sup>f</sup>Fixed during fitting.

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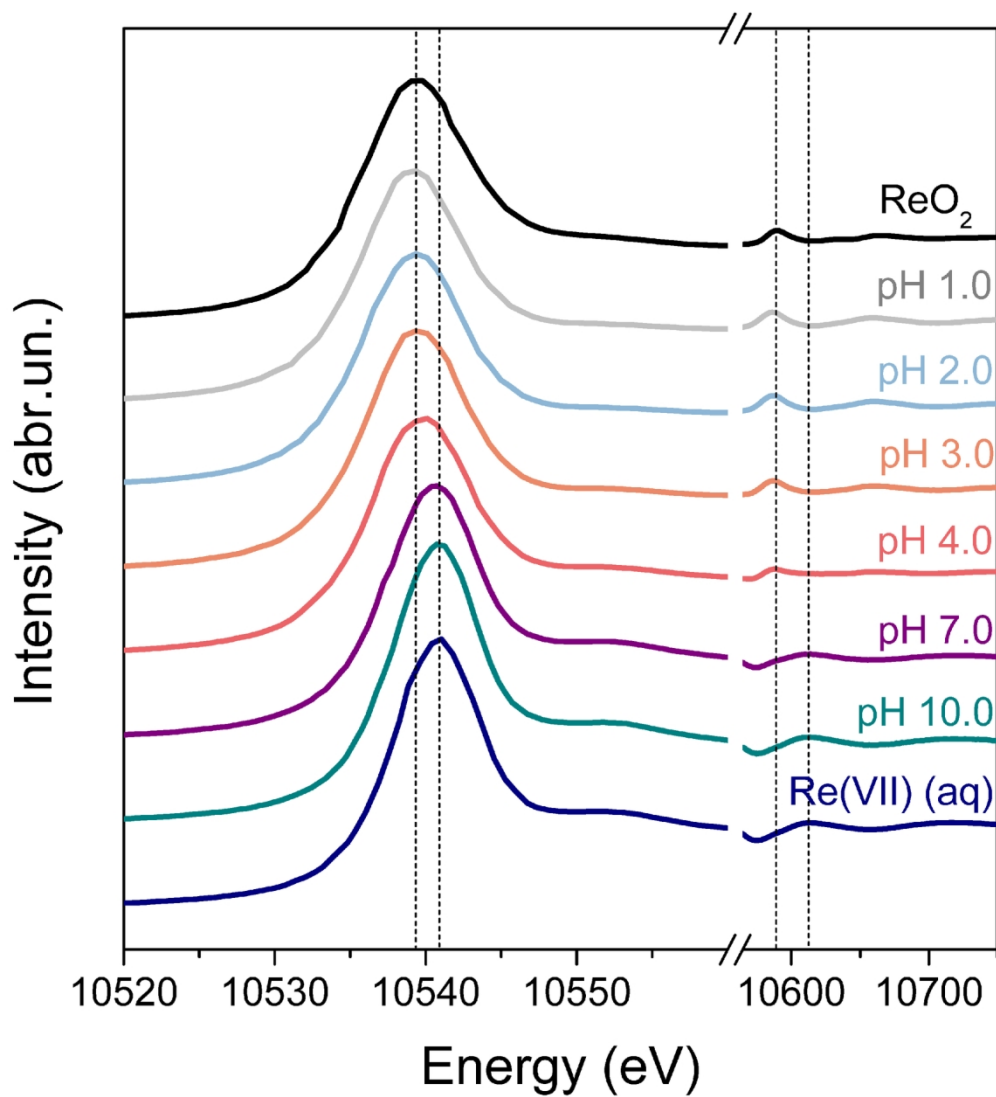
SEM images (A–D), powder XRD patterns (E), thermogravimetric weight-loss curves (F), and zeta potentials versus solution pH (G) of Ti<sub>2</sub>CTx and its nanocomposites.

146x120mm (300 x 300 DPI)



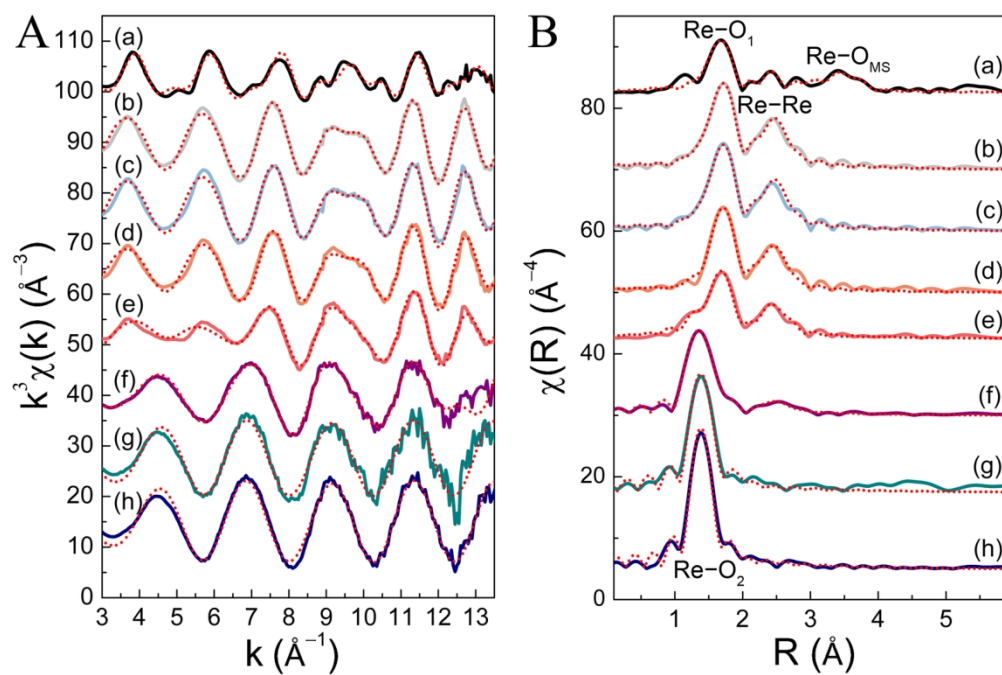
(A–C) Re(VII) removal from aqueous solution by Ti2CTx and PDDA hybrids as a function of pH, contact time and equilibrium Re(VII) concentration. (D) Effect of competitive anion species on Re(VII) removal by TCNS-P.

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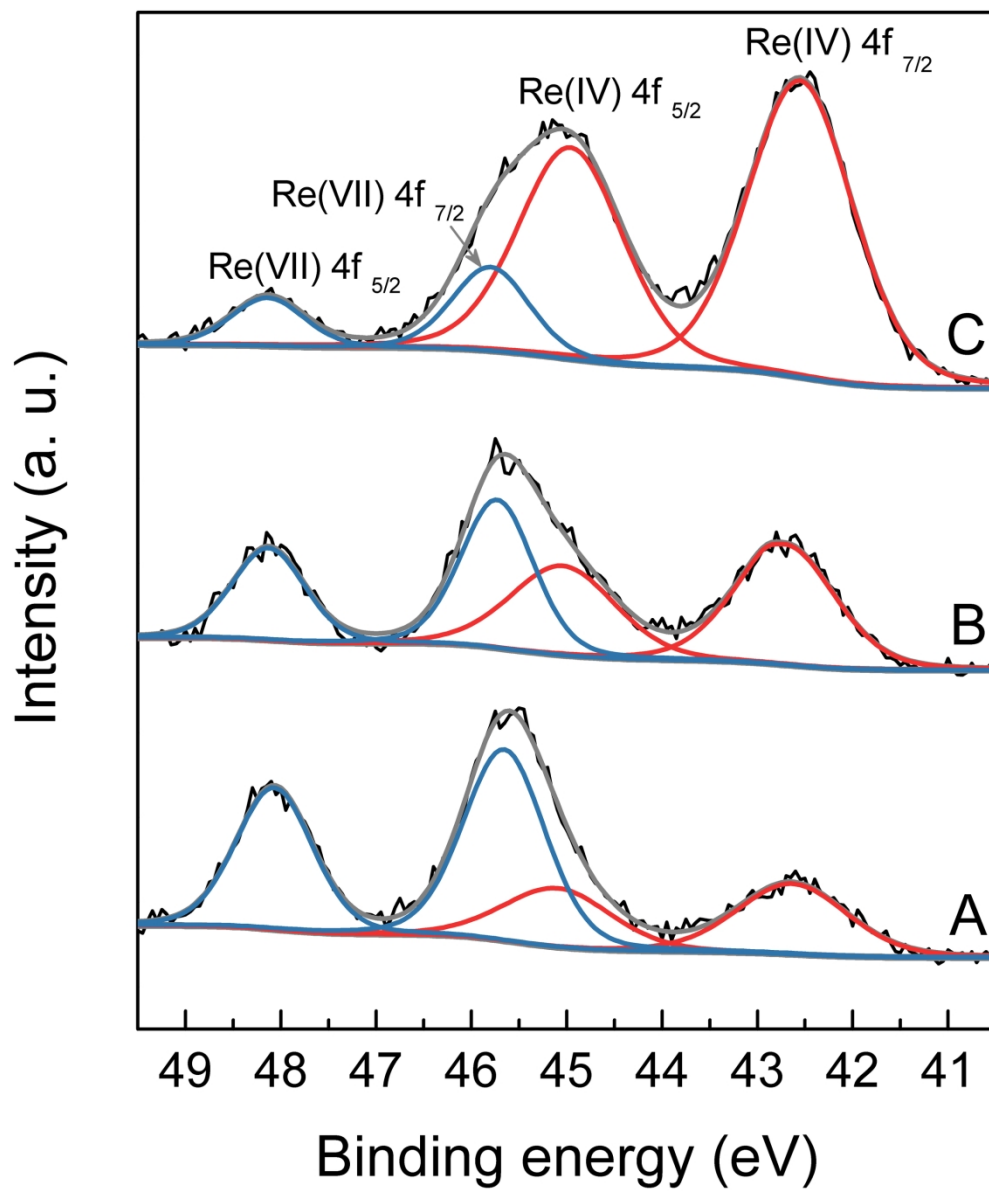
Re L3 edge XANES spectra of Re-loaded TCNS-P for different solution pH.

146x158mm (300 x 300 DPI)



(A) Re L 3 edge  $k^3$ -weighted EXAFS spectra (solid lines) and the best theoretical fits (dotted lines) for Re-loaded TCNS-P samples under different pH. (B) Corresponding non-phase shift corrected Fourier transforms.

146x97mm (300 x 300 DPI)



XPS spectra of Re-loaded TCNS-P for contact times at pH 3.0.

203x240mm (300 x 300 DPI)

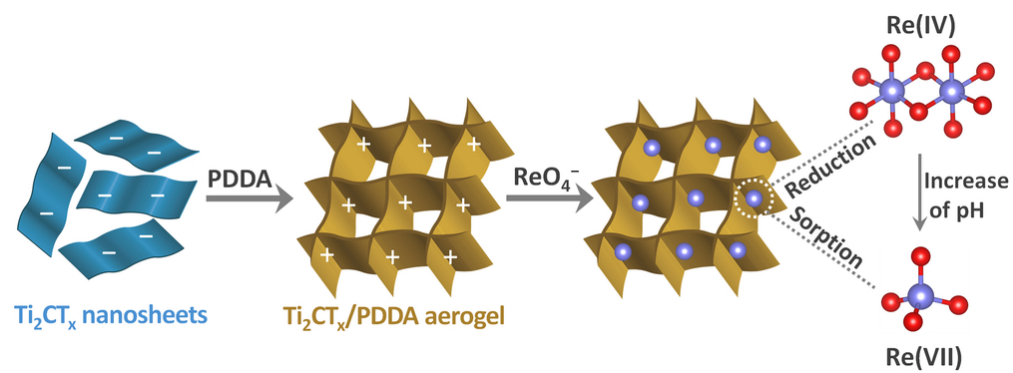


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