



BNL-223242-2022-JAAM

Remote Sensing from Unoccupied Aerial Systems: Opportunities to Enhance Arctic Plant Ecology in a Changing Climate

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To be published in "Journal of Ecology"

July 2022

Environmental and Climate Sciences Department
Brookhaven National Laboratory

U.S. Department of Energy

USDOE Office of Science (SC), Biological and Environmental Research (BER) (SC-23)

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1 **Journal:** Journal of Ecology, Grime Reviews

2 **Title:** Remote Sensing from Unoccupied Aerial Systems: Opportunities to Enhance Arctic Plant Ecology
3 in a Changing Climate

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32 **Figures:/Table Record:** 5 figures and 3 tables

33 **Word Count:** 8123

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54 **Remote Sensing from Unoccupied Aerial Systems: Opportunities to Enhance Arctic Plant Ecology in**
55 **a Changing Climate**

56 **Abstract:**

- 57 1. The Arctic is warming at a faster rate than any other biome on Earth, resulting in widespread changes
58 in vegetation composition, structure, and function that have important feedbacks to the global climate
59 system. The heterogeneous nature of arctic landscapes creates challenges for monitoring and improving
60 understanding of these ecosystems, as current efforts typically rely on ground, airborne, or satellite-
61 based observations that are limited in space, time, or pixel resolution.
- 62 2. The use of remote sensing instruments on small Unoccupied Aerial Systems (UASs) has emerged as
63 an important tool to bridge the gap between detailed, but spatially limited ground-level measurements,
64 and lower resolution, but spatially extensive high-altitude airborne and satellite observations. UASs
65 allow researchers to view, describe and quantify vegetation dynamics at fine spatial scales (1 – 10 cm)
66 over areas much larger than typical field plots. UASs can be deployed with a high degree of temporal
67 flexibility, enabling observation across diurnal, seasonal, and annual timescales.
- 68 3. Here we review how established and emerging UAS remote sensing technologies can enhance arctic
69 plant ecological research by quantifying fine-scale vegetation patterns and processes, and by enhancing
70 the ability to link ground-based measurements with broader-scale information obtained from airborne
71 and satellite platforms.
- 72 4. *Synthesis.* Improved ecological understanding and model representation of arctic vegetation is needed
73 to forecast the fate of the Arctic in a rapidly changing climate. Observations from UASs provide an
74 approach to address this need, however, the use of this technology in the Arctic currently remains
75 limited. Here we share recommendations to better enable and encourage the use of UASs to improve
76 the description, scaling, and model representation of arctic vegetation.

77

78 **Keywords:** Arctic, Fine-Scale, Landscape Ecology, Remote Sensing, Scaling, UAS, UAV, Vegetation

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83 1. Introduction

84 The Arctic is warming faster than any other region on Earth, with associated changes in temperature,
85 precipitation, surface albedo, sea ice, and ocean circulation (*IPCC*, 2019). Tundra ecosystems of the Arctic
86 are thus predicted to respond more rapidly to climate change than other terrestrial ecosystems (Chapin et
87 al., 2005; Hinzman et al., 2005). Over the past 40 years, long-term ecological monitoring and satellite
88 observations have indicated widespread changes in tundra vegetation composition, structure, and function
89 (Elmendorf et al., 2012; Pearson et al., 2013). Examples include a decadal ‘greening’ trend in satellite-
90 derived vegetation indices observed across the Arctic concurrent with a widespread increase in shrub and
91 tree cover (Elmendorf et al., 2012; Frost & Epstein, 2014; Ju & Masek, 2016; Myers-Smith et al., 2015;
92 Sturm et al., 2001). These changes, coupled with permafrost thaw (Lawrence et al., 2008), more frequent
93 tundra fires (Mack et al., 2011), and an altered hydrological cycle (Bring et al., 2016), are driving impacts
94 on the energy balance and carbon budget of the Arctic (DeMarco et al., 2014; McGuire et al., 2018; Myers-
95 Smith et al., 2011; Vowles & Björk, 2019).

96 The arctic tundra biome contains a high degree of spatial heterogeneity in vegetation distribution, land
97 surface structure, and environmental conditions (Myers-Smith et al., 2020; Virtanen & Ek, 2014; Fig. 1),
98 where vegetation interacts with the environment at very fine scales from several centimeters to multiple
99 meters (Assmann et al., 2020; Davidson et al., 2016; Siewert & Olofsson, 2020). These fine-scale
100 interactions result in strong patchiness in the direction and rate of vegetation changes across the Arctic that
101 are currently missed by coarser-scale observations but could aggregate to meaningful impacts on ecosystem
102 response to climate change (Bjorkman et al., 2018; Chen et al., 2020; Prevéy et al., 2018).

103 Traditional methods to characterize the variability in arctic vegetation involve intensive field surveys or
104 experimental manipulations, which are often limited in their spatial and temporal coverage (Metcalfé et al.,
105 2018; Schimel et al., 2015). Metcalfé et al. (2018) showed that the current pattern of ground sampling —
106 focused on just a few intensively studied locations — may inaccurately portray large-scale tundra processes,
107 hindering our ability to predict climate change impacts in the Arctic. In contrast, satellite and high-altitude
108 airborne remote sensing platforms have been widely used to observe a range of key vegetation properties,
109 dynamics, and changes (Beamish et al., 2020). The use of these platforms has complemented traditional
110 field measurements by providing wider spatial and temporal coverage (Shiklomanov et al., 2019).

111 However, the small-stature of most arctic plants (typically <1 m) and wide spatial variation in their
112 composition, structure, and function creates a strong scale-mismatch between the studied organisms and
113 the satellite and airborne observations, typically collected at >5 m spatial resolution (Assmann et al., 2020;
114 Siewert & Olofsson, 2020). This scale mismatch means that single pixels typically include a mixture of

115 information from many plant species (Somers et al., 2011; Wu & Li, 2009) and often non-photosynthetic
116 vegetation and other surface types including snow, surface water, bare soil, rock, and dead plant materials,
117 which makes the interpretation and ‘unmixing’ of the data difficult (Myers-Smith et al., 2020; Nelson et
118 al., 2022). In addition, due to differences in calibration, resolution, and revisit frequency, vegetation patterns
119 and phenology derived from different satellite platforms can show large discrepancies between each other
120 and with ground observations (Myers-Smith et al., 2020), introducing significant uncertainties in the
121 understanding of arctic vegetation dynamics. Therefore, to enhance our understanding of tundra vegetation
122 dynamics, measurements with a high spatial resolution and relatively broad scale are needed to bridge the
123 gap between traditional fine-scale ground sampling and broad-scale, low-resolution satellite images
124 (Myers-Smith et al., 2020).

125 To fulfill this need, small piloted airborne platforms have been used, which allows the collection of spatial
126 datasets at sub-meter resolutions, over large landscapes, and away from road systems (Nolan et al., 2015;
127 Heather et al., 2019; Cristóbal et al., 2021; Wainwright et al., 2021). However, piloted data collection is
128 often expensive, involves certified flight crews, and requires strategic flight planning (Chadwick et al.,
129 2020). In addition, weather conditions (e.g., wind, cloud, and rain) can change rapidly during the course of
130 a day in the Arctic, and conditions suitable for aircraft operation or remote sensing data collection
131 oftentimes last only from several minutes to a couple of hours (Assmann et al., 2019), challenging the use
132 of piloted airborne platforms.

133 The recent development and use of small unoccupied aerial systems (UASs; <25 kg) as a remote sensing
134 platform has revolutionized the way that ecologists quantify vegetation status and dynamics (Anderson &
135 Gaston, 2013; Assmann et al., 2019; Gaffey & Bhardwaj, 2020; Messina & Modica, 2020; Yao et al., 2019).
136 The use of remote sensing instrumentation on UASs has many advantages over traditional field sampling
137 or airborne/satellite platforms. For example, land surface observations can be easily obtained at a very high
138 resolution (1 - 10 cm), allowing for characterization of fine-scale details in a manner that closely mirrors
139 ground-based sampling but over larger spatial extents (Dainelli et al., 2021a, 2021b). Second, flight
140 missions can be deployed at a flexible time frame that is optimized to research objectives, such as capturing
141 the phenological cycle of target plants (D’Odorico et al., 2020) or repeatedly flying the same location over
142 the course of a day to capture diurnal vegetation dynamics or solar-induced fluorescence (e.g., SIF; Wang
143 et al., 2021). Third, aerosol and other atmospheric effects on remotely sensed imagery, that commonly
144 occur from high-altitude airborne and satellite platforms, can be largely avoided by flying UASs at low
145 altitude (Yao et al., 2019). Lastly, diverse types of remote sensing data are needed to describe the
146 composition, structure, and function of terrestrial vegetation; UASs can be used to collect these data using
147 a variety of sensors, such as optical red-green-blue (RGB) camera, multispectral and hyperspectral sensors,

148 thermal infrared (TIR) camera, and light detection and ranging (LiDAR, also commonly known as ‘lidar’)
149 sensors. Recent studies have shown that vegetation data collected with UASs can have a similar fidelity as
150 direct, in-situ measurements, demonstrating the potential of using UASs to characterize fine-scale
151 vegetation patterns and processes (Chang et al., 2020; Lucieer et al., 2014; Thomson et al., 2021; Yang et
152 al., 2020).

153 Thus far, the majority of UAS applications have focused on low-latitude ecosystems (Dainelli et al., 2021a,
154 2021b), but interest in using UASs in the Arctic has been steadily increasing. Some of the earlier examples
155 include Tømmervik et al. (2014), Juszak et al. (2017), and Fraser et al. (2016), where optical and
156 multispectral UASs were used to produce high-resolution maps of arctic plant species. Recently, other
157 remote sensing technologies have also been used with UASs in the Arctic, including spectroscopy
158 (Malenovský et al., 2017; Yang et al., 2020), TIR (Yang et al., 2020, 2021), and LiDAR (Collins et al.,
159 2020), all of which offer exciting new opportunities to advance research. Here, we review how established
160 and emerging approaches to UAS-based remote sensing can enhance arctic plant ecology research, increase
161 understanding of the fine-scale vegetation composition and function, and provide previously missing
162 information at a resolution that bridges the gap between traditional, ground-based measurements and broad-
163 scale, coarse resolution airborne and satellite remote sensing. Specifically, we first summarize the remote
164 sensing technologies that have been integrated with small UAS platforms and the key vegetation and surface
165 data that can be obtained through this integration. We then highlight some of the most impactful
166 applications of UAS-based remote sensing in plant ecology in the Arctic to date and provide examples of
167 how these data can be used to address ecological questions. Finally, we provide perspectives on the
168 remaining challenges that need to be addressed to extend and advance future UAS-based remote sensing in
169 the Arctic.

170 **2. UAS Remote Sensing Technologies**

171 **2.1 UAS Platforms**

172 The primary features that distinguish the variety of UASs available for use by the research community are
173 physical size, the sophistication of pilot control aids (e.g., auto hover, obstacle avoidance) and automatic
174 flight control systems (pre-planned, auto-piloted missions), and the power capacity (battery or fuel) which
175 limits the payload, operating altitude, and single flight duration (González-Jorge et al., 2017; Hardin et al.,
176 2019). Typically, a large UAS (e.g., >25 kg) is more capable of carrying heavy instrumentation and
177 covering large study areas, but their development and deployment is more expensive and requires complex
178 ground operations, as well as more rigorous pilot certifications. Here, we focus on small UASs (either fixed-
179 wing or copter) that can be easily transported and deployed in arctic environments and require simpler

180 remote pilot certifications for operation (e.g., in the United States, the Federal Aviation Administration Part
181 107 certification). These small UASs usually have a flight duration of <2 hours. Small UASs can fly safely
182 at lower altitudes (<100 m above ground level), enabling data collection at very high spatial resolutions
183 (<10 cm).

184 **2.2 UAS Sensors**

185 In conjunction with the increased interest in their use for high-latitude research, there has been a similar
186 increase in available sensors that can be deployed on small UAS platforms (Table 1). At present, the most
187 common sensor type employed with UASs is the standard RGB camera. However, more recent research
188 has started to utilize a range of other sensors, including multispectral cameras, imaging spectrometers,
189 thermal cameras, and LiDAR systems. Here, we refer interested readers to Colomina & Molina, (2014) and
190 Yao et al. (2019) for more details about these sensor types and their applications on UASs. The integration
191 of these state-of-the-art sensors with UASs allows researchers to remotely observe terrestrial vegetation
192 and land surfaces at very fine scales (<10 cm) and across the principal dimensions of plant biodiversity
193 (e.g., taxonomic, spectral, and structural; Dainelli et al., 2021a, 2021b).

194 A portfolio of vegetation and surface properties can be derived using the suite of instrumentation shown in
195 Table 1. These properties include species identity and plant functional type (PFT), vegetation structure
196 (e.g., canopy height and cover), plant traits, and phenology, as well as many other parameters that are useful
197 to depict vegetation distribution, energy balance, water cycling, and carbon sequestration (Table 2). Note
198 that this review focuses on vegetation applications; therefore, properties specific to other disciplines are not
199 discussed. Table 2 also specifies how each property can be derived from different data products created
200 using various remote sensing technologies. For example, high-density point clouds (a set of data points in
201 three-dimensional space) can be created from two main technologies: direct laser scanning (i.e., LiDAR)
202 and structure-from-motion (SfM) processing of optical RGB imagery (refer to Turner et al. (2012) and
203 Westoby et al. (2012) for more details about SfM). In turn, point clouds can be used to derive a number of
204 vegetation structure parameters, such as canopy height, cover, and biomass (Table 2; Wallace et al., 2016).

205 **3. Vegetation Applications of UAS-based Remote Sensing in the Arctic**

206 The use of remote sensing instrumentation with UASs spans a wide range of applications in the Arctic.
207 Here, we focus on two areas that can be particularly impactful for arctic ecology: 1) characterizing fine-
208 scale patterns in vegetation composition, structure, traits, and functions (sections 3.1 - 3.5); and 2) scaling
209 fine-scale vegetation patterns and processes to coarse-scale airborne or satellite platforms to enable a
210 broader-scale understanding of the Arctic (section 3.6; Fig. 1).

211 **3.1. Vegetation Composition and Diversity**

212 Temperature increase in the Arctic is driving species distributions to shift northward and to higher
213 elevations, altering historical biodiversity patterns that have strong links to ecosystem health and function
214 (Wang & Gamon, 2019; Wasowicz et al., 2020). Being able to characterize the spatial patterns and drivers
215 of change in arctic vegetation composition and diversity is important for forecasting how arctic ecosystems
216 will respond to climate change. However, the scale mismatch between the small arctic plants and the coarse
217 grain size of satellite or airborne observations (Davidson et al., 2016) results in significant biases in the
218 characterization of vegetation composition and biodiversity (Fig. 2; Gamon et al., 2019, 2020).

219 The primary advantage of using UAS-derived imagery for research into arctic vegetation composition and
220 diversity is its fine grain size, which allows individual plants or species to be identified. Early applications
221 that classify UAS optical RGB or multispectral imagery have proven useful for mapping a number of arctic
222 plant species or PFTs. For example, Fraser et al. (2016) classified and mapped nine tundra vegetation types
223 (i.e., willow, alder, birch, reindeer lichen, moss, sedge tussock, wet graminoid, and mixed dwarf shrub
224 heath) using RGB and height predictors. Recently, the use of imaging spectroscopy or the combined outputs
225 of multi-sensor UASs showed the potential to identify more tundra species and improve mapping accuracy.
226 For instance, Yang et al. (2021) used a combination of structural, spectral, and thermal information to
227 differentiate eight shrub species that are similar in spectral signatures but vary in canopy height and thermal
228 properties. That study also demonstrated that the use of TIR imaging helps identify species with unique
229 thermal characteristics, such as ‘hot’ canopy lichen species.

230 Together with high-resolution species or PFT maps, a number of key environmental parameters, including
231 terrain features (e.g., elevation, thaw slumps, and solifluction), surface water distribution (e.g., rivers,
232 drainages, thaw ponds), snow depth, soil moisture, and active layer depth (e.g., using synthetic aperture
233 radar) could be described at very high resolutions and across landscapes using UASs (Fraser et al., 2020;
234 Gunn et al., 2021; Xu & Zhu, 2018). This diversity of data can assist in analyzing vegetation distribution
235 patterns across fine-scale environmental gradients, reducing the need for intensive field measurements. For
236 instance, combining four years of UAS imagery with field surveys and time-series climate data, DelGreco
237 et al. (2018) characterized vegetation composition in a low-Arctic mire and found high heterogeneity in
238 vegetation dynamics, driven by permafrost thaw and associated increases in soil wetness. With the diverse
239 types of UAS data, specific species of interest (e.g., invasive species) can also be readily identified, which
240 is essential to understand the emerging impacts of species distribution or changes on vegetation composition
241 and diversity across arctic landscapes (Lucieer et al., 2014; Räsänen et al., 2020). Arctic vegetation
242 dynamics and landscape changes are increasingly driven by more frequent and severe disturbances such as
243 fire (French et al., 2015; McCarty et al., 2020) and rapid permafrost thaw or thermokarst (Jones et al., 2015;

244 Turetsky et al., 2019). Before-and-after UAS surveys can be used to study the impacts of disturbances on
245 arctic ecosystems and monitor post-disturbance vegetation recovery. Upscaling UAS data to link with time-
246 series satellite records could then enable the mapping of vegetation composition and status over time (see
247 *Section 3.6*). In a recent study by Siewert et al. (2021), the utility of UASs to capture herbivore (vole and
248 lemming) impacts on arctic vegetation composition and productivity was also investigated, which
249 represents an exciting new area of UAS application in the Arctic.

250 In addition to maps of plant species or PFTs, vegetation composition and diversity patterns, including
251 functional diversity, can also be inferred from analysis of spectral diversity using imaging spectroscopy
252 sensors installed on UASs (Rocchini et al., 2010, 2018; Wang & Gamon, 2019; Schweiger, 2018). This
253 approach assumes that genetic background and environment conditions result in differences in plant
254 physiology, biochemistry, and structure among individuals, species, lineages, or PFTs that are readily
255 expressed in spectral signatures (Ustin and Gamon et al., 2010; Cavender-Bares et al., 2020). The utility of
256 this approach has been demonstrated in a variety of biomes with spectral data collected from UASs (Baldeck
257 & Asner, 2013; Carlson et al., 2007), and has the potential to be as effective in the Arctic.

258 **3.2. Vegetation Structure**

259 There are a number of vegetation structural parameters that can be derived from UASs using point clouds
260 generated by either LiDAR or SfM (Table 1, Bergen et al., 2009). Here, our illustration focuses on canopy
261 height, vegetation cover, and biomass estimates that are needed to investigate the shrubification of arctic
262 ecosystems (Cunliffe et al., 2020; Greaves et al., 2015). Similar to high-resolution imaging, the main
263 interest of using UASs to quantify arctic vegetation structure is that point clouds can be generated at ultra-
264 high densities (>100 points/m²), a requirement to capture open-canopy, sparsely-distributed shrubs or to
265 penetrate dense, closed canopies as is necessary for constructing a reliable baseline digital elevation model
266 (DEM) and canopy height model (CHM; Fig. 3; Alonzo et al., 2020). For example, using UAS SfM, Fraser
267 et al. (2016) were able to obtain a point cloud density of $\sim 30,000$ points/m² at a low-arctic shrub tundra
268 landscape. Collins et al. (2020) explored the efficacy of UAS LiDAR for scanning arctic vegetation and,
269 while less dense than SfM, obtained point cloud densities of 300 - 500 points/m², which is more than 10
270 times as dense as typical airborne LiDAR point clouds (10 - 30 points/m²; Alonzo et al., 2020).

271 To apply UAS LiDAR or SfM for quantifying vegetation height, a series of filters must be applied to detect
272 data points returned from the bare ground surface (Andersen et al., 2003). Several methods exist for this
273 process, including the improved progressive Triangulated Irregular Network densification, but generally,
274 they combine highly automated processes with some manual corrections (Kilian et al., 1996; Kraus &
275 Pfeifer, 1998). The CHM is defined as the difference between the digital surface model (DSM) and a DEM

276 interpolated from the 'ground return' data points (Fig. 3). Fraser et al. (2016) and Yang et al. (2021)
277 validated the accuracy of UAS SfM for deriving tundra vegetation height against ground measurements in
278 high-Arctic (Tuktoyaktuk, Canada) and low-Arctic (Seward Peninsula, Alaska, USA) ecosystems, and
279 reported a RMSE of <0.11 m and <0.14 m, respectively. Alonzo et al. (2020) compared SfM-derived shrub
280 height with LiDAR installed on Goddard's LiDAR, Hyperspectral & Thermal (G-LiHT) Imager and found
281 excellent agreement between them (Pearson correlation coefficient = 0.89), indicating that both techniques
282 are appropriate for measuring low-stature arctic plants. However, it should be noted that detecting 'ground'
283 points in regions with dense graminoid cover (e.g., tussock tundra) could be challenging, as closed
284 graminoid canopies that might be tens of centimeters deep can cover the true bare ground (Wang et al.,
285 2016). In those regions, underestimations of shrub height may occur with either UAS LiDAR or SfM (Yang
286 et al., 2021).

287 Estimates of vegetation cover can be made using the fraction of point clouds (either LiDAR- or SfM-based)
288 returned from vegetation canopies, compared with non-vegetation returns (Lefsky et al., 2002; Nelson et
289 al., 1984). In some cases, the cover of different vegetation layers (e.g., tree, shrub, and grass layers) may
290 be derived by segmenting the point clouds or CHM into height classes that correspond to different
291 vegetation types. Similar to height, the detection of the ground surface (i.e., DEM) is an important aspect
292 of cover determination. If the ground surface elevation is overestimated, vegetation cover will be
293 underestimated, and vice versa. Point cloud density is another factor that influences cover determination.
294 A low point density could lead to either omission of small shrub canopies or an overestimation of cover
295 where gaps within dense shrub canopies cannot be detected (Fig. 3). Vegetation maps derived from UASs
296 can also be used to estimate the cover of total green vegetation (Riihimäki et al., 2019) or vegetation type
297 of interest (e.g., lichen; Macander et al., 2020; also see 3.1). This method is particularly useful in tundra
298 regions where the land surface is covered by a single vegetation layer, or the top layer is of interest.

299 The mapping of vegetation height and cover extends to many ecological applications in the Arctic. For
300 example, research into the impact of shrubification on tundra ecosystems involves a need for shrub height
301 which importantly determines vegetation-mediated feedbacks to climate warming, such as snow depth,
302 albedo, nutrient exchange, hydrology, and energy flux (Léger et al., 2019; Mekonnen et al., 2021; Myers-
303 Smith et al., 2011; Zhang et al., 2018). The estimate of vegetation cover, including lichens, is also useful
304 to understand the distribution and habitat of arctic herbivores, such as caribou (Joly et al., 2007, 2009). In
305 a recent study, Cunliffe et al. (2020) showed that aboveground biomass (AGB), another important structure
306 parameter needed for carbon cycle modeling, can be linearly estimated from canopy height ($R^2 = 0.92$), and
307 outperformed normalized difference vegetation index (NDVI; $R^2 < 0.23$), a commonly-used vegetation
308 index for estimating AGB (Berner et al., 2018; Reynolds et al., 2012). In fact, estimating AGB with LiDAR

309 or SfM has drawn extra interest for arctic research, as destructive ground sampling and laboratory
310 processing of AGB are extremely difficult in the remote Arctic (Cunliffe et al., 2020). In addition to canopy
311 height, a variety of point cloud-derived metrics and their combinations have also been tested to map AGB
312 (Alonzo et al., 2018, 2020). The performance of LiDAR and SfM was also compared by Alonzo et al.
313 (2020), and both showed a strong ability to predict AGB (SfM: $R^2 > 0.75$, RMSE $< 1.26 \text{ kg/m}^2$; LiDAR: R^2
314 > 0.65 , RMSE $< 1.48 \text{ kg/m}^2$).

315 One caveat to the use of UAS SfM is that SfM can only depict the outer surface of vegetation layers and
316 thus contains little or no information on sub-canopy vegetation or terrain in areas with dense canopies
317 (Lisein et al., 2013; Puliti et al., 2015). LiDAR can better penetrate dense canopies, but it often produces
318 less dense point clouds than SfM, typically lacks multispectral information, and is also relatively expensive
319 to collect (Collins et al., 2020). In addition, it is noted that plant species vary in their biomass allometry
320 (Berner et al., 2015), which may bias structure-based AGB estimates. Integrating structure and spectral
321 information into models has been shown promising for improving estimates of AGB (Alonzo et al., 2020),
322 and suggests that collection of different UAS data types at the same locations may be needed to enhance
323 the value of UAS remote sensing in the Arctic.

324 **3.3. Plant Traits**

325 Plant traits are key attributes of plant canopies or leaves that generalize the morphological, biochemical,
326 phenological, and physiological characteristics of an individual, a species, or a PFT (Cornelissen et al.,
327 2003; Violle et al., 2007). These attributes are one of the primary controls on the distribution and function
328 of plants and therefore underpin many vegetation-climate interactions (Myers-Smith et al., 2018; Reich,
329 2014). For example, traits related to the uptake and allocation of resources, such as leaf mass per area, leaf
330 longevity, and foliar nutrient content, play a key role in the regulation of plant growth rate, primary
331 productivity, and decomposition rates (Cornwell et al., 2008; Diaz et al., 2004; Lavorel & Garnier, 2002).
332 Similarly, canopy structural traits, such as leaf area and plant height, influence competition and light-
333 harvesting potential, as well as surface albedo, and are strong determinants of plant biomass and snow
334 dynamics that in turn, determine surface energy and water balance (Callaghan et al., 2004; Sturm, 2005).

335 Remote sensing, and spectroscopy in particular (Table 1 & 2), has been shown to provide an effective
336 method to remotely estimate a host of leaf and canopy traits across agricultural, grassland, and forest
337 ecosystems (Dahlin et al., 2013; Serbin & Townsend, 2020; Singh et al., 2015; Wang et al., 2019). This is
338 because spectroradiometers can measure very-high-resolution spectral reflectance across a large number of
339 narrow, near-contiguous wavebands (Gamon et al., 2019; Ustin et al., 2004) that allow for detection of the
340 subtle absorption features of biochemical and structural properties in leaves and canopies (Curran, 1989;

341 Kokaly et al., 2009; Fig. 4). Typically, these mapping efforts based on imaging spectroscopy rely on direct
342 connections between field plot sampling and remote sensing pixels (e.g., Singh et al., 2015). However, in
343 the Arctic, this direct plot-to-pixel scaling is much more challenging. The integration of spectroscopic
344 sensors with UASs is an ideal tool for this aim (Shiklomanov et al., 2019). By acquiring very-high-
345 resolution, cloud-free hyperspectral imagery (Fig. 4a) over landscapes, spectroscopic sensors on UASs
346 allow for a more direct connection between the sampled vegetation and pixel reflectance, which can be
347 used to develop trait scaling approaches and maps that can be used to train larger-scale trait models for
348 airborne or satellite sensors (Thomson et al., 2021).

349 There are a number of approaches that can be used to predict traits from optical and hyperspectral UAS
350 sensors. These methods range from relatively simple spectral vegetation indices (SVIs) which use the ratio
351 of two or more spectral bands to infer plant stress, phenology, species diversity, or pigment composition
352 (Gamon et al., 1997; Goward & Huemmrich, 1992; Mänd et al., 2010; Schweiger, 2020), to more complex
353 machine learning and latent variable methods, such as partial least squares regression (PLSR; Geladi &
354 Kowalski, 1986; Wold et al., 2001), that are used to link pixel reflectance spectra with the underlying traits
355 of interest (Burnett et al., 2021; Serbin & Townsend, 2020; Wang et al., 2021). It is noted that spectral data
356 present a high level of collinearity among wavelengths (Chen et al., 2011); latent variable methods like
357 PLSR are effective at handling this issue by projecting the large number of predictors (i.e., reflectance at
358 different wavelengths) to a small number of latent variables and, at the same time, maximize the correlation
359 between the response and latent variables. The inversion of radiative transfer models (RTMs), including
360 PROSAIL (Jacquemoud et al., 2009), present another method to infer plant traits using spectral data (Féret
361 et al., 2011) based on semi-mechanistic links between leaf properties, canopy structure, sun-sensor
362 geometry, and the resulting reflectance signature of leaves and canopies (Kuusk, 2018; Ollinger, 2011).
363 Using imaging spectroscopy from UASs together with RTMs could allow for fine-scale retrieval of some
364 key traits without requiring in-situ calibration data (Shiklomanov et al., 2019), as well as simulating and
365 testing retrievals across a range of sensor types (Shiklomanov et al., 2016).

366 In addition to spectroscopy, other technologies (e.g., LiDAR and TIR) can also be used to predict or
367 improve the prediction of plant traits. For example, combining thermal data with spectroscopy can help
368 better predict biochemical and physiological traits that affect or are affected by leaf temperature (Bishoyi
369 & Sudhakar, 2017; Maimaitijiang et al., 2017). The integration of spectroscopy with point cloud-derived
370 metrics (e.g., canopy height) is also useful to predict traits associated with vegetation structure (Ewald et
371 al., 2018). Also, being able to simultaneously map vegetation traits, temperature, and structure at a high
372 spatial resolution can aid understanding of the response of plant traits to thermal changes among species
373 and across tundra landscapes, as well as the vertical profile of plant traits within shrub or tree canopies.

374 However, it is noted that UAS-based remote sensing is not a “silver bullet” for trait collection, given its
375 limited spatial coverage. Trait models developed at a specific site with a particular set of species may not
376 be readily extrapolated to other regions or different research objectives (Burnett et al., 2021). The
377 development of generalized trait models, i.e., those that work across a wide range of species and
378 environments, is an ongoing area of research (e.g., Serbin et al., 2019; Yan et al., 2021; Wang et al., 2020;
379 Schweiger et al., 2020), and such models could aid the application and iterative improvement of trait
380 mapping in the Arctic using UASs. In addition, linking UAS data with hyperspectral data collected from
381 airborne or satellite platforms to develop trait models and estimates at larger scales represents a unique
382 opportunity to advance pan-Arctic retrieval of plant traits using remote sensing. This opportunity could
383 substantially benefit from the ongoing and forthcoming spectroscopy missions, such as those associated
384 with NASA’s Arctic - Boreal Vulnerability Experiment (ABOVE) and Surface Biology and Geology (SBG)
385 mission (Cawse-Nicholson et al., 2021).

386 **3.4. Vegetation Stress and Thermal Function**

387 Temperature is fundamentally important to a wide range of vegetation and ecosystem processes (Berry &
388 Bjorkman, 1980; Chapin, 1983; Körner, 2006). The temperature of plant leaves (T_{leaf}) and canopies (T_{canopy})
389 directly influences a variety of processes, including the rate of enzyme-catalyzed reactions, membrane
390 fluidity, and the diffusion and solubility of CO_2 and O_2 , which together control the rate of photosynthesis
391 and respiration and, subsequently, the short-term and chronic responses of plants to changes in their
392 environment (Jones, 1992; Still et al., 2019, 2021). Therefore, characterizing T_{leaf} and T_{canopy} is particularly
393 useful for investigating vegetation function and health, and quantifying terrestrial vegetation responses to
394 climate change (Gersony et al., 2016; Krishna et al., 2021; Westermann et al., 2011; Yan et al., 2020). In
395 terms of surface temperature variation, T_{canopy} is typically quantified with remote sensing platforms that
396 retrieve land surface temperature (LST, Table 2) using a TIR sensor or camera (Table 1).

397 There is a rich history of using TIR imagery to quantify temperature variation across managed or natural
398 ecosystems and to assess plant-environment interactions (Costa et al., 2013; Jones & Leinonen, 2003). We
399 refer interested readers to Krishna et al. (2021) and Still et al. (2019 & 2021) for a review of the theory and
400 general applications of TIR. Here we focus on the use of TIR sensors on UASs for studying arctic ecology,
401 plant function, and ecological scaling. The heterogeneity of arctic ecosystems is mirrored by a large spatial
402 variation in LST and energy balance properties (Dietrich & Körner, 2014; Scherrer & Körner, 2009; Yang
403 et al., 2021). For example, in a ground-based study using a thermal camera, Scherrer & Körner, (2009)
404 detected a surface temperature variation of up to 20 °C along a 100-m subarctic hillslope during clear-sky,
405 mid-summer days. There is also a strong seasonality in LST and surface energy exchanges in the Arctic

406 (Westermann et al., 2011) that has a strong control on regional to global climate feedback (Chae et al.,
407 2015; Zhang et al., 2018).

408 The high degree of spatial and temporal variation in LST means that traditional remote sensing platforms
409 (>60 m resolution, e.g., Landsat) may not adequately characterize the fine-scale variation in LST and its
410 drivers across arctic landscapes, and thus leads to a mischaracterization of underlying surface biophysical
411 changes in response to warming conditions (Soliman et al., 2012; Westermann et al., 2011). In particular,
412 the mixing of different plant species and non-vegetation surfaces (e.g., rocks, soil, water, snow) in coarse-
413 resolution satellite pixels makes it challenging to interpret the biological and environmental controls on
414 LST (Cable et al., 2016). UAS-borne TIR could fill this gap by providing data at spatial resolutions that are
415 high enough to capture LST variation across different surface components, but at the same time, can be
416 repeated with a flexible time frame to capture LST dynamics (Simpson et al., 2021). In Fig. 5, we show an
417 example of UAS-collected LST at a low-arctic tundra site. Even in this small landscape of ~1 ha, a high
418 degree of variation in T_{canopy} (~5 - 20 °C) is observed, strongly tied to fine-scale patterns within and across
419 plant community types (Breen et al., 2020). The description of these fine-scale details will allow researchers
420 to link the variation in LST with surface and belowground features (e.g., vegetation type, non-vegetation
421 components, soil moisture, permafrost, and disturbance gradients) to better understand the patterns and
422 drivers of the spatiotemporal variation in LST (Yang et al., 2021), and to scale up or disaggregate coarse
423 remote sensing signals to improve large-scale monitoring and modeling efforts in the Arctic (Lara et al.,
424 2020).

425 In addition to characterizing fine-scale patterns and drivers of LST, linking UAS-collected LST with other
426 observations could yield important new insights on the relationships between arctic ecosystem functioning
427 and vegetation patterns. For example, quantifying the difference between T_{canopy} and T_{air} is useful for
428 diagnosing spatial and/or temporal patterns in plant thermal regulation (Dietrich & Körner, 2014; Jin &
429 Dickinson, 2010; Novick & Katul, 2020; Scherrer & Körner, 2009; Zhang et al., 2020). In the Arctic,
430 considerable uncertainties remain in our understanding of thermal regulation, the role of plant stature and
431 aerodynamics in energy cycling, and how future warming could impact plant function and fitness (Scherrer
432 & Körner, 2009; Lawrence & Swenson 2011; Bhatt et al. 2017; Aalto et al. 2018). In a recent study using
433 a UAS-borne TIR, Yang et al. (2021) found strong spatial variation in thermal decoupling across shrub
434 tundra landscapes that varied by PFT, patterns that have important implications for biodiversity, energy
435 balance, permafrost thaw of arctic ecosystems (Myers-Smith et al., 2011; Zhang et al., 2018). For example,
436 the T_{canopy} of deciduous tall shrubs was found to be significantly lower than other PFTs (e.g., moss, lichen,
437 graminoid, and low-lying shrubs) and typically below T_{air} , which leads to localized cooling during the
438 growing season and exerts negative feedback to permafrost thaw (Blok et al., 2010; Frost et al., 2018) and

439 plant diversity (Yang et al., 2021). Furthermore, using UAS-borne TIR in the footprint of eddy covariance
440 towers could improve the characterization and scaling of surface energy exchanges and water cycling from
441 site to ecosystem or biome scale (Ellsäßer, 2020; Hoffmann et al., 2016).

442 It is also worth mentioning that warmer and drier conditions in the Arctic are likely to increase plant stress,
443 either directly through increased water stress or indirectly by prompting insect pest disturbances (Bjerke et
444 al., 2014). Early detection of such stresses with satellite platforms is often complicated by an initial patchy
445 landscape response, which signifies a potentially important and increasing role of UAS-borne TIR for
446 identifying plant water or physiological stress in the Arctic (Still et al., 2019). In many other ecosystems
447 (e.g., agriculture and forest), T_{canopy} , $T_{\text{canopy}} - T_{\text{air}}$, and thermal stress indices (e.g., canopy water stress index)
448 derived from UASs have all been used to identify fine-scale water stress or insect pest outbreaks that create
449 thermal anomalies in T_{canopy} (Costa et al., 2013; Jones & Leinonen, 2003; Krishna et al., 2021). The
450 combination of spectroscopy with TIR may further improve the fidelity of this practice (Jones & Schofield,
451 2008; Krishna et al., 2021; Maimaitiyiming et al., 2020), aiding in the management of essential natural
452 resources in the Arctic. For example, the retrieval of plant traits from spectroscopy may allow researchers
453 to examine the impact of invasive insects on different plant species, especially where insect foraging
454 preference is strongly determined by plant biochemical traits (Wang et al., 2020). In addition, research on
455 arctic ecophysiology (e.g., photosynthetic and stomatal response to temperature) also requires knowledge
456 of the thermal and reflectance characteristic of arctic plants (Nelson et al., 2022; Chapin et al., 2012), which
457 can be simultaneously obtained from UASs.

458 **3.5. Vegetation Seasonality and Phenology**

459 In a recent review, Myers-Smith et al. (2020) discussed the extraordinary complexity of capturing plant
460 seasonality and phenology (e.g., leaf emergence, development, senescence, and abscission; Schwartz, 2013)
461 in the Arctic, involving spatial heterogeneity, scaling processes, and data availability, and highlighted
462 persistent challenges to addressing this complexity with traditional remote sensing. Particularly, the low
463 revisit frequency (>8 days) and the strong cloud and fog contamination (Tjernström et al., 2015) curtails
464 the acquisition of a suitable amount of clear-sky satellite imagery needed to establish a robust time-series
465 observation for the curve fitting or thresholding used to derive phenometrics (Gu et al., 2009). Not
466 surprisingly, seasonal phenological patterns and decadal trends derived from different satellite platforms
467 often do not align with each other (Myers-Smith et al., 2020).

468 The utility of UASs for studying vegetation seasonality and phenology lies in the ability that flight missions
469 can be easily repeated with a flexible time frame with revisit time optimized to the phenological cycle of
470 target species (Anderson & Gaston, 2013; Getzin et al., 2012). This flexibility is particularly useful in the

471 context of arctic environments, as frequent revisits can be made in spring and autumn (key phenological
472 stages) to counter the influence of cloud and fog on data availability. In a recent study, Assmann et al.
473 (2020) explored the use of a multispectral sensor on a UAS for capturing seasonal dynamics in NDVI and
474 revealed high spatial heterogeneity in tundra greenness and phenology not captured by satellites. In
475 particular, a notable loss in the seasonal variation of NDVI was observed when grain size increased from
476 ultra-fine UAS (5 cm) to medium-size satellite pixels (30 m), highlighting a need to investigate
477 spatiotemporal scaling processes in arctic plant seasonality and phenology.

478 Notably, the collection of reliable time-series UAS observations could face practical challenges in the
479 Arctic. Site access can be limited in early spring due to snow coverage. Optimal imaging conditions also
480 are rarely present with fluctuating cloud cover, rain, and wind conditions within a day and throughout the
481 year, all of which require careful calibration among UAS flights (Assmann et al., 2020). Thus far, limited
482 work has been done with UAS remote sensing of plant seasonality and phenology in the Arctic.
483 Nevertheless, as arctic researchers begin to use UASs (e.g., the High Latitude Drone Ecology Network,
484 HiLDEN), time-series observations will be more readily available. Moreover, given the various types of
485 instrumentation that can be mounted on a UAS, repeated observations provide an opportunity to expand
486 seasonality and phenology studies beyond simple vegetation indices which have shown limitations in
487 tracking arctic phenology (Assmann et al., 2020; Myers-Smith et al., 2020; Wang et al., 2018). For example,
488 the seasonal changes in leaf pigments, functional traits, canopy structure, and thermal properties can all be
489 potentially explored with UAS spectroscopy, LiDAR, and thermal imaging (D'Odorico et al., 2020; Keenan
490 et al., 2014; Liu et al., 2015; Still et al., 2019). Furthermore, given the high resolution of UAS data,
491 phenological diversity across key plant species and environmental gradients can be investigated. This is
492 important for predicting the range dynamics of arctic vegetation under future climate change, as plant
493 phenology importantly controls plant survival (Chuine, 2010) while being highly sensitive to micro-climate
494 and varying strongly across plant species (Andresen et al., 2018; Collins et al., 2021; Prevéy et al., 2018).

495 **3.6. Ecological Scaling**

496 Typically, airborne and satellite platforms provide excellent seasonal and long-term monitoring of
497 ecosystems, but they are limited in the ability to identify underlying surface processes (Anderson, 2018;
498 Lechner et al., 2012). In contrast, ground-based measurements provide detailed information on vegetation
499 structure, composition, and dynamics, but are limited in spatial extent given most observations are point or
500 plot measurements and come from only a few specific regions (Metcalf et al., 2018; Schimel et al., 2015;
501 Siewert & Olofsson, 2020). This mismatch between the scales (both spatial resolution and extent) of in-situ
502 and airborne/satellite observations, plus current sampling biases, makes it challenging to scale, map, and

503 describe broad-scale vegetation changes in the heterogeneous Arctic (Assmann et al., 2020; Myers-Smith
504 et al., 2020; Siewert & Olofsson, 2020; Yang et al., 2021).

505 UAS remote sensing offers unique opportunities to bridge this scale gap (Siewert & Olofsson, 2020). In
506 sections 3.1-3.5, we illustrated that UAS data can capture many sources of vegetation and surface
507 heterogeneity that are present in arctic ecosystems, providing a tool to depict fine-scale vegetation patterns
508 and processes, similar to traditional ground surveys but over larger spatial extents. These fine-scale details
509 enable an easy connection with in-situ vegetation surveys to extrapolate ground observations over larger
510 areas, which in turn facilitates landscape-scale understanding of vegetation dynamics in the Arctic. For
511 example, Siewert & Olofsson, (2020) showed that NDVI derived from ground measurements in the
512 northern Arctic agrees better with estimates derived from UAS imagery with a 12 cm resolution ($R^2 = 0.89$)
513 than satellite imagery at 10 m, 30 m, and 250 m resolutions ($R^2 = 0.2, 0.16, 0.01$, respectively). This scale-
514 dependency is propagated to strongly influence the estimation of AGB and GPP (gross primary
515 productivity) using NDVI (e.g., % bias of estimated AGB: 17.0% 30 m, 21.0% 75 m).

516 The integration of ground observations with UAS data can expand the spatial extent of ecological studies,
517 but UASs are not intended nor expected to collect observations over the large areas needed for monitoring
518 vegetation changes across the Arctic (Myers-Smith et al., 2020). However, UAS data can be a useful proxy
519 of ground measurements to further calibrate models built with airborne or spaceborne observations to assess
520 larger-scale ecological patterns, which represents an exciting future research opportunity in the Arctic
521 (Myers-Smith et al., 2020). For example, Thomson et al. (2021) explored the feasibility of using trait maps
522 developed with a multispectral UAS to upscale plant water content to the wider landscape using Sentinel-
523 2A imagery (10 m). Similarly, Riihimäki et al. (2019) used UAS maps as training data and built SVI-based
524 models to estimate green vegetation cover from Planet Cubesat (3m), Sentinel-2A (10 m), and Landsat 8
525 OLI (30 m).

526 Integrating UAS data with satellite and airborne imagery can also help improve our mechanistic
527 understanding of the links between fine-scale vegetation dynamics and broader-scale ecological patterns
528 and trends (Myers-Smith et al., 2020). For example, using datasets collected with a multi-sensor UAS, Yang
529 et al. (2021) showed that landscape-scale variation in vegetation thermoregulation and canopy structure is
530 largely driven by PFT composition, as well as trait variation within each PFT. By linking vegetation
531 properties with UAS data at different scales, the scale at which dynamic processes occur and the drivers of
532 large-scale variation can also be determined (Assmann et al., 2020; Siewert & Olofsson, 2020). For
533 example, Assmann et al. (2020) captured plant growth dynamics across tundra landscapes by investigating
534 the seasonal change in tundra NDVI with a multispectral UAS. They identified that a resolution of ~50 cm
535 is the optimal grain size for monitoring arctic greening in dryas-vetch and tussock-sedge communities and

536 showed a loss of seasonal variation in the spatial heterogeneity of landscape greenness when aggregating
537 from UAS pixels (50 cm) to medium-grained satellite pixels (10 – 30 m). These types of applications present
538 new opportunities to identify how fine-scale vegetation and surface heterogeneity influences the
539 spatiotemporal patterns of coarse remote sensing signals, improving our ability to interpret and describe
540 remotely sensed changes across the Arctic (Myers-Smith et al., 2020).

541 **4. From Describing Vegetation to Answering Ecological Questions**

542 How does UAS data can be used to improve ecological approaches and address fundamental ecological
543 questions pertinent to the Arctic? Ecologists seek to understand how organisms (i.e., plants in this Review)
544 interact with each other and their abiotic environment (Sutherland et al., 2013; Tansley, 1935). In the Arctic,
545 this objective is typified by the need to quantify, understand, and predict how vegetation is responding to a
546 changing climate and the impact of these changes on the larger arctic biome. However, presently, our ability
547 to address this objective has been limited by significant data and knowledge gaps that hinder our ecological
548 understanding of the Arctic and increases model uncertainty associated with predicting the fate of the Arctic
549 (Fisher et al., 2018; Metcalfe et al., 2018).

550 In Table 3, we summarize some of the most pressing ecological questions that could be potentially
551 addressed through the use of UASs. These questions span five key research areas, including fine-scale
552 vegetation and surface heterogeneity, shrubification, arctic ‘greening’, disturbance, and process model
553 uncertainty (Fisher et al., 2018; Mekonnen et al., 2021; Myers-Smith et al., 2011; Rogers et al., 2022). For
554 each ecological question, we specify the measurement need to address the question and how UASs can fill
555 this need by using the technologies or data presented in *Section 3*.

556 One key aspect of arctic ecology research that UAS remote sensing could revolutionize is the
557 parameterization and benchmarking of process models used to simulate arctic vegetation. These are key
558 steps to reducing model uncertainty and are required for robust prediction of change in the Arctic (Fisher
559 et al., 2014 & Fisher et al., 2018). In process models, the diversity of plant species and their traits are
560 typically binned into PFTs. However, the current classification of tundra PFTs has focused on a few primary
561 classes, e.g., evergreen and deciduous shrubs, graminoids, forbs, moss, and lichen (Wullschleger et al.,
562 2014). The parameterization of these PFTs also largely relies on scant ground measurements or assumptions
563 from temperate species (Rogers et al., 2017, 2019), which leads to significantly higher model uncertainties
564 in the Arctic than other biomes. Here, we expect that as its applications extend in the Arctic, measurements
565 from UASs could play an important role in filling this gap. For example, the detailed identification and
566 mapping of PFTs with UASs that consider functional and structural differences across plants can be directly
567 used to inform landscape-scale models; these fine-scale classifications can also be scaled up to create
568 vegetation composition maps using airborne or satellite platforms to inform regional or biome-scale models.

569 Similarly, the diverse and species-specific structure, traits, and function that can be derived from UASs
570 provide a relatively easy avenue to parameterize PFTs, given sufficient UAS data are collected in key
571 locations in the Arctic or methods are developed to connect UAS information to broad-scale remote sensing
572 data (e.g., Thomson et al., 2021).

573 **5. Perspectives, Challenges, and Future Directions**

574 In an era of unprecedented change in the Arctic, understanding plant responses to novel environmental
575 conditions at local, watershed, and larger scales is essential for our capacity to forecast the fate of these
576 ecosystems (Fisher et al., 2014; Mekonnen et al., 2021; Rogers et al., 2022). In sections 3 and 4, we
577 highlighted some of the most impactful applications of UASs in the Arctic, and below we lay out the next
578 key steps, as well as the persistent challenges facing widespread use of UASs in the Arctic.

579 **5.1. Challenges and Caveats of Flying UASs in the Arctic**

580 The short growing season, typically characterized by weather that is unsuitable for UAS operations, means
581 that flying a UAS in the Arctic is particularly challenging. To minimize variation in solar angle, flying
582 around solar noon is usually suggested (Assmann et al., 2019). However, in some areas, windy conditions
583 can hamper effective flight control, increase battery usage, delay or ground UAS flights. On par with these
584 logistics challenges, technical issues, such as aircraft material failure, compass issues, and software failure,
585 are also not uncommon in the Arctic – all of which significantly increase time and cost while challenging
586 UAS flight planning and operation (Assmann et al., 2019).

587 To maximize UAS data coverage and impact on larger-scale research, it is important to optimize flight
588 plans and operations. In practice, this can be challenging as small UASs tend to have short flight times
589 (usually 15-20 minutes). In order to cover a large study site, successive flights are often needed, which rely
590 on consistent weather conditions and multiple battery sets. The general lack of accessible power at remote
591 locations means that most researchers will either need a large number of batteries, a generator for on-site
592 recharge, or both. In the future, these challenges may be mitigated by new battery technology, more efficient
593 electric motors and controllers, as well as new platform designs or options (e.g., fixed-wing vs copter) that
594 may be able to extend flight times for specific mapping missions. The increased development and use of
595 multi-sensor UASs would also largely reduce the complexity of obtaining multiple data types by facilitating
596 simultaneous collection (Yang et al., 2020).

597 Another key consideration when leveraging UASs for ecological applications is the high spatial resolution
598 of the data. While this is a primary benefit of UAS observations, it also raises challenges for image analysis.
599 Raw UAS imagery may suffer from “salt-and-pepper” noise artifacts (an impulse type of noise that

600 commonly exists in high-resolution images, especially when the pixel size is smaller than the studied object;
601 Azzeh et al., 2018). Given this, most studies exploring ecological patterns and processes should use object-
602 based methods, like species mapping which requires the consideration of object size (Yang et al., 2020,
603 2021). When stitching multiple flights together is required to cover a large study area, it is also important
604 to conduct calibration or standardization to ensure consistency across different times of day, illumination
605 conditions, or changes unrelated to changes in the surface properties being observed (Hakala et al., 2018).
606 The use of ground control points may also be needed to connect datasets from multiple flights and confirm
607 correct geolocation.

608 **5.2. Challenges and Next Steps for UAS Data Processing and Sharing**

609 The spatial resolution of UAS imagery creates significant challenges related to data volume and processing.
610 At high resolutions (<10 cm), even flights over small areas (e.g., 200 × 300 m) using a standard RGB
611 camera can generate data in excess of tens of gigabytes (Wyngaard et al., 2019). UAS flights are often
612 conducted as successive overlapping missions to cover a study area (Gillan et al., 2021), and in some cases
613 will include data collected from multiple instruments simultaneously (e.g., Yang et al., 2020). To store,
614 process, and use these data, large local or cloud storage, fast disc access, and high-performance computers
615 are usually needed.

616 Different UAS platforms and sensors may also require different software and workflows to post-process
617 data. For basic SfM processing, commercial software packages are commonly used, and while these
618 software packages have rapidly evolved to provide reasonably efficient processing, the memory and storage
619 requirements may still exceed standard end-user computers, and these software can be expensive.
620 Alternatively, open-source platforms for SfM processing have become more common, including
621 OpenDroneMap (<https://www.opendronemap.org/>), which can be run using a web interface, where
622 processing jobs can be submitted to local or remote compute clusters. The capacity to set up ad-hoc, on-
623 demand UAS data processing frameworks using open-source tools such as OpenDroneMap represents an
624 important direction in the development of regular UAS monitoring of critical ecosystems.

625 For other instrumentation, post-processing may have other challenges. For example, the processing of
626 LiDAR may require linking flight data with ground calibration targets and the interpolation of ground return
627 points to get accurate vegetation height information (Lefsky et al., 2002). For TIR, the use of a calibration
628 constant or in-image standard to generate absolute temperature retrievals is required, and ambient
629 conditions and other aspects related to sun-sensor geometry should be considered during processing
630 (Messina & Modica, 2020). Similarly, the processing of UAS-borne imaging spectroscopy may require the

631 collection of calibration targets and atmospheric corrections to retrieve at-surface reflectance (Adão et al.,
632 2017).

633 UAS data is usually stored onboard the platform and downloaded after a flight. This can raise data
634 provenance challenges, as it is important that the integrated UAS and flight control systems correctly link
635 the key metadata for each measurement that is necessary to establish data quality assurance and control,
636 and for spatial referencing during post-processing. Managing in-flight metadata (flight planning, mission
637 logs, atmospheric optical conditions) is also important for down-stream remote sensing normalization and
638 for atmospheric correction of multi- and hyper-spectral data into surface reflectance data products.

639 Long-term data preservation is also an important and often overlooked aspect of UAS research. Community
640 adoption of common data and metadata reporting formats is necessary to aid in the interoperability of UAS
641 data products. Currently, there are no dedicated archives for UAS data storage, nor are there common
642 standards for baseline metadata to ensure long-term data preservation. The increased use of UASs for arctic
643 research will require new means for data archiving, sharing, and access in order to facilitate wider use and
644 allow larger synthesis activities (Cunliffe et al., 2021). Traditional data sharing approaches (e.g., static data
645 archiving) are insufficient for effective storage, discovery, and sharing of large datasets, like UAS data.

646 UAS data should be accessible through an external, well-documented application programmer interface
647 (API) that can be coupled with open-source tools for data discovery, subsetting, and extraction directly
648 within data analysis workflows. New or expanded investments in open-source, cloud-based, secure, and
649 version-controlled data storage platforms that have hierarchical storage capabilities should facilitate ease
650 of discovery and use of UAS data. A notable example is the CyVerse Open Science Workspace
651 (<https://cyverse.org/>), which supports hierarchical data storage, includes file and archive version control,
652 provides digital object identifiers (DOIs) for data, and uses an API for data search, discovery, and retrieval
653 (Gillan et al., 2019, 2021; Swetnam et al., 2017). Platforms like CyVerse also provide cloud computation
654 and analysis support, facilitating a fully cloud-based data processing, storage, and publishing workflow
655 (e.g., Gillan et al., 2021). UAS data products can also be hosted over internet protocols such as HTTP(S),
656 cloud storage buckets (Amazon S3 and Google Cloud Services), and image collections in the Google Earth
657 Engine (GEE). Similarly, UAS datasets stored in the cloud benefit from a simple additional step to store
658 these files as Cloud Optimized GeoTIFFs which are optimized for data storage, subsetting, and processing
659 in the cloud without any additional server software requirements (<https://www.cogeo.org/>). Using these
660 tools and repositories can then facilitate rapid integration into other cloud-based tools, including GEE, as
661 well as popular extensions of these tools into R and Python, allowing for fully cloud-based analyses. For
662 example, by storing UAS data in a Google Cloud bucket, it is possible to share images and image collections

663 through GEE, thus making it easier for the broader use of datasets in scripted workflows that leverage other
664 spatial datasets.

665 **5.3. Future Directions of UAS Remote Sensing in the Arctic**

666 The expanded use of UASs is an important next step to improve our understanding of the fine-scale patterns
667 and drivers of plant composition, structure, function, and change in the Arctic. By combining these
668 platforms with traditional field surveys and plant ecology research, as well as integrating with other airborne
669 and satellite data, we will be able to develop the new techniques and methods necessary to better monitor
670 and model this globally important and climatically sensitive region (Cunliffe et al., 2020; Siewert &
671 Olofsson, 2020; Yang et al., 2021). Therefore, we strongly advocate for the increased use of optical RGB
672 and multispectral UASs, but also encourage efforts to use other sensors, such as thermal cameras, imaging
673 spectrometers, LiDAR, and SIF sensors. In particular, SIF sensors on UASs have already been shown to be
674 especially useful for capturing photosynthesis-related vegetation properties and functions, like GPP, in low-
675 latitude ecosystems (Chang et al., 2020; Wang et al., 2021). The widespread use of a greater range of sensors
676 will accelerate our ability to understand fine-scale variation in the form and function of arctic plants and
677 bridge the gap between ground and satellite observations.

678 Using UASs to investigate fine-scale vegetation patterns and foster ecological scaling represents their major
679 applications in the Arctic, but the collection of UAS data also extends to many other ecological applications
680 or different disciplines that are not included in this review (Gaffey & Bhardwaj, 2020). For example, the
681 construction of a reliable DEM is important to hydrological studies that capture or model surface or soil
682 water across arctic landscapes (Vélez-Nicolás et al., 2021). In addition, by flying UASs both prior to and
683 after snowmelt, snow depth can also be mapped, which is important to understand shrub-snow interactions
684 in the Arctic (Lawrence & Swenson, 2011). Further exploring the use of UASs in different areas, and
685 synthesizing them with vegetation applications, will be valuable to understanding the patterns, drivers of
686 change, and impacts of vegetation dynamics in the Arctic.

687 In light of the practical challenges of working in the Arctic, new strategies can be used to increase the
688 spatial and temporal coverage of UAS data. In the past decade, citizen science has grown immensely and
689 is regarded as an important tool for studies in ecology (Dickinson et al., 2010; McKinley et al., 2017). With
690 the advent of low-price UASs and sensors, citizen science holds a great potential to increase site
691 accessibility and foster long-term ecosystem monitoring of arctic ecosystems with UASs.

692 To sum up, with the fastest warming on Earth, widespread vegetation and land surface change is occurring
693 in the Arctic. In order to accurately assess these changes and project their impacts on arctic ecosystems, we
694 must address the challenge pertinent to this region – a high degree of spatial heterogeneity in vegetation

695 distribution, land surface structure, and environmental conditions. The deployment and use of new
696 technologies, like UASs, is an important part of the solution to address this challenge. Through this review,
697 we hope to shed light on the research opportunities provided by UASs and facilitate a broader use of this
698 technology to improve our description, understanding, modeling, and prediction of arctic ecosystems.

699 **Acknowledgment:**

700 This work was supported by the Next-Generation Ecosystem Experiments (NGEE Arctic) project that is
701 supported by the Office of Biological and Environmental Research in the United States Department of
702 Energy, Office of Science, and through the Department of Energy contract No. DE-SC0012704 to
703 Brookhaven National Laboratory. PRN was supported by NASA's Arctic Boreal Vulnerability Experiment
704 (ABoVE) Grant NNX15AU05A. We thank Editor Jason Fridley, Editor Emily Lines, and Editor Tommaso
705 Jucker for their encouragement to write this review and the great comments on the proposal and the draft
706 of this manuscript.

707 **Conflicts of Interest:**

708 The authors declare that they have no known competing financial interests or personal relationships that
709 could have appeared to influence the work reported in this manuscript.

710 **Author Contributions**

711 DY, SPS, BDM, KJD, and JL conceived the idea. All authors contributed to the writing of the manuscript.

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Figures and Tables:

Fig. 1: Illustration of arctic biomes (bottom panel), Unoccupied Aerial System (UAS) remote sensing technologies (middle panel), and the key applications of UAS remote sensing for plant research (top panel). The arrows between the elements of the middle and top panels indicate the connection between ecological applications and UAS remote sensing technologies. The arrows starting from the edge of the entire white box in the middle panel indicate that all available UAS technologies can be used for the specified ecological applications.

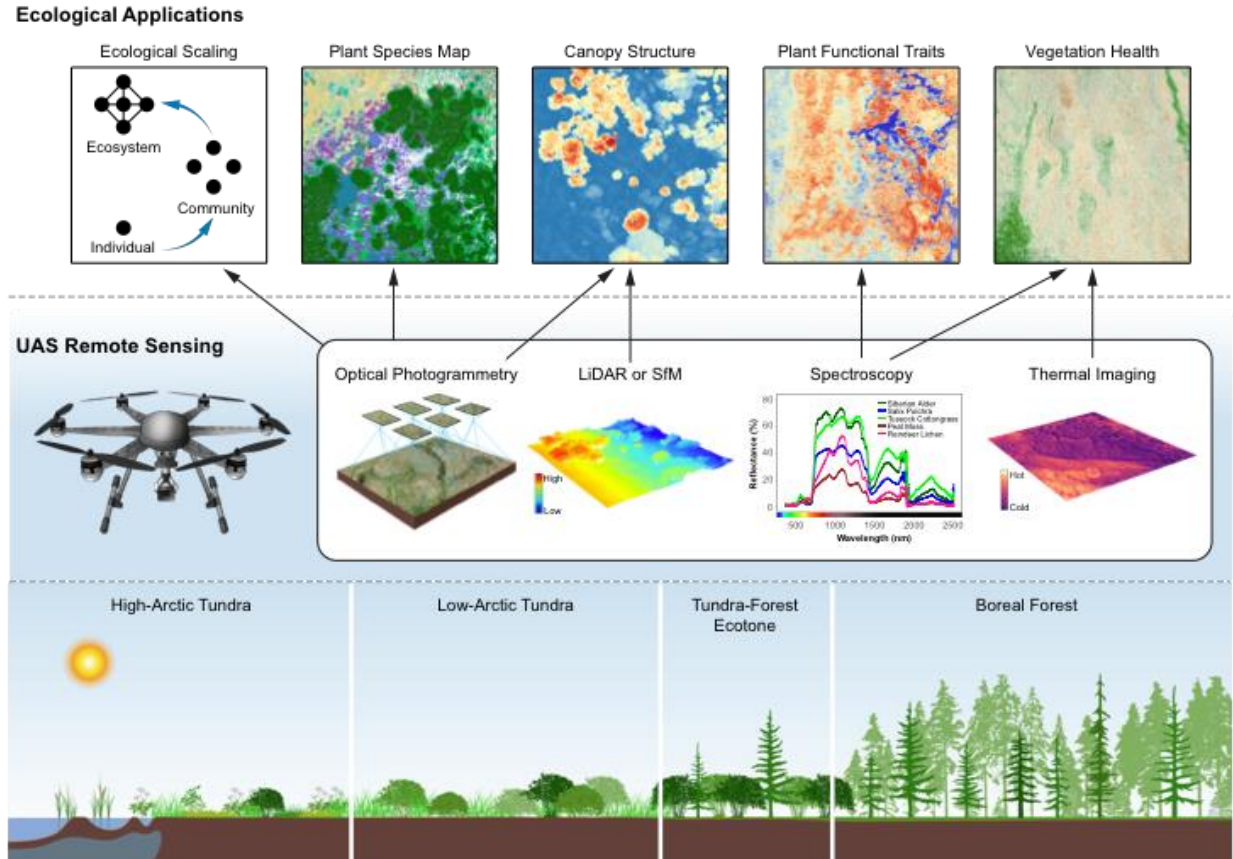


Fig. 2: Scale effects on vegetation composition and diversity analysis in the Arctic. The top 6 panel pairs show remote sensing images acquired at different spatial resolutions (i.e., 0.01, 0.1, 0.5, 1, 5, and 10 m) in a shrub landscape and the species maps derived from these images using a random forest classification. The bottom 2 panels are vegetation composition and Shannon diversity index calculated using the species maps shown in the top 6 panel pairs. A fusion of UAS-derived RGB ortho-mosaic, canopy height model, and canopy temperature was used for the classification. Non-vegetation components were excluded for calculating the Shannon diversity. Noticeably, decreasing spatial resolution significantly confuses species classification, biases vegetation composition analysis, and reduces Shannon diversity. Data used for this figure can be found in Serbin et al., (2021).

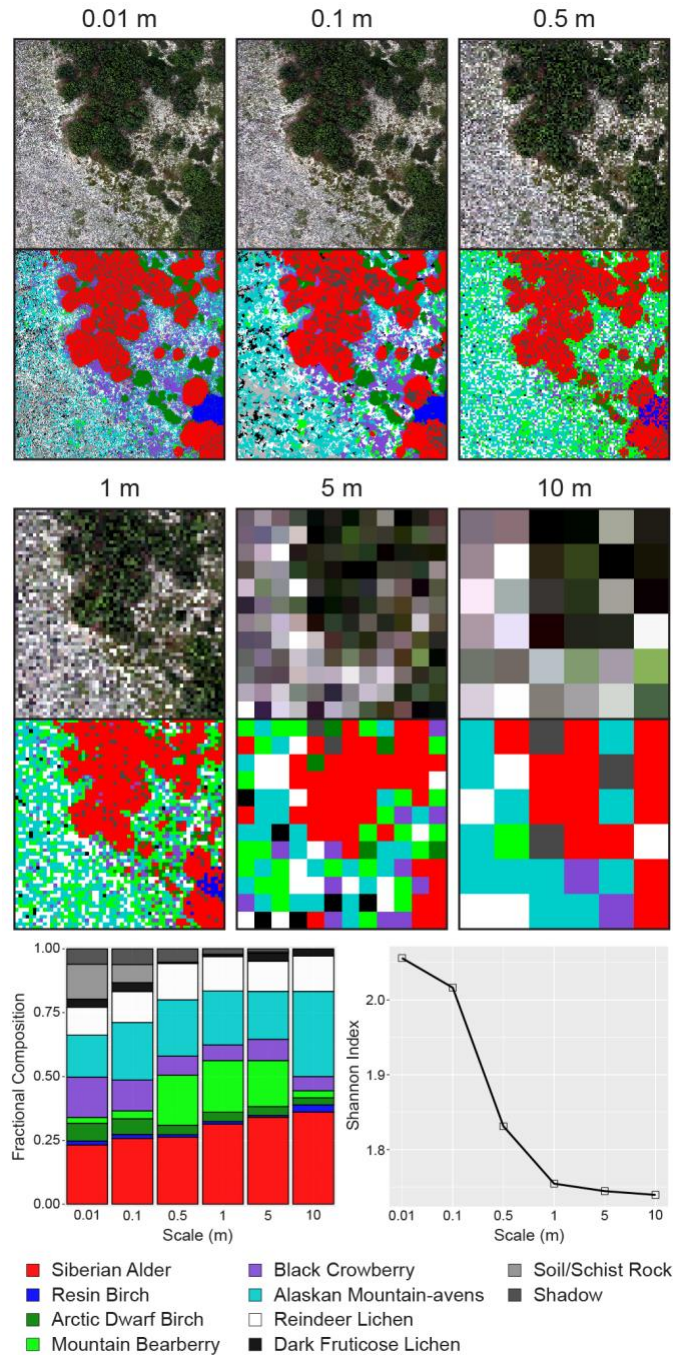
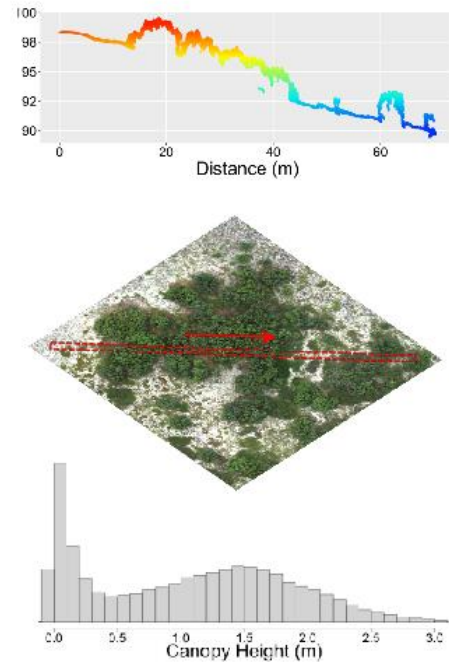


Fig. 3: Example application of UAS-derived point clouds for deriving canopy height and shrub cover in two representative arctic plant communities (a) Alder Tall Shrubland and (b) Alder Savannah. The effects of point cloud density (i.e., 1, 10, 100 points/m²) on deriving canopy height and shrub cover are demonstrated in the right three columns (Point Clouds, Canopy Height, and Shrub Cover). The height profiles (top-left panel in (a) and (b), respectively) correspond to the transects indicated by the dashed red lines shown in the RGB images. The histograms (bottom-left panel in (a) and (b), respectively) show the canopy height distribution of the entire landscapes indicated by the RGB images. The elevation of the point clouds and canopy height are both measured in meters. Data used for this figure can be found in Serbin et al., (2021).

(a) Alder Tall Shrubland

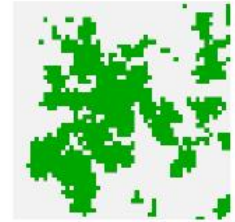
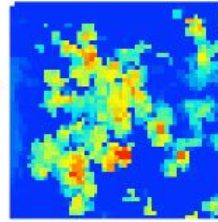


Point Clouds

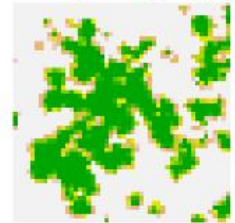
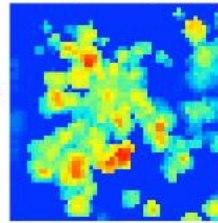
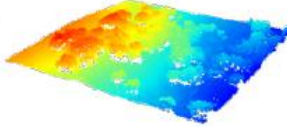
Canopy Height

Shrub Cover

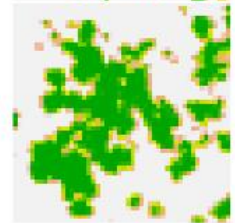
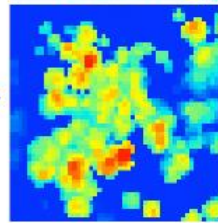
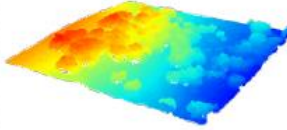
1 point/m²



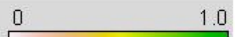
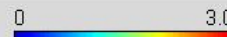
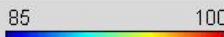
10 points/m²



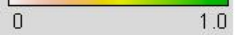
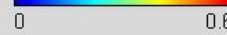
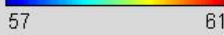
100 points/m²



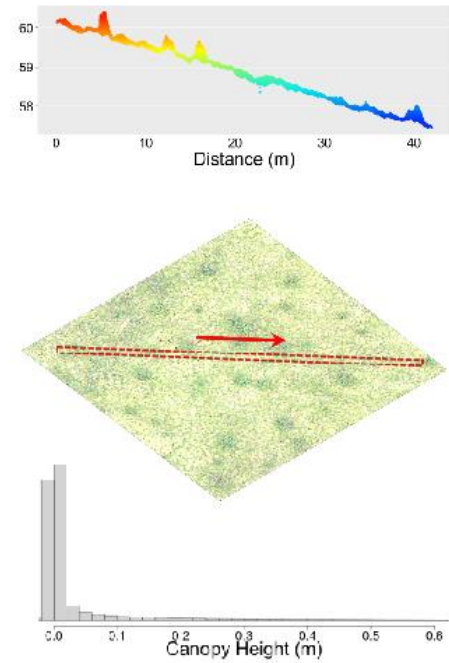
Alder Tall Shrubland



Alder Savannah



(b) Alder Savannah

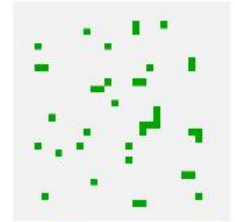
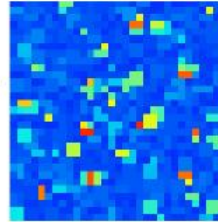


Point Clouds

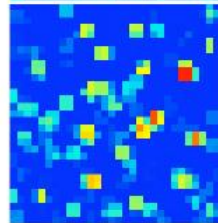
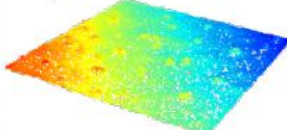
Canopy Height

Shrub Cover

1 point/m²



10 points/m²



100 points/m²

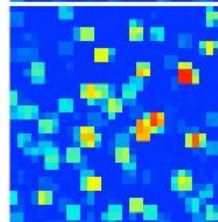
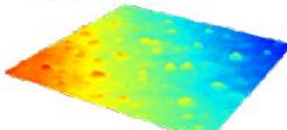


Fig. 4: Example of hyperspectral UAS imagery from a shrub landscape (a) and spectral signatures of key arctic tundra plants identified from the UAS imagery (b). The spectral signatures sensitive to different leaf and canopy traits are illustrated in (b). The hyperspectral UAS imagery displayed in (a) is acquired at a 5 cm spatial resolution. Data used for this figure can be found at: <https://github.com/nelsopet/lecospec>.

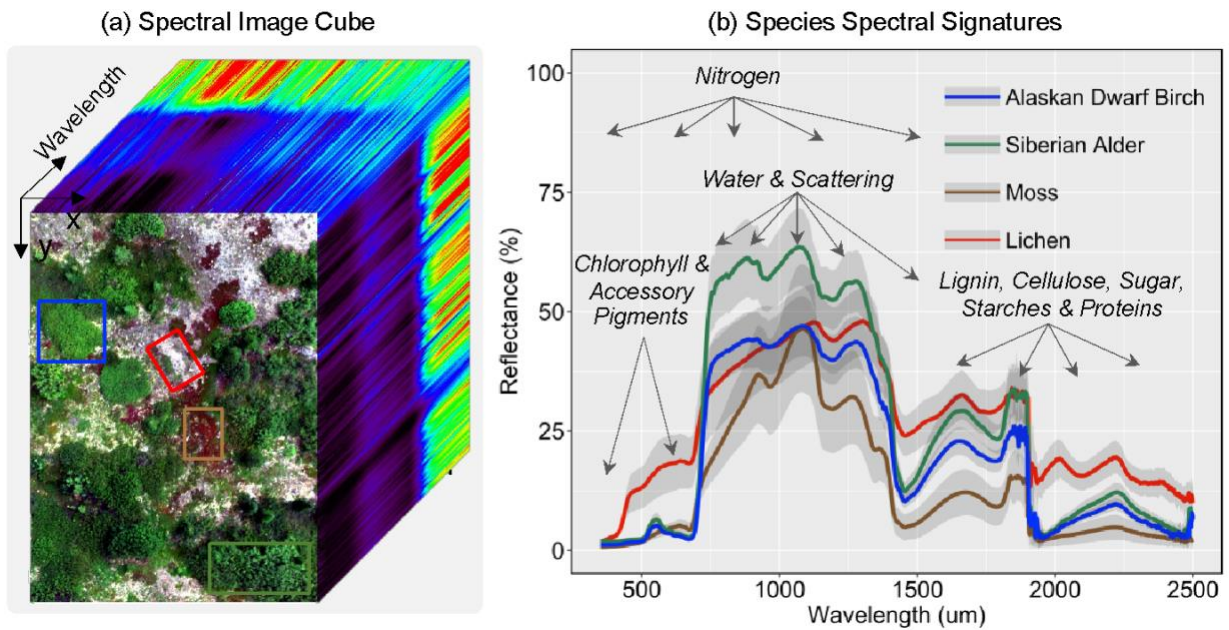


Fig. 5: Example application of high-resolution UAS thermal imagery for depicting thermal variation across an arctic tundra landscape. The thermal profile and histogram of four representative plant communities are demonstrated in (a) Willow Shrub, (b) Willow Birch Shrub, (c) Wet Meadow Tundra, (d) Ericaceous Dwarf Shrub Tundra. The thermal profiles (line plots) in (a) - (d) correspond to the transects (1 m wide) indicated in each panel. The black line in the thermal profile plots represents the mean temperature every 1 m step along the transect, and the ribbon region indicates temperature standard deviation within each 1 m step. The arrows on the thermal images indicate the direction of the transects, from left to right, that correspond with the thermal profiles. The data used for this analysis are collected using a multi-sensor UAS developed by Yang et al. (2020). Data can be found in Serbin et al., (2021).

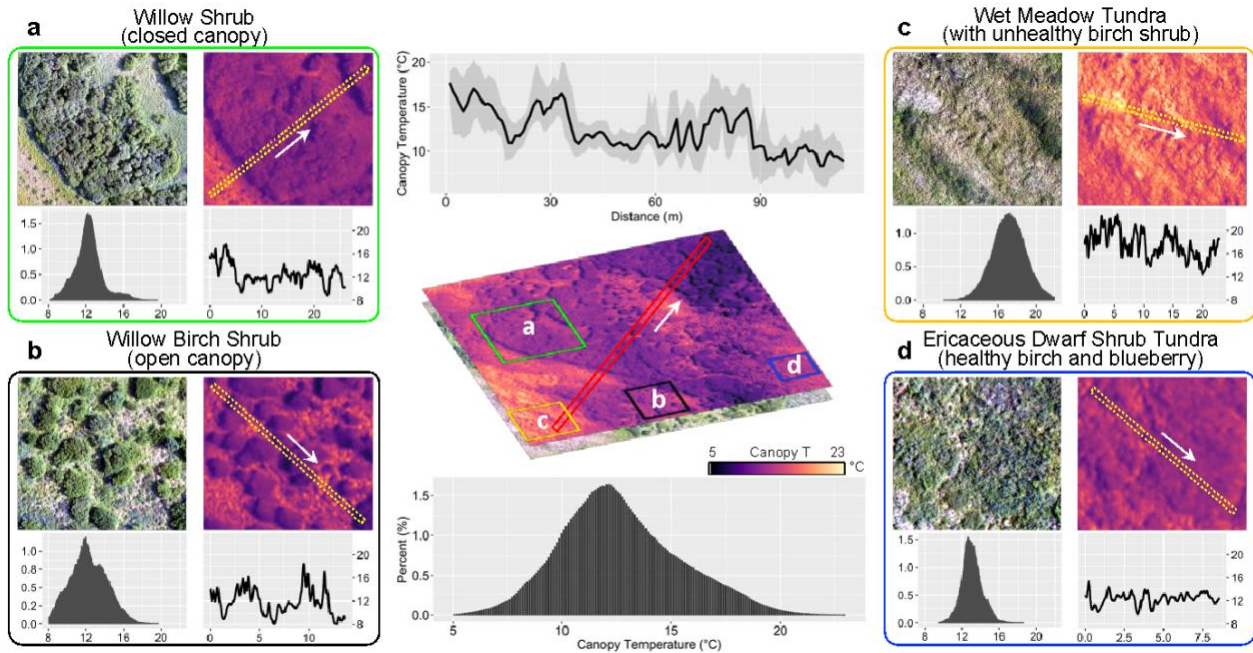


Table 1: Types of remote sensing sensors that have been implemented on UASs

Sensor Type	Description	References
Optical RGB camera	The most basic sensor type implemented with UAS, equipped with a standard complementary meta oxide semiconductor (CMOS) sensor through which red-blue-green colored images are collected.	Fraser et al. (2016); Yang et al. (2020)
Multispectral camera	Captures image data within specific wavelength ranges across the electromagnetic (EM) spectrum, commonly including blue, green, red, red-edge, and near infrared.	Assmann et al. (2019); Juszak et al. (2017)
Spectroradiometer	Measures the reflectance (or backward scattering) of solar radiation from an object or the emission (fluorescence) of the EM radiation from an object. A full range spectrometer covers the EM spectrum of visible to shortwave infrared (0.4 – 2.5 μm)	Chang et al. (2020); Malenovsky et al. (2017); Lucieer et al. (2014)
Thermal camera	A device that detects infrared radiation of a surface. Infrared cameras are sensitive to wavelengths from about 1 μm - 14 μm .	Yang et al. (2020); Hoffmann et al. (2016); Hoffmann et al. 2016); Ellsäßer, (2020)
Light detection and ranging (LiDAR)	An active remote sensing technology that uses light in the form of a pulsed laser to measure ranges (variable distances) to the Earth, to generate three-dimensional information about surface characteristics.	Lefsky et al. (2002); Collins et al. (2020)

Table 2: Key vegetation properties that can be derived from a UAS and the corresponding remote sensing techniques can produce each property. Note that spectral reflectance can be derived from two types of spectrometers: 1) point spectrometers, 2) imaging spectrometers. In this table, spectral reflectance indicate reflectance curves derived from both types of spectrometers, while hyperspectral imagery only indicates products from imaging spectrometers.

Key surface or vegetation property	Data products that can be used to derive the vegetation property	Sensor type	References
Plant species, plant function type, composition, and diversity	RGB imagery, Multispectral imagery, Hyperspectral imagery, Thermal infrared imagery, Canopy height model, Point clouds	Optical RGB camera, Multispectral camera, Imaging spectroradiometer, Thermal camera, LiDAR	Lucieer et al. (2014); Fraser et al. (2016); Juszak et al. (2017); Alonzo et al. (2016); Yang et al. (2020 & 2021); Thomson et al. (2021)
Surface or vegetation albedo	Spectral reflectance, Multispectral imagery	Multispectral camera, Imaging or point spectroradiometer	Canisius et al. (2019); Xu et al. (2020)
Plant functional traits	Spectral reflectance, Thermal infrared imagery	Multispectral camera, Imaging or point spectroradiometer, Thermal camera	Shiklomanov et al. (2019); Thomson et al. (2021)
Water content	Thermal infrared imagery, Spectral reflectance	Thermal camera, Imaging or point spectroradiometer, Multispectral camera	Ellsaber et al. (2020); Chan et al. (2021); Thomson et al. (2021)
Land-surface or canopy "skin" temperature	Thermal infrared imagery	Thermal camera	Jones et al. (2003); Costa et al. (2013); Yang et al. (2020); Still et al. (2019 & 2021)
Solar-induced fluorescence (SIF)	Very-fine spectral resolution reflectance	Imaging or point spectroradiometer	Chang et al. (2020)
Canopy height, cover, biomass	Point clouds	LiDAR, Optical RGB camera	Andersen & Gaston (2003); Alonzo et al. (2016 & 2020); Cunliffe et al. (2020 & 2021)
Digital Elevation Model (DEM)	Point clouds	LiDAR, Optical RGB camera	Fraser et al. (2016); Yang et al. (2020); Alonzo et al. (2020)
Seasonality and phenology	RGB imagery, Multispectral imagery, Hyperspectral imagery, Thermal infrared imagery	Optical RGB camera, Multispectral camera, Imaging spectroradiometer, Thermal camera	Assmann et al. (2020)

Table 3: Key arctic ecological questions or knowledge gaps that UAS can help or partially help with.

Research area	Key ecological questions or knowledge gaps	Measurement needs	Role of UAS platforms
Vegetation distribution and surface heterogeneity	How does the distribution of Arctic vegetation vary among species and locations in the Arctic?	Species-specific vegetation maps across Arctic landscapes	Provide data to improve identification and mapping of plant species (3.1) at fine-scales and provide training data for larger mapping efforts
	How do plant biophysical properties differ among vegetation types in the same locality?	Species-specific or fine-scale estimates of plant biophysical properties	Very high-resolution maps of vegetation structure, traits, and function (3.2 - 3.5)
	What is the connection between fine-scale vegetation and surface patterns with larger ecosystem processes?	Detailed information on vegetation composition, structure, and function as it relates to fine-scale surface features (e.g., topography, moisture, and disturbance events).	Linking UAS measurements of vegetation and surface properties with ecosystem-scale measurements from flux towers or satellites can help explain the drivers of variation in ecosystem properties (3.1 - 3.6)
Shrubification	What controls Arctic shrubification?	High-fidelity shrub cover and type maps linked with environmental gradients and disturbance history	Fine-scale characterization of shrub species, traits, biophysical properties. Connect fine-scale variation with landscape features (e.g., topography, water, and disturbances) to identify abiotic and disturbance controls on shrub distribution (3.1 - 3.5).
	How does shrubification affect plant biodiversity?	Accurate vegetation composition and diversity maps together with shrub fractional cover information	High-resolution maps of vegetation fractional cover that can be used to train regional mapping efforts (3.1)
	How does shrubification interact with snow accumulation and permafrost thaw?	Spatially-detailed snow depth and thaw measurements in relation to shrub distribution, cover, and structure.	Create high-resolution maps of shrub structure, distribution, and cover to link with other remotely sensed (SAR, LiDAR) or field survey measurements of snow and thaw depth data (3.1 & 3.2)
	How does shrubification affect surface energy and water exchange?	Accurate characterization of shrub LAI, albedo, and surface “skin” temperature across shrub types and environmental gradients	A fine-scale understanding of the influence of shrubs on landscape vegetation dynamics and sensible heat exchanges, to inform broader scaling, mapping, and modeling efforts (3.2 & 3.4)

Arctic greening	What are the landscape controls on the rate of “Arctic greening”?	Improved understanding of the local-scale drivers of the larger regional variability in Arctic greening and its connection with changes in vegetation and abiotic environments	Collection of UAS data for select areas can help resolve fine-scale drivers (e.g., shrubification, sub-pixel disturbance, changes in species composition) of the larger scale greening signal in coarse resolution satellite data. Collection of UAS data at key phenophases can be used to explore how vegetation seasonality influences greenness. UAS observations can also be used to parameterize radiative transfer models to simulate how different landscape features influence the emergent reflectance patterns at coarse resolution (3.1-3.3 & 3.5 - 3.6)
	How does scale (both spatial and temporal) affect our understanding of greening in the Arctic?		
Disturbance	How do the effects of disturbance type and extent vary across different vegetation and ecosystem types?	High-resolution disturbance area mapping in relation to pre-disturbance vegetation type, topography, and soil conditions	Paired UAS flights before and after disturbance can be used to investigate patterns of change in relation to disturbance extent (3.1)
	How does disturbance severity control recovery patterns of tundra vegetation?	Time-series and high-resolution vegetation composition and status mapping before and after disturbance (fire, permafrost thaw, or pest outbreaks)	Repeat UAS flights to train larger upscaling methods to enable time-series monitoring of vegetation composition and status during recovery (3.1 & 3.6)
	After permafrost thaw, what is the successional trajectory of the vegetation?		
Model uncertainty	What PFTs currently populate the Arctic?	Improved estimates of fractional cover of PFTs	High-resolution and detailed PFT mapping and upscaling to satellite platforms (3.1)
	What is the current structure and stand biomass in the Arctic?	Better estimates of standing biomass and canopy structure and in relation to disturbance and land-use history.	Characterize standing vegetation height, structure, and biomass properties (3.2). Validation of larger biomass mapping efforts (3.3)
	How do we improve the parameterization of terrestrial biosphere models in the Arctic?	Measure key parameters currently in models but poorly represented and characterize their variation along climatic gradients	UAS platforms can be used to develop fine-scale maps of plant traits and then link the variation in traits with environmental and climatic gradients in the Arctic (3.3)

Vegetation phenology is not driven by biology or abiotic conditions but is prescribed.	Monitoring of vegetation seasonality at a fine-resolution at landscape scales	Repeat flights can be used to define phenological timing. Fine-scale maps of surface topography and snow distribution from UASs can be connected with phenological observations to study how abiotic features regulate vegetation seasonality (3.3)
Green leaves do not always equate with full photosynthetic capacity.	Improved understanding of the controls on photosynthetic capacity and stress tolerance including abiotic drivers, and biotic and temporal variation	Imaging spectroscopy with UASs will enable high resolution mapping of photosynthetic traits and linking with surface and environmental gradients (3.3). Thermal imaging and SIF from UAS allows for the characterization of photosynthetic activity at the scale of individual plants.
When are Arctic plants photosynthesizing?		
How consistent is photosynthetic temperature response across the landscape?		
What is the impact of scale on model predictions?	Multi-scale characterization of vegetation and environmental measurements	Fill the scaling gap between leaf/individual scale measurements of traits and other airborne and spaceborne remote sensing platforms